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Allelopathic interactions between phytobenthos and meiofaunal community in an Adriatic benthic ecosystem: Understanding the role of aldehydes and macroalgal structural complexity

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1 2	Allelopathic interactions between phytobenthos and meiofaunal community in an Adriatic benthic ecosystem: understanding the role of aldehydes and macroalgal
- 3 4	structural complexity
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25 Abstract

Macroalgae produce several allelopathic substances, including polyunsaturated aldehydes (PUAs), which may inhibit photosynthesis and growth rates of other algal species, and grazing. Additionally, macroalgal structural complexity is an important factor in determining abundance patterns and size structure of epiphytic organisms.

In this study the PUAs production of two Mediterranean macroalgae, *Dictyopteris polypodioides*, 30 31 (DP, Phaeophyceae, Dictyotales) and Cystoseira compressa (CC, Phaeophyceae, Fucales), was characterized to clarify the relationships between the meiobenthic and microphytobenthic 32 communities. Results showed a higher PUAs production and a diverse qualitative profile for DP, 33 34 which reported long-chain compounds (i.e. C14-C16) as main aldehydes, than CC, with the short-35 chain C6:2 as main compound, as well as variability among sampling times. A clear separation of the meiofauna and microphytobenthos assemblages was found for the macroalgae, but with different 36 37 temporal trends. Dissimilarities were due to five microalgal orders, namely Naviculales, Lyrellales, Gonyaulacales (i.e. Ostreopsis), Bacillariales, and Licmophorales, and to the meiofaunal groups 38 39 nematodes, copepods, and copepod nauplii, which were more abundant on DP than on CC. Results indicate that macroalgal complexity is a major determinant of the meiofaunal community structure 40 41 (accounting to 26% of the variation), rather than PUAs production itself (17%). PUAs effects seem 42 species-specific, thus affecting some grazers instead of the entire community. Conversely, microphytobenthos affected the meiofauna assemblages, particularly harpacticoids, confirming the 43 role of these organisms as primary food source of all marine food chain producers. Since PUAs are 44 45 produced also by several epiphytic diatoms, the understanding of their effects on the community structure and on the relationships among taxa in the field is complicated and requires further in-depth 46 investigations in simplified systems (i.e. microcosms). 47

48

49 Keywords

50 Brown macroalgae; PUAs; microalgae; meiofauna; harpacticoid copepods; chemical ecology.

52 1. Introduction

53 Chemical compounds produced by aquatic organisms and microorganisms have received increasing 54 attention for the important role that may have in modulating the interactions in marine ecosystems 55 (Paffenhöfer et al., 2005; Ianora et al., 2012). Many organisms produce allelochemicals, secondary 56 metabolites that affect survival and growth of other species and may give a competitive advantage to 57 their producers (Legrand et al., 2003; Fistarol et al., 2004; Tillmann, 2004; Allen et al., 2016).

Several studies regarding chemical interactions between various marine organisms, including
copepods, urchins, sea stars, algae, mollusks, and polychaetes, were focused on direct responses of
target species to allelochemicals and have been performed mainly under laboratory conditions (e.g.
Ianora et al., 2004; Adolph et al., 2004; Caldwell et al., 2004; Vardi et al., 2006 Ribalet et al., 2007;
Taylor et al., 2007).

63 Algae produce a variety of these allelopathic substances, such as phenolic compounds, alkaloids, peptides, oxoacids, and polyketides, among which there are toxins synthesized by several species 64 65 (Snyder et al., 2003; Agostini-Costa et al., 2012; Pistocchi et al., 2012), polyunsaturated fatty acid (PUFAs) (Grima et al., 1995; Tonon et al., 2002; Patil et al., 2007) and their derivatives, such as 66 67 polyunsaturated aldehydes (PUAs) (Wichard et al., 2005; Ianora et al., 2012; Pezzolesi et al., 2017). 68 Consequences of allelopathic interactions between different algal species can be of various nature and magnitude, including loss of motility, cell deformation (Tang and Gobler, 2010; Pichierri et al., 69 2017), pigmentation loss, cytoplasm aggregation, formation of vesicles and cellular lysis (Fistarol et 70 71 al., 2004). However, the majority of allelochemicals have milder effects, for example inhibition of photosynthesis, reduction of growth rate, and inhibition of grazing (Legrand et al., 2003). 72

