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Stream diatom biodiversity in islands and continents—A global perspective on effects of area, isolation and environment

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1	Stream diatom biodiversity in islands and continents – a global
2	perspective on effects of area, isolation and environment
3	
4	Short running title: Diatom in islands and continents
5	
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48

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## 67 Abstract

85

68	Aim. The species-area relationship (SAR) is one of the most distinctive biogeographic
69	patterns, but global comparisons of the SARs between island and mainland are lacking
70	for microbial taxa. Here, we explore whether the form of the SAR and the drivers of
71	species richness, including area, environmental heterogeneity, climate and
72	physicochemistry, differ between islands and similarly sized areas on mainland, referred
73	to as continental area equivalents (CAEs).
74	Location. Global.
75	Major taxa studied. Stream benthic diatoms.
76	Methods. We generated CAEs on six continental datasets and examined the SARs of
77	CAEs and islands (ISAR). Then, we compared CAEs and islands in terms of total
78	richness and richness of different ecological guilds. We tested the factors contributing to
79	richness in islands and CAEs with regressions. We used structural equation models to
80	determine the effects of area vs. environmental heterogeneity, climate and local
81	conditions on species richness.
82	Results. We found a non-significant ISAR, but a significant positive SAR in CAEs.
83	Richness in islands was related to productivity. Richness in CAEs was mainly dependent
84	on area and climate, but not directly on environmental heterogeneity. Species richness

Main conclusions. Ecological and evolutionary processes shaping diatom island
biogeography do not depend on area at the worldwide scale probably due to the presence
of distinct species pool across islands. Conversely, area was an important driver of

within guilds exhibited inconsistent relationships with island isolation and area.

diatom richness in continents, and this effect could be attributed to dispersal. Continents
had greater richness than islands, but this was a consequence of differences in
environmental conditions such as specific island climatic conditions. We stress the need
for more island data on benthic diatoms, particularly from archipelagos, to better
understand the biogeography of this most speciose group of algae.

94

## 95 Keywords

96 ecological guilds, freshwater diatoms, island biogeography, macroecology, species-area

97 relationship, streams

#### 98 Main Text

#### 99 Introduction

100 A fundamental ecological law that describes how the number of species increases with 101 area is the species-area relationship (SAR, Arrhenius, 1921). The SAR belongs to a few, truly robust generalizations in ecology detected in a wide range of ecosystems and taxa 102 (Connor & McCoy, 1979; Rosenzweig, 1995; Drakare et al., 2006). Islands represent 103 perhaps the most straightforward study setting to explore the SAR because of their well-104 defined area. Unlike most mainland habitat patches, islands are surrounded by an 105 inhospitable matrix for continental taxa, which cannot be colonized and, consequently, 106 107 cannot serve as a source of immigrants. This peculiar feature of islands inspired MacArthur & Wilson to develop the theory of island biogeography (MacArthur & 108 Wilson, 1967), which has contributed enormously to modern biodiversity theory (Chase 109 110 & Leibold, 2003), metapopulation biology (Hanski & Gaggiotti, 2004), community 111 ecology (Mittelbach & McGill, 2019), landscape ecology (Farina, 2008) and biodiversity conservation (Prugh et al., 2008). 112

Island biogeography investigates how species richness on islands varies spatially and through time (Whittaker & Fernandez-Palacios, 2007). It postulates that larger and less isolated islands host more species than small and remote islands because larger area decreases extinction and proximity to mainland increases immigration. Larger islands may also encompass more species because they provide a larger target for immigration, higher habitat diversity (Lack, 1976) and have higher speciation rates (Whittaker & Fernandez-Palacios, 2007). Lastly, since island age affects diversification and erosion, it

120 may also determine species richness, which tends to be the highest in islands of 121 intermediate age according to the general dynamic model (Whittaker et al., 2008, 2017). 122 In the light of this knowledge, Chase et al. (2019) recently presented a framework for the ecological mechanisms underlying the island SAR (ISAR). They suggested that passive 123 sampling (i.e. larger islands passively sample more individuals and species from the 124 125 regional pool than smaller islands), disproportionate effects (e.g. different colonization and extinction rates in larger vs. smaller islands) and habitat heterogeneity (greater in 126 127 large islands) would be the main drivers of ISAR. Nevertheless, the major patterns and 128 drivers of island vs. mainland SAR are still poorly understood, particularly for the species rich microorganisms. 129

Given the importance of environmental heterogeneity and dispersal on the SAR (Chase et 130 al., 2019), functional groups varying in resource utilization and dispersal can have 131 132 different SARs (Lomolino & Weiser, 2001; Báldi, 2008; Schrader et al., 2020). For example, the SAR slope was steeper for specialist than for generalist bird species 133 (Matthews et al., 2014) and functional traits related to dispersal explained the SAR 134 variation in plant communities (Schrader et al., 2020). Thus, evaluating the SAR of 135 136 different ecological guilds may improve the knowledge of the niche- vs. dispersal-related processes behind the SAR patterns. As functional diversity may have a distinct 137 (Jamoneau et al., 2018; Schrader et al., 2020) and even stronger response to 138 environmental variation than species diversity (Krause et al., 2014; Abonyi et al., 2018), 139 140 the SAR for different functional groups may elucidate how community assembly processes operate through space and time (Tilman et al., 1997). 141

142	The ISAR has been tested with larger-bodied organisms, including terrestrial arthropods
143	(Simberloff & Wilson, 1969) and reptiles (Algar & Losos, 2011), birds (Kalmar &
144	Currie, 2006, 2007), vascular plants (Kreft et al., 2008), and fish (Sandin et al., 2008).
145	However, ISAR patterns are still poorly understood for microorganisms. Earlier
146	microbial field studies that used microcosms (Smith et al., 2005), lakes (Reche et al.,
147	2005), trees (Bell et al., 2005; Peay et al., 2007) or spring ecosystems (Teittinen &
148	Soininen, 2015) as surrogates of islands, reported significantly positive ISARs in almost
149	all systems (but see Teittinen & Soininen, 2015). However, investigations on
150	microorganismal diversity in real islands at a global scale are, to our knowledge, still
151	missing.
152	Rosenzweig (1995) hypothesized that islands should have lower local and regional
153	species richness than similarly sized continental regions due to isolation (lower mass- and
154	rescue effect), but steeper SAR slopes. This is because area tends to be a more critical
155	factor for biota on islands than on continents due to its stronger effects on extinction and
156	colonization (Kreft et al., 2008). However, in an extensive meta-analysis, Drakare et al.
157	(2006) did not find evidence for steeper SARs on islands (ISARs) than on mainland
158	across multiple species groups. The SAR patterns are typically explored within
159	archipelagos due to the presence of a common species pool, allowing assessment of the
160	pure area effect. However, there are also more general models for the SAR at the global
161	scale, searching for broader influences on the SAR (Kalmar & Currie, 2006; Kreft et al.,
162	2008; Triantis et al., 2015), including differences in evolutionary history (Rosenzweig,
163	1995).

165 Here, we adopted a similar perspective and investigated freshwater diatom SAR at a worldwide scale, given that diatoms have large distributions (Finlay, 2002) and are 166 strongly controlled by environmental conditions (Soininen et al., 2016). We compared 167 168 SARs, total species richness, and species richness drivers between islands and corresponding areas on five continents, referred to as continental area equivalents 169 170 (CAEs). For this comparison, we devised a novel method based on island-mainland pairs (Fig. 1), assuming that terrestrial area is a good surrogate for area of freshwater habitat 171 (see Appendix S1 in Supporting Information). The CAEs corresponded to the sampling 172 173 area of 18 islands. We then examined (1) if SAR slopes differed between islands (ISAR) 174 and continents, (2) whether islands showed overall lower diatom species richness than 175 CAEs, (3) if species richness of island was related to environment, spatial isolation or 176 island age, and (4) whether habitat diversity, passive sampling or disproportionate effects explained the SAR. We investigated these research questions separately for total diatom 177 178 species richness and species richness of ecological guilds, differing in dispersal capacity and tolerance to nutrient limitation and disturbance (Passy, 2007, 2016), all expected to 179 influence the SAR (Matthews et al., 2014; Schrader et al., 2020). 180

## 181 Materials and Methods

## 182 Biological and environmental datasets

183 In total, we included 18 island datasets (Corsica, Cyprus, Guadeloupe, Iceland, Ireland,

184 Kauai, La Réunion, Martinique, Madeira, Majorca, Mayotte, New Caledonia, North New

185 Zealand, Oahu, Possession, São Miguel, Sardinia and South New Zealand) and six

- 186 continental datasets (China, Finland, France, French Guiana, Kenya and USA) in our
- 187 study (see Appendix S2). Diatoms were sampled from hard substrates (typically stones)

or macrophytes, generally during the low flow period (see Appendix S2 for details).

189 Although diatoms in some datasets were collected over several years, we did not expect a

190 substantial effect of interannual variation in our study, because we were interested in

regional diversity patterns and included environmental variables to account for this

192 potential variation.

Diatoms were cleaned with acid or hydrogen peroxide. A total of 400-700 diatom valves 193 were counted for each sampling site, which is sufficient for reliable estimates of total 194 195 diversity (Heino & Soininen, 2005). As the number of counted valves varied somewhat 196 among the samples, we studied if this would affect our richness estimates. We estimated species richness with 300 valves and tested the correlation with the observed species 197 198 richness. We observed a very strong relationship between the estimated and the observed species richness ( $R^2_{ai} = 0.98$ ). Also, valve counts did not differ significantly between 199 200 islands and continents (Cliff test difference for large dataset, delta=-0.15). We thus believe that the number of counted valves has only marginal impact on our richness 201 results. 202

203 Diatoms were generally identified up to species level, except in some rare case where some of the valves were identified only to genus level (representing less than 5% of the 204 entire dataset). Homogenization of the taxonomy among regions was performed using the 205 OMNIDIA database (Lecointe et al., 1993, updated in November 2020). To ensure that 206 we have a proper estimate of the diversity, we i) evaluated the proportion of observed 207 208 species compared to the size of the species-pool in each region using basic Chao equation (Chao, 1987) and calculated a 'corrected' species richness according to this ratio (i.e. the 209 observed species richness was increased relative to the proportion of missing species 210

211 estimated from the species pool) and ii) calculated a genus-based richness assuming that 212 genus level identification varies much less among diatomists than species identification. We then ran analyses with observed species richness, corrected species richness and 213 genus richness (see Data analyses section). In total, our datasets comprised 1967 taxa, 214 further classified into four ecological guilds: low profile (species of short stature), high 215 216 profile (species of tall stature, typically filamentous, colonial or branched), motile (species moving freely in the biofilm) and planktonic species (species not innate to the 217 benthos but originating from planktonic sedimentation) (Passy, 2007; Rimet & Bouchez, 218 219 2012; Soininen et al., 2016). Contrary to motile and high-profile species, low-profile species are tolerant to nutrient limitation and disturbance and exhibit wider distributions 220 221 (Passy, 2016), suggesting potentially higher dispersal capabilities (Heino & Soininen, 222 2006). Planktonic species may indicate important features of the sites such as low current velocity and large rivers. 223

Physico-chemical data of each sampling site included pH, conductivity (µS.cm<sup>-1</sup>), total 224 phosphorus (mg.l<sup>-1</sup>) and water temperature (°C), with the exception of Finland and 225 226 Possession island (with no water temperature data) and Ireland, Kenya and New Zealand (with no total phosphorus data). Physico-chemical data were collected up to two months 227 228 before the diatom sampling. Climate data were obtained from WorldClim database at 0.5 229 minutes resolution (Hijmans et al., 2005), including annual precipitation (mm), seasonality in precipitation (%), annual temperature (°C), and temperature seasonality 230 (standard deviation of monthly mean temperatures). For each sampling site, we also 231 232 extracted elevation from the Global Multi-resolution Terrain Elevation Data 2010 (Danielson & Gesch, 2011) and computed terrain slope as a proxy for current velocity. 233

For islands, we determined age of formation from the literature (see Appendix S2) and isolation using the isolation index of Dahl (Dahl, 1991, Gillespie et al., 2008). This index (equation 1) is based on the sum of square root distances to the nearest equivalent or larger island ( $d_i$ ), the nearest island group or archipelago ( $d_a$ ), and the nearest continent ( $d_c$ ).

239 Isolation index = 
$$\sqrt{(d_i)} + \sqrt{(d_a)} + \sqrt{(d_c)}$$
 (1)

241 For a reliable comparison of species-area relationships between islands and continents,

242 which are vastly different in size, we generated CAEs, comparable in size to the islands

by taking subsets of the continental data (see Algar & Losos, 2011) for a related

approach). The method used to create these CAEs (Fig. 1) was as follows.

245 We first computed the geographical centroid of each island and calculated  $D_{c-i}$ , a vector representing the Euclidean distance between the centroid and each island sample site *i*. 246 Second, for each continent, we calculated  $D_{jj}$  the Euclidean distance matrix between 247 248 sample sites *j*. All Euclidean distances were calculated from geographical coordinates expressed in a projected geographical system adapted for each region (see Appendix S2). 249 250 Third, we treated all continental sites as candidate CAE centroids and calculated D<sub>v-i</sub> representing the Euclidean distance between the candidate CAE centroid v and all other j 251 continental sites. We then computed a matrix  $DD_{\nu_j,c-j}$  (equation 2), which represented the 252 253 absolute difference between i) the distance between a candidate CAE centroid and all other sites in the focal continent  $(D_{\nu_{ij}}, i.e. raw of the matrix D_{ij})$  and ii) the distance 254 between the island centroid and all other sites in the focal island  $(D_{c-i})$ . 255

256 
$$DD_{v-j,c-i} = |D_{v-j} - D_{c-i}|$$
 (2)

Note that the minimum value of  $DD_{v:j,c-i}$  is theoretically 0, indicating that the distance between a centroid and an island site *i* is identical to the distance between a CAE centroid and a continent site *j*. Thus, smaller  $DD_{v:j,c-i}$  equates to similar distances between an island centroid and island sites and the distances between a candidate CAE centroid and continental sites. We then assigned for each centroid-island site distance a unique corresponding CAE centroid-continent site distance ( $\Delta_{c-i,v:j}$ , i.e. the minimum value of the column of  $DD_{v:j,c-i}$ , equation 3).