The effect of PUAs have been mostly investigated for planktonic microalgae and only a few studies have evidenced PUAs production by benthic species, such as diatoms (Jüttner et al., 2010; Scholz and Liebezeit, 2012; Pezzolesi et al., 2017), and macroalgae (Kajiwara et al., 1996; Akakabe et al., 2003; Pezzolesi et al., 2021). Macroalgae are among the most important components of marine coastal

ecosystems because they are highly productive (Pinckney and Zingmark, 1993), have a high 77 78 taxonomic diversity, and may act as foundation species providing habitat for different organisms 79 (Cacabelos et al., 2010). The shape and structural complexity of macroalgae are important factors in determining the abundance patterns and size structure of the epiphytic organisms (McAbendroth et 80 81 al., 2005). The more structurally complex macroalgal species show abundant and various populations of invertebrates because they provide greater surface area for the colonization of epifaunal 82 83 assemblages and epiphytic microalgae (Chemello and Milazzo, 2002). Among epifaunal assemblages, macrofauna is the most investigated, while meiofaunal communities are overlooked. 84 Meiofauna represent the most abundant and taxonomically diversified metazoans on Earth (Giere, 85 86 2009) and on hard substrates can overcome macrofauna in terms of abundance (Gibbons and Griffiths, 1986). Meiofauna plays a dominant role in the exchange of organic matter (Sandulli et al., 2014; 87 Semprucci et al., 2016) as part of the "small food web" (size class 45-1000 µm). Moreover, it supports 88 89 most of the higher trophic levels (Giere, 2009), being an important food resource for macrofauna, small fish, juveniles and other epibenthic predators (Chardy and Dauvin, 1992). In phytal 90 91 environments, harpacticoid copepods are the dominant meiofauna group (Hicks, 1977; Coull and Wells, 1983; Hall and Bell, 1993) and show high diversity (Sarmento and Santos, 2012). They feed 92 mainly on diatoms and, for this reason, they have a high content of fatty acids and play a key 93 94 nutritional role for fish, carnivorous crustaceans (prawns and their larvae), and polychaetes (Coull, 1999; Giere, 2009). 95

Indeed, few studies have focused on the benthic environment and included the potential role of macroalgae both in terms of production of allelochemical compounds and regulation of interactions between various organisms (Kajiwara et al., 1996; Akakabe et al., 2003), while only one laboratory study analyzed the effects of PUAs on a species of harpacticoid (i.e. *Tisbe holothuriae*) (Taylor et al., 2007). PUAs strongly impair the reproduction of various potential grazers in *in vitro* studies (Poulet et al., 1994; Ianora et al., 2004b), while in the field the relationship between aldehyde production and reproductive failure of other higher-level organisms (i.e. copepods) remains unclear (Leflaive and Ten-Hage, 2009). Some studies performed to investigate the allelopathic activity of fresh thalli or extracts of *Ulva* spp. reported inhibitory effects on the growth of several microalgae, including harmful species (Tang and Gobler, 2011). The production of aldehydes and their potential consequences on biotic interactions may thus be explored, especially in macroalgae. This would help to determine whether these compounds, which are present across phylogenetic and environmental barriers, may play a role in the ecology of their producers (Leflaive and Ten-Hage, 2009), perhaps affecting the development of epiphytic organisms, such as meiofauna or microalgae.

The main aims of this study were to i) analyze the qualitative and quantitative production of aldehydes by two different macroalgae over a period of some months (Mediterranean spring and summer); ii) understand the relationship between meiobenthic communities, with particular interest to harpacticoid copepods, and the microphytobenthos present on the two macroalgal species, considering the structural complexity of macroalgae and the potential role of aldehydes in regulating their interaction.

Specifically, this study was performed in a benthic environment of the North-western Adriatic Sea 116 (Piscinetta del Passetto, Ancona, Italy), characterized by the presence of a rich phytobenthic 117 assemblage and by annual blooms of a toxic dinoflagellate, Ostreopsis cf. ovata Fukuyo, 1981. Two 118 macroalgal species, Cystoseira compressa (Esper) Gerloff & Nizamuddin and Dictyopteris 119 polypodioides A.P. De Candolle J.V. Lamouroux, commonly present in this site (Rindi et al., 2020) 120 and representative of the Mediterranean phytobenthic community were selected, based on their 121 different complexity and PUAs composition (Pezzolesi et al., 2021). To our knowledge, this is the 122 first study where macroalgal allelochemistry has been investigated to understand both the 123 microphytobenthos and meiofauna community structure and their interactions, in relation to the algal 124 complexity. 125

126

127 2 Material and methods

128 *2.1 Sampling area and sampling procedure*

The study was performed in a semi-enclosed and shallow (mean depth 1.5 m) inlet called Piscinetta del Passetto (Conero Riviera, Italy, northern Adriatic Sea: 43°37'09" N, 13°31'54" E), described and showed in Pezzolesi et al. (2021). Sampling was carried out at six different times from May to September 2018 (see table S2), with monthly frequency except for September (when sampling was performed twice). In the Mediterranean area, this period corresponds to late spring and summer. Surface temperature and salinity were measured with a multiparameter water probe HQ30d (Hach-Lange GmbH) and a refractometer Atago S-10, respectively.