264 
$$\Delta_{c-i,v-j} = min(DD_{.,c-i}) (3)$$

Then, we considered that the CAE centroid could be considered as the centroid of a CAE 265 only if at least N=15 of the selected CAE centroid-continent site distances  $\Delta_{c-i,v-j}$  were 266 267 below a threshold value ( $\theta$ ) set to 5 km. Thus, theoretically, the number of sites in each CAE could vary between 15 and the total number of sites in each island. Note that 268 269 because Kauai and Possession islands have less than 15 sites, N was set to 10 for the 270 creation of their CAEs. Finally, to avoid pseudoreplication within the sites of CAEs, we selected for each continent-island pair only CAE separated by a distance of at least twice 271 the mean distance between centroids and their corresponding sites. Due to this procedure, 272 273 the size of CAEs could be, in some rare case, much smaller than the corresponding island size. 274

The CAEs, corresponding to the sampling area of an island, were successfully created in all continents (see appendix S3). Kenya was an exception because we were unable to create CAEs corresponding to Corsica, Iceland, Ireland, New Caledonia, Sardinia, North and South New Zealand, which were larger in size. Also, following our methodology, it

279 was not possible to create CAEs corresponding to the island of Mayotte in USA and

280 Finland, the island of São Miguel in Finland and USA and Possession Island in China and

USA because their continental sites were more spread out than the island sites.

282 Consequently, the total number of continent-island pairs for creating CAEs was 96.

Randomization procedure for calculation of species richness and other environmental
variables

For each continent-island pair, we randomly selected 15 sites within the CAEs and 15 285 sites within each island (10 sites for Kauai and Possession and their respective CAEs; 20 286 287 iterations) to achieve comparable sampling effort for islands and CAEs. For each random subset, we calculated species richness as the total number of species observed among the 288 15 sites, and area from the convex hull around these 15 sites. We found that the areas in 289 290 islands estimated using convex hulls were good surrogates for whole island areas (see appendix S4). We also calculated median values for each environmental variable for the 291 15 sites in the CAEs and islands and computed their environmental heterogeneity as the 292 293 multivariate dispersion of all environmental variables using the average distance of all 294 samples to the sample centroid in the multivariate space with the *betadisper* function in the *vegan* package. Environmental variables used in the analyses and computation of 295 environmental heterogeneity were selected because they are known to be important for 296 stream diatom distributions (Soininen, 2007; Soininen et al., 2016). For the computation 297 298 of environmental heterogeneity in the Kenya and Ireland dataset, we respectively used total nitrogen and orthophosphate concentrations instead due to the lack of total 299

301 heterogeneity for New Zealand and Possession islands due to missing data).

302 Data analyses

303 We conducted separate analyses for CAEs and islands to examine the relationship between species richness and area (SAR). We used linear mixed models (LMMs) for 304 CAEs to account for continental influences that may underlie differences in species pools 305 306 and the potential lack of independence among CAEs, given that multiple CAEs were created within a continent (i.e. continents were included as random factors). We 307 performed traditional linear models for islands. We tested SAR with three commonly 308 309 used models (DeMalach et al., 2019), including power (Arrhenius model), logarithmic (Gleason model) and Michaelis-Menten, and selected the best model based on the lowest 310 Akaike Information Criterion. We also tested relationships between area and the 311 312 'corrected' species richness (according to the size of the species pool) and genus richness 313 to ensure that the sampling effort or the taxonomic resolution did not influence our results. 314

To test for passive sampling, we estimated species richness from rarefaction curves based 315 on species occurrence. For each CAE and island, we pulled at random 15 sites and 316 317 randomly selected 130 species occurrences without replacement, thus ensuring that the 318 maximum occurrence of each species did not exceed 15. Species richness was then 319 estimated from the 130 occurrences and used to generate the SAR, which was fit with mixed models for CAEs and traditional linear models for islands. According to Chase et 320 al. (2019), failure to detect SAR using this estimation of species richness would suggest 321 322 that SAR is caused by passive sampling only. However, the reverse is not true, and

significant SAR observed with this estimation of species richness does not necessarily
prove the absence of passive sampling (Chase et al., 2019).

Then, to test for the effect of area on species richness after controlling for environmental 325 variation, we first computed global LMMs for total and guild species richness and eight 326 environmental variables (pH, conductivity, elevation, annual temperature, annual 327 precipitation, temperature seasonality, precipitation seasonality and environmental 328 heterogeneity). Models were constructed using the median values of species richness as 329 330 the response variable and median environmental variables obtained from the subsampling 331 procedure as explanatory variables (N = 851, i.e. one value for each 833 CAEs and each 332 18 island). Prior to analyses, explanatory variables were log-transformed to improve 333 normality when necessary and standardized, but we did not treat for multicollinearity here, as this does not affect the fit of the model. Second, residuals from these regressions 334 were regressed against log-transformed area with LMMs for CAEs and simple linear 335 models for islands. 336

The number of islands in our study is comparable to the number of islands in many other 337 SAR studies (see data used in Matthews et al., 2019) but admittedly not very high (N=18) 338 for a study at the worldwide scale (Kalmar & Currie, 2006). Therefore, we performed a 339 sensitivity test with our continental datasets to determine the number of CAEs required 340 for observing a significant SAR, acknowledging that the number of islands and CAEs 341 necessary to detect a SAR may be different. We used the median values of the species 342 343 richness and area obtained from the randomization procedure for each continent-island pair (N=96), and randomly sampled (1000 times) K continent-island pairs. Each time we 344 fit the SAR with the best SAR model (logarithmic) and extracted the probability (P) of 345

346 observing a significant SAR, as well as the median values of model coefficients. We varied K from 11 to 96, i.e. the total number of continent-island pairs available in our 347 dataset. We then identified the minimum number of 'islands' needed to observe a 348 349 significant SAR with our data (P > 95%). We performed these analyses with both traditional linear models and LMMs (e.g. assuming a common species pool). 350 We compared species richness of islands and species richness of their corresponding 351 CAEs with Cliff's non-parametric effect size statistic (Romano et al., 2006; Tecchio et 352 353 al., 2016), due to the large number of data points resulting from the randomization 354 procedure (i.e., decreasing variance around the mean). We also used Cliff's tests to compare the species richness of each ecological guild between CAEs and islands. 355 To compare species richness of CAEs and islands after removing the effect of 356 environment, we computed LMMs as above but also included all the values of random 357 358 subsampling (x20, N=17020). We therefore used a nested design in the random factors of the models, so that subsampling values are nested within each continent/island. Residual 359 richness values were then extracted from the models and compared between CAEs and 360 islands with Cliff's tests. 361 We used linear mixed models for CAEs and traditional linear models for islands to 362 363 examine the relationship between species richness, environmental heterogeneity, the 364 median of all environmental variables and the median values of latitude and longitude. 365 Environmental explanatory variables were log-transformed to improve normality when

366 necessary and we run separate regression models with each environmental factor and

367 species richness to avoid multicollinearity. We also tested for non-linear relationships

368 separately with all environmental variables with the same procedure.

369 Finally, to disentangle the possible drivers of the SAR for continents, we implemented piecewise structural equation modeling (SEM, Lefcheck, 2016) using linear mixed 370 models with continental dataset as a random factor. We could not implement such models 371 for islands due to an insufficient number of data points. We assumed an a priori model 372 (Fig. 2) predicting species richness as directly influenced by area, environmental 373 374 heterogeneity (as defined above), local environmental conditions and climatic conditions. We used conductivity and elevation as predictors of local conditions, temperature 375 seasonality and annual precipitations as predictors of climate, as they were significant 376 377 predictors of species richness in global LMMs and exhibited low collinearity in pairwise correlations tests (see Appendix S5). We assumed that the effect of area on species 378 379 richness could also be indirect through environmental heterogeneity, according to the 380 habitat diversity hypothesis (Lack, 1976). Finally, we also assumed that temperature seasonality and precipitation are directly influenced by elevation. We included a 381 382 correlation between temperature seasonality and precipitation as well as between conductivity and precipitation (see Appendix S5). We used the Fisher's C statistic to test 383 the consistency of the theoretical model with the data. All analyses were run for total 384 385 richness and separately for richness of each ecological guild.

All analyses were conducted with R (R Core Team, 2019) using packages 'vegan'

387 (Oksanen et al., 2019), 'spatstat' (Baddeley et al., 2015), 'raster' (Hijmans, 2019), 'sf'

388 (Pebesma, 2018), 'lmerTest' (Kuznetsova et al., 2017), 'lme4' (Bates et al., 2015),

<sup>389</sup> 'effsize' (Torchiano, 2020) and 'piecewiseSEM' (Lefcheck, 2016).

390 Results

392 We found a significant positive SAR for total species richness in CAEs, but not in islands 393 (Fig. 3a). The best model describing the SAR in CAEs was the logarithmic model (see Appendix S6). The observed  $R^2$  values were relatively low compared to values usually 394 observed for islands but comparable to those found in continental areas (Kreft et al., 395 2008). Similar results emerged with rarefied richness (see Appendix S7), 'corrected' 396 species richness given the size of the species pool (see Appendix S8), genus richness (see 397 398 Appendix S9) and also after removing the effect of environmental variation (see Appendix S10). 399 400 The sensitivity analysis revealed that a minimum of 52 continent-islands pairs is needed 401 to observe a significant SAR with our data. This number dropped to 16 when using mixed models with continent (a surrogate for the species pool) as a random effect (see Appendix 402 403 S11).

About half of the 1967 identified species belonged to the motile guild (see Appendix
S12), followed by the high profile and low-profile guilds. Planktonic species and species
with variable guilds represented a minor part of the communities. Species richness within
all guilds was significantly and positively related to area in CAEs (Fig. 3b-e) and this
relationship persisted for all but the high-profile guild after controlling for the
environment (see Appendix S10).

410 Comparison of species richness of islands and CAEs

411 Overall, species richness was significantly lower in the islands than in the respective

412 CAEs for more than 50% of all continent-island pairs (N = 96) (Fig. 5a). Similar results

emerged for the guilds, especially for the planktonic guild, where over 70% of the
comparisons had significantly higher species richness in CAEs. The only exception was
the low-profile guild whose species richness tended to be higher in islands (ca. 60%).
Importantly, however, when environmental variation was accounted for, the species
richness differences between CAEs and islands disappeared in more than 80% of cases
(Fig. 5b).

## 419 Ecological variables driving species richness

In islands, we found significant relationships between species richness and isolation for 420 total, low profile species richness (U-shaped pattern) and planktonic guild (negative 421 422 linear pattern and a weak non-linear pattern) (Fig. 4, see Appendix S13). There was no 423 relationship between richness and age of island for the total community or any of the ecological guilds (see Appendix S13). Apart from isolation, total species richness in 424 425 islands was significantly related only to phosphorus concentration. Species richness of 426 guilds was also significantly related to some other environmental variables depending on the guild considered (see Appendix S13). 427

428 Total species richness in CAEs was significantly related to environmental heterogeneity,

429 pH, conductivity, phosphorus concentration, all climatic variables and longitude (see

430 Appendix S13). The piecewise SEM models (Lefcheck, 2016) disentangled the effects of

- 431 the influencing factors and demonstrated that diatom species richness in CAEs was
- 432 related to area, habitat heterogeneity, physicochemistry, elevation, and climate. The data
- 433 fitted well the *a priori* model (Fig. 6) for total species richness and species richness of all
- 434 ecological guilds. The marginal  $R^2$  (variance explained by the fixed effects only) for total
- 435 species richness was 0.71 and varied between 0.14 (for low-profile species) to 0.59 (for

436 motile species). In the SEMs, area explained species richness independently, without any indirect effect through environmental heterogeneity, except for the motile species 437 richness where the effect of environmental heterogeneity was negative. Indeed, total 438 439 species richness and richness of the motile guild were mainly driven by area and temperature seasonality (Fig. 6a, d). Low-profile species richness was only explained by 440 441 area and precipitation (Fig. 6b). High-profile species richness was mainly explained by area, elevation and conductivity, while climate had no direct effect (Fig. 6c). Finally, 442 planktonic species richness was solely determined by elevation and was thus the only 443 444 group without a significant relationship with area.

#### 445 Discussion

Here, we conducted the first comparative analysis of island vs. mainland species-area 446 relationship for microbes, providing insight into the roles of area, environmental 447 448 heterogeneity, isolation and island age on species richness patterns. We showed for 449 freshwater diatoms that: (i) there was a significant SAR in continents but not in islands (except for high profile), (ii) regional species richness was higher in continents than in 450 451 islands, but this difference was explained entirely by environmental conditions (iii) the 452 effect of isolation varied among diatom guilds and (iv) area and median environmental conditions but not environmental heterogeneity were significant predictors of diatom 453 richness. Next, we will discuss the main findings in more detail and highlight our major 454 conclusions about total community and guild richness. 455

## 456 Drivers of species richness in islands

457 The lack of a significant SAR in islands may be due to low sample size (N = 18) or may

458 represent a real biogeographical pattern. Sensitivity analyses performed for CAEs

revealed that 16 islands are needed to detect a SAR given a common species pool. This
result is consistent with numerous studies on other organisms, reporting ISAR for a
relatively small number of islands within archipelagos (Matthews et al., 2019). However,
at a global scale, a much higher number of islands (N=52, Appendix S11) may be
required for detection of diatom ISAR.

The absence of diatom ISAR may have evolutionary and ecological causes. First, diatoms 464 may have distinct species pools across the globe (Soininen et al., 2016) and differences in 465 466 island area may not be sufficient to predict richness on islands that differ greatly in species pool. As the size of the species pool influences the shape of the SAR (Catano et 467 al., 2021), future analyses on archipelagos will be essential for determining whether 468 469 ISAR exists for diatoms (but see Jüttner et al., 2018). Second, environmental heterogeneity, which increased with island size (Fig. 3, and see Appendix S5) and is 470 471 recognized as an important driver of SAR (Lack, 1976; Chase et al., 2019), had no direct impact on island species richness. Third, island richness was related only to total 472 phosphorus, suggesting that productivity is a key factor explaining island diatom species 473 richness at this scale. Note however, that due to data availability, only phosphorus 474 concentration was considered as a resource factor for explaining species richness. The 475 476 consideration of other nutrient resources, known to influence diatom diversity (e.g. nitrogen, iron, Passy, 2007, Soininen, 2007), may improve the understanding of diatoms 477 species richness in islands. 478