Apical parts (i.e. tips) of the thalli (first 5-8 cm) of two macroalgae, i.e., *Cystoseira compressa* (Phaeophyceae, Fucales) and *Dictyopteris polypodioides* (Phaeophyceae, Dictyotales), were sampled at a depth of approximately 0.5 m. At the study site, during the time of the year in which sampling was carried out, these species are present and mostly occur in a well-developed habit. For each macroalgal species, six replicates were collected by snorkeling at each sampling time using 50 mL polypropylene tubes (VWR International), avoiding the dispersion of the associated epiphytic organisms.

Water samples for assessment of dissolved aldehydes and nutrient analysis were collected in 2 L 143 144 polyethylene bottles (VWR International) in the proximity of the sampled macroalgae. Water was subsequently filtered using GF/F Whatman filters (0.7 µm porosity, 47 mm) and stored at -22 °C until 145 analysis. Macroalgae were treated to remove all associated benthic organisms as described in 146 Pezzolesi et al. (2021) using filtered seawater. Briefly, each tube containing the thallus and their 147 148 storage water was vigorously shaken to separate the macroalga from the epiphytic microalgae and the meiofauna. Then the tube was rinsed with filtered seawater and vigorously washed several times until 149 150 epiphytic organisms were completely removed. The seaweed thalli were dried with absorbent paper, then weighed to determine fresh weight (g fw) and photographed; finally, they were stored at -80 °C 151 in new tubes. The total volume of washing seawater of each sample (approximately 150 mL) was 152 measured and then divided into two aliquots, one for the microphytobenthos and the other for the 153

meiofauna analyses. Aliquots for microphytobenthos (about 75 mL) were fixed with Lugol and stored 154 155 in 250 ml dark glass bottles. Aliquots for meiofauna analysis were sieved through a 1000-µm mesh and a 45-µm mesh; the meiobenthic organisms retained on the finer mesh sieve were fixed with 70% 156 alcohol and stored in a 50 ml falcon until subsequent analyses. Each apical part of macroalga was 157 placed on a white surface with a reference scale and then photographed using a digital camera Canon 158 EOS 750D. To take into account the different morphology and complexity of the two algal species, 159 pictures were processed using the program ImageJ v1.53u. For each image the scale was fixed and 160 transformed to binary format with a pixel width of 1 cm. Then the area, the perimeter, and the fractal 161 complexity (D) were measured based on the image and using a method analogous to the grid method 162 163 (boundary dimension) proposed by Sugihara and May (1990). In this method, the fractal dimension is the slope of the linear fit of $\log N(s)$ versus $\log (1/s)$; where s represents the scale of analysis and 164 N(s) is the number of objects observed at that scale. 165

166

167 *2.2 Aldehydes*

168 2.2.1 Aldehydes (PUAs) produced by macroalgae

The extraction and quantification of PUAs produced by the different macroalgae was carried out as 169 described in Pezzolesi et al. (2021) by gas chromatography-mass spectrometry (GC-MS). 170 Specifically, a portion of the apical part of the thallus (about 0.2-0.8 g f. wt.) was shredded with 171 mortar and pestle, in liquid nitrogen. The powder thus obtained was transferred into 10 mL tubes. 172 aldehydes performed with Derivatization of the polyunsaturated was O-(2,3,4,5,6-173 pentafluorobenzyl)hydroxylamine hydrochloride solution (PFBHA HCl) and quantification was 174 based on the internal standard (i.e. benzaldehyde). All reagents were purchased from Sigma-Aldrich 175 (Milan, Italy) and used without any further purification. 176

177

178 2.2.2 Dissolved aldehydes (dPUAs) in seawater

Dissolved aldehyde concentration was determined following the protocol described by Vidoudez and Pohnert (2008) with slight modifications, using PFBHA HCl in Tris–HCl 100 mM pH 7.2 as derivatizing reagent and benzaldehyde as internal standard, as described in Pezzolesi et al. (2021).