We found that isolation might have some effect on species richness in islands. Two of the
most isolated islands (Oahu and New Caledonia) actually showed high species richness,
resulting in a U-shaped relationship between species richness and isolation for total and

482 low-profile species richness. Oahu and New Caledonia still had the highest species richness when the latter is corrected by species pool but the U-shaped relationship is only 483 marginally significant (p = 0.09, see Appendix S8). Greater speciation in the most 484 485 isolated islands, which have many endemic species, e.g. New Caledonia has been dubbed "Galapagos of diatoms" (Moser et al., 1998), may explain their higher richness 486 487 considering that endemic and total species richness are typically correlated (Kallimanis et 488 al., 2010). We could, however, not exclude the fact that some other unmeasured environmental factors, particularly related to islands conditions, may also be responsible 489 490 for this pattern. Finally, our finding further suggests that the biogeographical drivers of diatom richness on real islands are trait dependent. 491

Following Rosenzweig (1995), we hypothesized that islands would harbour lower species richness than continents due to diminished dispersal and rescue effects. While species richness was indeed lower in islands compared to continents, this difference disappeared when we accounted for environmental differences. Thus, annual precipitation, higher in islands than continents, was associated with lower species richness (see Appendix S10), likely because of its positive effect on current velocity, and subsequently, shear stress (Heino & Soininen, 2007).

#### 499 Drivers of species richness in continents

500 We tested whether SARs in continents could result from passive sampling and

environmental heterogeneity, which are major drivers of the SAR (Lack, 1976;

502 Rosenzweig, 1995; Stein et al., 2014; Chase et al., 2019). Surprisingly, species richness

in continents was not directly explained by environmental heterogeneity in the SEM.

504 Although area was strongly related to environmental heterogeneity (but poorly related to other environmental variables, see Appendix S5), none of the SEM models showed a 505 direct effect of habitat heterogeneity on either total or guild species richness, except for 506 motile species. For the latter, the direct effect of habitat heterogeneity was negative. 507 contrary to the results observed in univariate regressions (see Appendix 13) due to the 508 509 strong collinearity between area and heterogeneity. Given that we still observed a significant SAR with species richness estimated from the rarefaction curves, passive 510 sampling cannot be completely ruled out (Chase et al., 2019). However, the impact of 511 512 area on richness in continents might also be due to disproportionate effects, including dispersal, extinction and speciation. While extinction and speciation have been less 513 514 studied in diatoms, dispersal and mass effects (whereby species maintain their presence 515 in unfavorable conditions via immigration, Shmida & Wilson, 1985) were shown to have a notable influence on regional to subcontinental diatom communities (Soininen, 2007; 516 Jamoneau et al., 2018; Leboucher et al., 2020). For continental diatoms, larger areas may 517 thus increase the probability of immigration from the surrounding landscape, particularly 518 for species with high dispersal capabilities (mass-effect species), thereby increasing 519 520 CAE's diversity.

Environmental factors, such as nutrients, climate and elevation, were also important predictors of total and guild species richness. Total species richness decreased with temperature seasonality, as did the species richness of motile species, which represented ca. 50% of the whole community (see Appendix S12). As motile species are generally warm-water species (Pound et al., 2021) and high seasonality occurs in colder areas, it is possible that motile guild richness was limited by unfavorable temperatures. Species

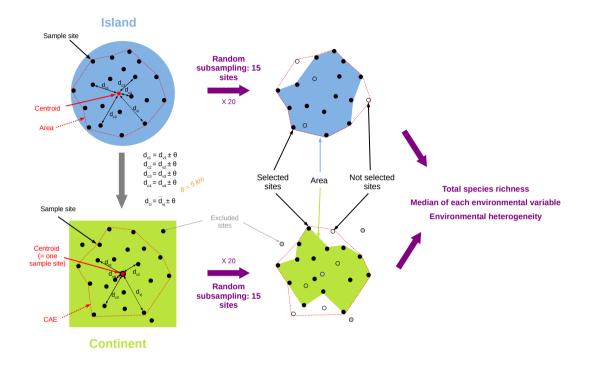
richness of the high profile and planktonic guilds was the lowest at high elevation. For
high-profile species, high elevation is stressful due to increased current velocity and
probability for dislodgement. For planktonic species, high elevations do not provide
sufficient habitat, given that these species require large water bodies. Species richness of
the low-profile guild is positively influenced by annual precipitation probably because
this guild is tolerant to physical disturbance (Passy, 2007), which should increase its
richness in the community.

#### 534 Conclusions

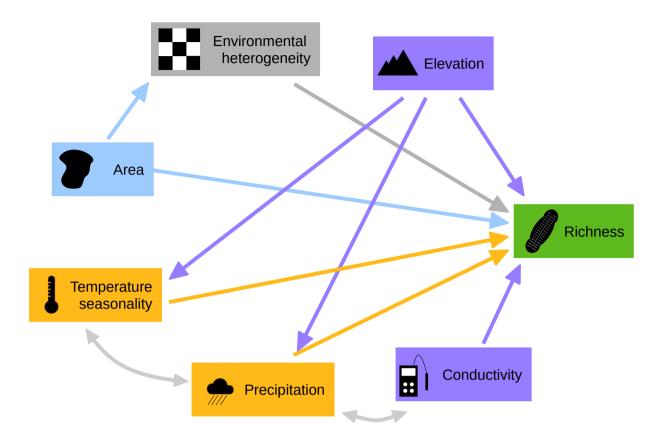
We examined diatom ISARs and compared them with the SARs of similarly sized 535 continental area equivalent across five continents. Contrary to most previous studies, we 536 did not find significant ISAR for total species richness but detected significant 537 relationships of richness with total phosphorus. These results imply that diatom richness 538 539 in islands is not related to area but is controlled by productivity. However, the lack of 540 ISAR may be due to distinct species pool across islands in our study. Species richness was typically higher in continental areas than in similarly sized islands, most probably 541 due to differences in climate and related environmental conditions, such as current 542 velocity. The significant SAR for continents may originate from disproportionate effects, 543 such as mass effect, but not from environmental heterogeneity. Isolation influenced the 544 richness of the whole community and some diatom guilds in islands. These finding 545 indicate that there are important differences in richness responses to island properties 546 547 among ecological guilds and between the community level and the functional level. Finally, the proposed new method for species-area comparisons between islands and 548 continental area equivalents will advance research on biogeography of islands vs. 549

- 550 mainland. We advocate obtaining global diatom data, particularly from archipelagos to
- 551 better understand the drivers of island species diversity.

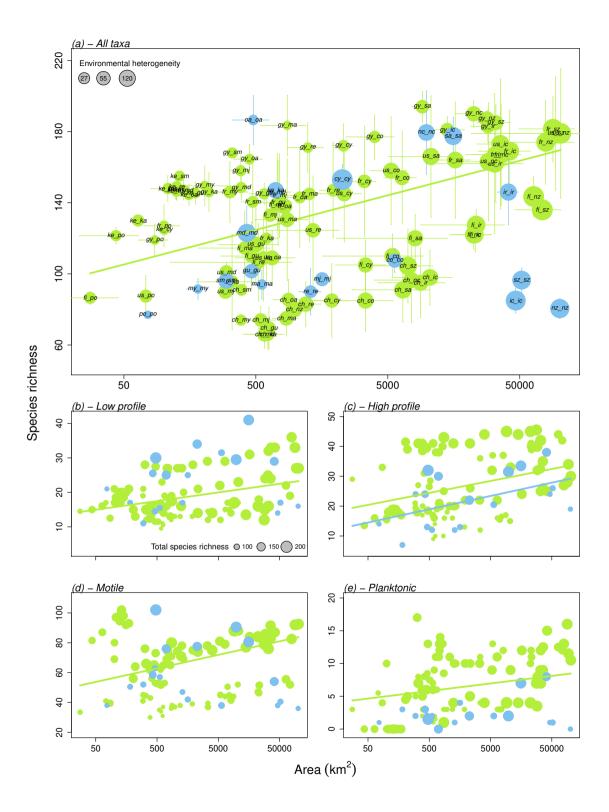
## 552 Figures legend



- 554 Figure 1: Descriptive diagram of the methods. Diagram describing the methodological
- 555 process used for creation of continental area equivalents (CAE) and subsampling of both
- 556 islands and CAEs.



- 558 Figure 2. A priori model explaining diatom species richness. Species richness is
- 559 modeled as a function of area, environmental heterogeneity, local environmental
- 560 conditions (purple) and climate (orange).



562 Figure 3. Species-area relationships for continents and islands. Species-area relationships for continental area equivalents (CAE) (in green, N=96) and islands (in 563 blue, N=18) for total species richness (a), and richness of low profile (b), high profile (c), 564 motile (d) and (e) planktonic species. Green regression lines represent significant linear 565 fits in mixed models for CAEs: richness = 8.44x + 72.24,  $R^2_m = 0.22$  for total species 566 richness, 1.09x + 10.69,  $R_m^2 = 0.13$  for low-profile species, 1.74x + 13.61,  $R_m^2 = 0.11$  for 567 high-profile species, 3.94x + 38.36,  $R^2_m = 0.11$  for motile species and 0.50x + 2.69,  $R^2_m =$ 568 569 0.05 for planktonic species, where x = log(area). The blue regression line represents significant linear fit for high-profile species of islands: 1.93x + 6.95,  $R^2_{ai} = 0.23$ . Dot 570 sizes are proportional to environmental heterogeneity (in log) for all taxa (a) and 571 proportional to total species richness for functional groups (b-e). Error bars represent 572 standard deviation estimated from the subsampling procedure. Text in dots indicate the 573 dataset used for computing species richness and area. For example, 'fr my' indicates the 574 position of Mayotte CAE in France. Continental datasets are indicated by 'fr' for France, 575 'us' for US, 'fi' for Finland, 'ch' for China, 'ke' for Kenya and 'gy' for French Guiana 576 and islands indicated by 'ic' for Iceland, 'co' for Corsica, 'gu' for Guadeloupe, 'ma' for 577 Martinique, 're' for La Réunion, 'my' for Mayotte, 'nz' for North New Zealand, 'sz' for 578 579 South New Zealand, 'nc' for New Caledonia, 'ka' for Kauai, 'oa' for Oahu, 'po' for 580 Possession, 'cy' for Cyprus, 'ir' for Ireland, 'md' for Madeira, 'mj' for Majorca, 'sm' for São Miguel and 'sa' for Sardinia. 581

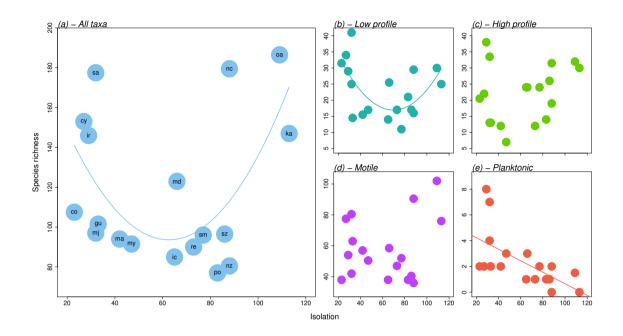
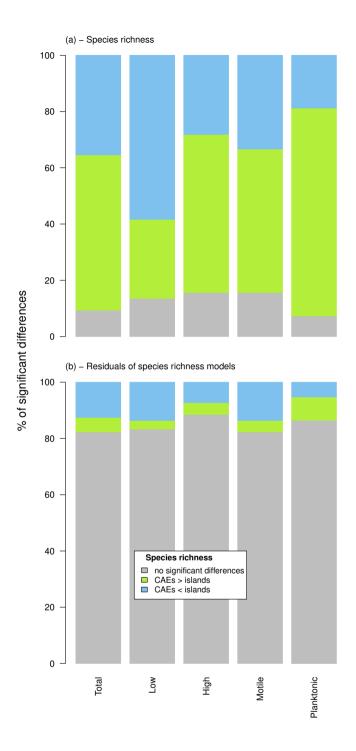
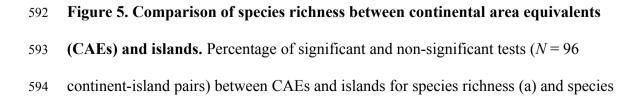


Figure 4. Relationships between island species richness and isolation. Relationship 583 between total species richness (a) and species richness of each ecological guild (b-e) with 584 585 island isolation for islands (N = 18). Significant linear and quadratic relationships (p<0.05) are shown by regression fits (only the fit with the lower AIC is shown if both 586 are significant, see Appendix S13):  $0.03x^2 - 3.79x + 212.32$ ,  $R^2_{aj} = 0.26$  for total richness, 587  $0.01x^2 - 0.96x + 50.44$ ,  $R^2_{aj} = 0.32$  for low-profile and -0.04x + 5.12,  $R^2_{aj} = 0.34$  for 588 planktonic species. For island names, see Fig. 1. Isolation is based on index defined by 589 Dahl (Dahl, 1991). 590





614	richness residuals (b). Tests were performed for total and guild species richness. Species
615	richness residuals were estimated from linear mixed models with species richness as the
616	dependent variable, and pH, conductivity, elevation, mean annual temperature and
617	precipitation, temperature and precipitation seasonality and environmental heterogeneity
618	as explanatory variables and continent as a random factor. Comparisons of values (i.e.,
619	species richness or residuals of species richness) were performed with Cliff's test,
620	whereby tests with delta >0.33 indicated significant differences (Romano et al., 2006).