182

183 *2.3 Nutrients*

184 Nitrate and phosphate analyses were performed on filtered seawater aliquots (GF/F Whatman

185 filters) and analyzed spectrophotometrically according to Strickland and Parsons (1972).

186

187 *2.4 Microphytobenthic community*

188 The quali-quantitative analysis of the microphytobenthos associated with the two macroalgae was performed using an inverted optical microscope (Zeiss Axiovert 100) at 320x and 200x magnification. 189 Sub-samples (3-5 ml) of epiphytic communities fixed with Lugol were settled in counting chambers 190 191 after homogenization, according to the Utermöhl's sedimentation method (Utermöhl, 1958; Edler and Elbrächter, 2010). Counting was performed in different ways. The microphytobenthos community 192 was examined at 320x magnification on 30 random fields or 4-5 transects; then a counting at 200x 193 of the organisms present on the whole sedimentation chamber was performed to obtain a correct 194 evaluation of uncommon taxa. The microphytobenthos was identified and recognized following 195 196 various manuals and identification keys (e.g. Tomas, 1997; Kraberg et al., 2010). The identification of individuals was based exclusively on observable morphological characters (such as shape, size, 197 number of chloroplasts); the current taxonomic status for the microalgae was confirmed following 198 199 AlgaeBase (Guiry and Guiry, 2021).

200

201 *2.5 Meiofauna and harpacticoid copepods*

Meiofaunal organisms of each sample were counted and identified at higher taxonomic levels. All harpacticoids (Order Harpacticoida Sars G.O., 1903) were collected under a stereomicroscope (Nikon SMZ 1500) and stored in 70% alcohol inside 1.5 mL-Eppendorf labeled for subsequent identification. Harpacticoids were identified to species level using Lang (1948; 1965), and Boxshall and Halsey(2004).

207

208 2.6 Data analyses

PUAs produced by the macroalgae and dissolved in the seawater were analyzed and expressed as µg 209 g⁻¹ fw and µg L⁻¹, respectively. Microphytobenthos was analyzed in terms of total epiphytic cells and 210 main orders or genera; data were expressed as the number of cells per gram of fresh macroalgae 211 weight (cells g⁻¹ fw). Meiofauna and harpacticoid communities were analysed according to total 212 density (N), standardized towards the fresh weight of macroalgae (ind. g⁻¹ fw), the total number of 213 214 taxa or species (S), and Shannon diversity (H'). For each variable, the average value of the six replicates ± standard error is reported. All univariate variables were analyzed by a 2-way crossed 215 ANOVA; the factors considered were the macroalgal species (fixed, two levels: Cystoseira 216 217 compressa (CC) and Dictyopteris polypodioides (DP)) and time (fixed, 6 levels). Cochran's test was used to check for the homogeneity of variances and data were transformed, if necessary (Underwood, 218 219 1996). If even after transformation it was not possible to obtain homogeneity of the variances, untransformed data were analysed and results were considered robust if significant at p<0.01 to 220 compensate for the increased probability of type I error (Underwood, 1996). When significant main 221 222 effects or interactions were detected, the Student-Newman-Keuls (SNK) test was used for pairwise a posteriori comparisons. 223

The community structure of each assemblage (microphytobenthos, meiofauna and harpacticoids) was analyzed by non-metric multidimensional scaling (nMDS) based on Bray-Curtis similarity of square root-transformed data. Differences in community structures were assessed by permutational nonparametric multivariate analysis of variance (PERMANOVA) following the same experimental design adopted for ANOVA (Anderson, 2001; 2005). When significant main effects or interactions were detected, the specific procedure provided within PERMANOVA was used for pairwise a posteriori comparisons. The analyses were performed using unrestricted permutation of the raw dataand 9999 permutations.

Taxa that mostly contributed to the dissimilarity/similarity among/within macroalgal species and times were identified using the SIMPER analysis (Clarke, 1993).

Relationships between macroalgae complexity, PUAs, dominant microphytobenthic taxa, and both meiofauna and harpacticoid copepod assemblages were analyzed by the distance-based linear model

236 (DistLM) procedure in PERMANOVA+ (Anderson et al., 2008).