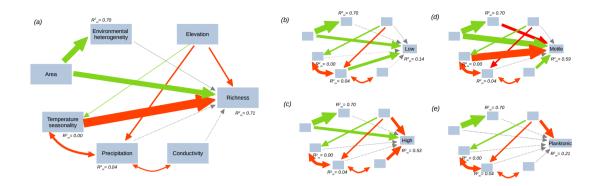


Figure 6. Structural equation models explaining species richness in continental area 623 equivalents (CAEs). Structural equation models for total species richness (a), low profile 624 625 (b), high profile (c), motile (d) and planktonic (e) species richness in continents (N = 96CAEs). Green and red arrows represent significant positive and negative relationships, 626 627 respectively, whereas gray-dashed arrows represent non-significant relationships. Arrow widths are proportional to the standardized regression coefficients and  $R_m^2$  values 628 represent marginal  $R^2$  from a linear mixed model. All models fitted well the *a priori* 629 model, i.e. the model including all shown causal relationships (Fisher's C = 14.99, df =630 14, p = 0.38 for all models). 631

## 632 Data Availability Statement

633 Data are available under the following link: https://doi.org/10.57745/ZPBSLT

634

#### 635 **References**

- Abonyi, A., Horváth, Z., & Ptacnik, R. (2018). Functional richness outperforms taxonomic richness in predicting ecosystem functioning in natural phytoplankton communities. *Freshwater Biology*, 63(2), 178–186.
- Algar, A. C., & Losos, J. B. (2011). Evolutionary assembly of island faunas reverses the classic island–mainland richness difference in Anolis lizards. *Journal of Biogeography*, 38(6), 1125–1137.
- Arrhenius, O. (1921). Species and area. The Journal of Ecology, 95–99.
- Baddeley, A., Rubak, E., & Turner, R. (2015). *Spatial Point Patterns: Methodology and Applications with R.* Chapman and Hall/CRC Press.
- Báldi, A. (2008). Habitat heterogeneity overrides the species–area relationship. *Journal of Biogeography*, 35(4), 675–681.
- Bates, D., Maechler, M., Bolker, B., Walker, S., Christensen, R. H. B., Singmann, H., Dai, B., & Grothendieck, G. (2015). *Ime4: Linear Mixed-Effects Models using "Eigen" and S4* (1.1-8) [Computer software].
- Bell, T., Ager, D., Song, J.-I., Newman, J. A., Thompson, I. P., Lilley, A. K., & Gast, C. J. van der. (2005). Larger Islands House More Bacterial Taxa. *Science*, 308(5730), 1884–1884.
- Catano, C. P., Grman, E., Behrens, E., & Brudvig, L. A. (2021). Species pool size alters species–area relationships during experimental community assembly. *Ecology*, *102*(1), e03231.

- Chao, A. (1987). Estimating the Population Size for Capture-Recapture Data with Unequal Catchability. *Biometrics*, *43*(4), 783–791.
- Chase, J. M., Gooriah, L., & May, F. (2019). A framework for disentangling ecological mechanisms underlying the island species–area. *Frontiers of Biogeography*.
- Chase, J. M., & Leibold, M. A. (2003). *Ecological Niches: Linking Classical and Contemporary Approaches*. University of Chicago Press.
- Connor, E. F., & McCoy, E. D. (1979). The Statistics and Biology of the Species-Area Relationship. *The American Naturalist*, *113*(6), 791–833.
- Dahl, A. L. (1991). *Island Directory* (UNEP Regional Seas Directories and Bibliographies No. 35; p. 573). UNEP.
- Danielson, J. J., & Gesch, D. B. (2011). Global multi-resolution terrain elevation data 2010 (GMTED2010) (U.S. Geo - Logical Survey Open-File Report No. 2011– 1073; p. 26).
- DeMalach, N., Saiz, H., Zaady, E., & Maestre, F. T. (2019). Plant species–area relationships are determined by evenness, cover and aggregation in drylands worldwide. *Global Ecology and Biogeography*, 28(3), 290–299.
- Drakare, S., Lennon, J. J., & Hillebrand, H. (2006). The imprint of the geographical, evolutionary and ecological context on species–area relationships. *Ecology Letters*, 9(2), 215–227.
- Farina, A. (2008). Principles and Methods in Landscape Ecology: Towards a Science of the Landscape. Springer.
- Finlay, B. J. (2002). Global Dispersal of Free-Living Microbial Eukaryote Species. Science, 296(5570), 1061–1063.
- Gillespie, R. G., Claridge, E. M., & Roderick, G. K. (2008). Biodiversity dynamics in isolated island communities: Interaction between natural and human-mediated processes. *Molecular Ecology*, 17(1), 45–57.

- Hanski, I. A., & Gaggiotti, O. E. (2004). Ecology, Genetics and Evolution of Metapopulations. Elsevier Academic Press.
- Heino, J., & Soininen, J. (2005). Assembly rules and community models for unicellular organisms: Patterns in diatoms of boreal streams. *Freshwater Biology*, 50(4), 567–577.
- Heino, J., & Soininen, J. (2006). Regional occupancy in unicellular eukaryotes: A reflection of niche breadth, habitat availability or size-related dispersal capacity? *Freshwater Biology*, 51(4), 672–685.
- Heino, J., & Soininen, J. (2007). Are higher taxa adequate surrogates for species-level assemblage patterns and species richness in stream organisms? *Biological Conservation*, 137(1), 78–89.

Hijmans, R. J. (2019). raster: Geographic Data Analysis and Modeling.

- Hijmans, R. J., Cameron, S. E., Parra, J. L., Jones, P. G., & Jarvis, A. (2005). Very high resolution interpolated climate surfaces for global land areas. *International Journal of Climatology*, 25(15), 1965–1978.
- Jamoneau, A., Passy, S. I., Soininen, J., Leboucher, T., & Tison-Rosebery, J. (2018).
  Beta diversity of diatom species and ecological guilds: Response to environmental and spatial mechanisms along the stream watercourse. *Freshwater Biology*, 63(1), 62–73.
- Jüttner, I., Vijver, B. V. de, Williams, D. M., Lange-Bertalot, H., & Ector, L. (2018). The genus Eunotia (Bacillariophyta) in the Falkland Islands and species-area relationships in sub-Antarctic islands. *Diatom Research*, 33(4), 413–452.
- Kallimanis, A. S., Bergmeier, E., Panitsa, M., Georghiou, K., Delipetrou, P., & Dimopoulos, P. (2010). Biogeographical determinants for total and endemic species richness in a continental archipelago. *Biodiversity and Conservation*, *19*(5), 1225–1235.

- Kalmar, A., & Currie, D. J. (2006). A global model of island biogeography. *Global Ecology and Biogeography*, 15(1), 72–81.
- Kalmar, A., & Currie, D. J. (2007). A Unified Model of Avian Species Richness on Islands and Continents. *Ecology*, 88(5), 1309–1321.
- Krause, S., Le Roux, X., Niklaus, P. A., Van Bodegom, P. M., Lennon, J. T., Bertilsson,
  S., Grossart, H.-P., Philippot, L., & Bodelier, P. L. E. (2014). Trait-based
  approaches for understanding microbial biodiversity and ecosystem functioning. *Frontiers in Microbiology*, 5.
- Kreft, H., Jetz, W., Mutke, J., Kier, G., & Barthlott, W. (2008). Global diversity of island floras from a macroecological perspective. *Ecology Letters*, 11(2), 116–127.
- Kuznetsova, A., Brockhoff, P. B., & Christensen, R. H. B. (2017). ImerTest Package: Tests in Linear Mixed Effects Models. *Journal of Statistical Software*, 82(1), 1–26.
- Lack, D. (1976). *Island biology: Illustrated by the land birds of Jamaica*. Blackwell Scientific Publications, Oxford, UK.
- Leboucher, T., Tison-Rosebery, J., Budnick, William. R., Jamoneau, A., Vyverman, W., Soininen, J., Boutry, S., & Passy, S. I. (2020). A new metacommunity approach for detecting species influenced by mass effect. *Journal of Applied Ecology*.
- Lecointe, C., Coste, M., & Prygiel, J. (1993). "Omnidia": Software for taxonomy, calculation of diatom indices and inventories management. *Hydrobiologia*, *269–270*(1), 509–513.
- Lefcheck, J. S. (2016). piecewiseSEM: Piecewise structural equation modelling in r for ecology, evolution, and systematics. *Methods in Ecology and Evolution*, 7(5), 573–579.
- Lomolino, M. V., & Weiser, M. D. (2001). Towards a More General Species-Area Relationship: Diversity on All Islands, Great and Small. *Journal of Biogeography*, 28(4), 431–445. JSTOR.

- MacArthur, R. H., & Wilson, E. O. 1967. (1967). *The theory of island biogeography*. Princeton University Press.
- Matthews, T. J., Cottee–Jones, H. E., & Whittaker, R. J. (2014). Habitat fragmentation and the species–area relationship: A focus on total species richness obscures the impact of habitat loss on habitat specialists. *Diversity and Distributions*, 20(10), 1136–1146.
- Matthews, T. J., Rigal, F., Triantis, K. A., & Whittaker, R. J. (2019). A global model of island species–area relationships. *Proceedings of the National Academy of Sciences*, 116(25), 12337–12342.
- Mittelbach, G. G., & McGill, B. J. (2019). Community Ecology. Oxford University Press.
- Moser, G., Lange-Bertalot, H., & Metzeltin, D. (1998). Insel der Endemiten. Geobotanisches Phänomen Neukaledonien—Island of Endemics. New Caledonia a geobotanical phenomenon (J. Cramer). Schweizerbart'sche Verlagsbuchhandlung.
- Oksanen, J., Blanchet, F. G., Friendly, M., Kindt, R., Legendre, P., McGlinn, D., Minchin, P. R., O'Hara, R. B., Simpson, G. L., Solymos, P., Stevens, M. H. H., Szoecs, E., & Wagner, H. (2019). *vegan: Community Ecology Package*.
- Passy, S. I. (2007). Diatom ecological guilds display distinct and predictable behavior along nutrient and disturbance gradients in running waters. *Aquatic Botany*, 86(2), 171–178.
- Passy, S. I. (2016). Abundance Inequality in Freshwater Communities Has an Ecological Origin. *The American Naturalist*, 187(4), 502–516.
- Peay, K. G., Bruns, T. D., Kennedy, P. G., Bergemann, S. E., & Garbelotto, M. (2007). A strong species–area relationship for eukaryotic soil microbes: Island size matters for ectomycorrhizal fungi. *Ecology Letters*, 10(6), 470–480.
- Pebesma, E. (2018). Simple Features for R: Standardized Support for Spatial Vector Data. *The R Journal*, *10*(1), 439–446.

- Pound, K. L., Larson, C. A., & Passy, S. I. (2021). Current distributions and future climate-driven changes in diatoms, insects and fish in U.S. streams. *Global Ecology and Biogeography*, 30, 63–78.
- Prugh, L. R., Hodges, K. E., Sinclair, A. R., & Brashares, J. S. (2008). Effect of habitat area and isolation on fragmented animal populations. *Proceedings of the National Academy of Sciences*, 105(52), 20770–20775.
- R Core Team. (2019). *R: A Language and Environment for Statistical Computing*. R Foundation for Statistical Computing.
- Reche, I., Pulido-Villena, E., Morales-Baquero, R., & Casamayor, E. O. (2005). Does Ecosystem Size Determine Aquatic Bacterial Richness? *Ecology*, 86(7), 1715– 1722.
- Rimet, F., & Bouchez, A. (2012). Life-forms, cell-sizes and ecological guilds of diatoms in European rivers. *Knowledge and Management of Aquatic Ecosystems*, 406, 01.
- Romano, J., Kromrey, J. D., Coraggio, J., & Skowronek, J. (2006). Appropriate statistics for ordinal level data: Should we really be using t-test and Cohen's d for evaluating group differences on the NSSE and other surveys. *Annual Meeting of the Florida Association of Institutional Research*, 1–33.
- Rosenzweig, M. L. (1995). *Species diversity in space and times*. Cambridge: Cambridge University Press.
- Sandin, S. A., Vermeij, M. J. A., & Hurlbert, A. H. (2008). Island biogeography of Caribbean coral reef fish. *Global Ecology and Biogeography*, 17(6), 770–777.
- Schrader, J., König, C., Triantis, K. A., Trigas, P., Kreft, H., & Weigelt, P. (2020). Species–area relationships on small islands differ among plant growth forms. *Global Ecology and Biogeography*, 29(5), 814–829.
- Shmida, A., & Wilson, M. V. (1985). Biological determinants of species diversity. *Journal of Biogeography*, 12(1), 1–20.

- Simberloff, D. S., & Wilson, E. O. (1969). Experimental Zoogeography of Islands: The Colonization of Empty Islands. *Ecology*, 50(2), 278–296.
- Smith, V. H., Foster, B. L., Grover, J. P., Holt, R. D., Leibold, M. A., & deNoyelles, F. (2005). Phytoplankton species richness scales consistently from laboratory microcosms to the world's oceans. *Proceedings of the National Academy of Sciences*, 102(12), 4393–4396.
- Soininen, J. (2007). Environmental and Spatial Control of Freshwater Diatoms—A Review. *Diatom Research*, *22*(2), 473–490.
- Soininen, J., Jamoneau, A., Rosebery, J., & Passy, S. I. (2016). Global patterns and drivers of species and trait composition in diatoms. *Global Ecology and Biogeography*, 25(8), 940–950.
- Stein, A., Gerstner, K., & Kreft, H. (2014). Environmental heterogeneity as a universal driver of species richness across taxa, biomes and spatial scales. *Ecology Letters*, 17(7), 866–880.
- Tecchio, S., Chaalali, A., Raoux, A., Tous Rius, A., Lequesne, J., Girardin, V., Lassalle, G., Cachera, M., Riou, P., Lobry, J., Dauvin, J.-C., & Niquil, N. (2016).
  Evaluating ecosystem-level anthropogenic impacts in a stressed transitional environment: The case of the Seine estuary. *Ecological Indicators*, *61, Part 2*, 833–845.
- Teittinen, A., & Soininen, J. (2015). Testing the theory of island biogeography for microorganisms patterns for spring diatoms. *Aquatic Microbial Ecology*, 75(3), 239–250.
- Tilman, D., Knops, J., Wedin, D., Reich, P., Ritchie, M., & Siemann, E. (1997). The Influence of Functional Diversity and Composition on Ecosystem Processes. *Science*, 277(5330), 1300–1302.

Torchiano, M. (2020). effsize: Efficient Effect Size Computation.

- Triantis, K. A., Economo, E. P., Guilhaumon, F., & Ricklefs, R. E. (2015). Diversity regulation at macro-scales: Species richness on oceanic archipelagos. *Global Ecology and Biogeography*, 24(5), 594–605.
- Whittaker, R. J., & Fernandez-Palacios, J. M. (2007). *Island Biogeography: Ecology, Evolution, and Conservation*. OUP Oxford.
- Whittaker, R. J., Fernández-Palacios, J. M., Matthews, T. J., Borregaard, M. K., & Triantis, K. A. (2017). Island biogeography: Taking the long view of nature's laboratories. *Science*, 357(6354).
- Whittaker, R. J., Triantis, K. A., & Ladle, R. J. (2008). A general dynamic theory of oceanic island biogeography. *Journal of Biogeography*, *35*(6), 977–994.

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## 637 Biosketch

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