A total of 14 explanatory variables, grouped in three sets, were considered: macroalgae complexity 237 descriptors (perimeter/area, D, Area), PUA concentrations (C14:5, C16:4, C16:3, C6:2), and 238 239 microphytobenthic taxa abundances (Navicula spp. J.B.M. Bory de Saint-Vincent, 1822, Lyrella spp. N.I. Karayeva, 1978, Cocconeis spp. C.G. Ehrenberg, 1837, Cylindrotheca spp. L. Rabenhorst, 240 1859/Nitzschia spp. A.H. Hassall, 1845, other diatoms, Ostreopsis cf. ovata, other dinoflagellates). 241 242 The relationships between meiobenthic taxa data and harpacticoid abundance (square root transformed data) and the three groups of variables were analyzed by explicitly examining the 243 244 proportion of variation in the taxa data (or harpacticoid species abundance) that was explained by PUAs concentrations and microphytobenthic taxa abundances over and above the amount explained 245 246 by macroalgae complexity descriptors.

247 The Akaike Information Criterion with correction (AICc) was used to select the model. Prior to the DistLM, a draftsman plot and correlation matrices were performed to detect possible skewness of the 248 variables and/or strong correlation among pairs of variables (Anderson and Robinson, 2001). 249 Variables were not strongly correlated, so all variables were entered in the analysis. Concentrations 250 of Lyrella spp., and Cocconeis spp were square root transformed. Finally, Pearson correlation analysis 251 was used to test which macroalgal attribute (i.e. complexity descriptors such as D and macroalgal 252 surface (S), and PUAs composition) explained the variation of some representative genera of 253 microphytobenthos, of the main meiofauna taxa and of the most representative harpacticoid species. 254

Densities of microphytobenthic genera, main taxa and harpacticoid species were square roottransformed.

Significance level was set at 0.05 (5%) for all tests. ANOVA, Cochran test, SNK test, and correlations
were performed by R (version 3.5.3) using packages Lme4; all multivariate analyses were conducted
with PRIMER v7 (Clarke and Gorley, 2015) with the PERMANOVA + add-on (Anderson et al.,
2008).

261

262 *3.* **Results**

263 *3.1 PUAs production by macroalgae*

The interpretation of the chromatograms and relative mass spectra obtained by GC-MS gave the quali-quantitative profile of the main aldehydes produced by *Dictyopteris polypodioides* (DP) and *Cystoseira compressa* (CC).

The major aldehydes production was found in T1 (May) for both macroalgae, with values of 225.5 \pm 35.6 and 17.1 \pm 4.0 µg g⁻¹ fw for DP and CC, respectively. Conversely, the lowest amounts were measured at times T3 (July) for DP and T2 (June) for CC, with a concentration of 44.3 \pm 5.3 and 2.0 \pm 0.3 µg g⁻¹ fw, respectively (Table 1). From a quantitative point of view, ANOVA results (Table S1) showed a higher PUAs production in DP than in CC and variability among different sampling times (F=17.73, P<0.001).

From a qualitative point of view, the main compounds detected were the short-chain PUA hexadienal 273 (C6:2), which was present in both algae and resulted as the main compound in CC, and some long-274 chain PUAs, namely hexadecatrienal (C16:3), hexadecatetraenal (C16:4), and tetradecapentaenal 275 (C14:5), which were detected exclusively in DP (Fig. S1). Specifically, C14:5 was the most abundant 276 compound in DP, with average concentrations ranging from a maximum of $115.3 \pm 20.8 \ \mu g \ g^{-1}$ fw at 277 time T1 to a minimum of $13.2 \pm 2.4 \ \mu g \ g^{-1}$ fw at time T3, corresponding to the 50% and 32% of the 278 total aldehydes (Table 1; Fig. S1). C16:4 had the maximum average concentration at time T1 with 279 $36.4 \pm 8.8 \ \mu g \ g^{-1}$ fw, decressing gradually to a minimum of $5.9 \pm 1.5 \ \mu g \ g^{-1}$ (relative abundance of 280

8-29%) at time T6. C16:3 showed a similar pattern and an average maximum concentration of 30.9 281 \pm 9.3 µg g⁻¹ fw at T1 (relative abundance of 12%); then, its amount decreased to a minimum of 8.7 \pm 282 3.1 µg g⁻¹ fw at time T4, and it was not detected at all at T6. In DP the most variable compound was 283 C6:2, with a maximum average concentration of $16.1 \pm 2.2 \ \mu g \ g^{-1}$ fw at time T1; this aldehyde was 284 not detected at times T2 and T5. C6:2 was the main aldehyde in CC, accounting for 64-91% of the 285 total amount, and its average concentration had the maximum and minimum values of 15.4 ± 3.5 and 286 $1.0 \pm 0.1 \ \mu g \ g^{-1}$ fw at time T1 and T2, respectively. This short-chain aldehyde resulted not 287 significantly different between the two algae at the different sampling times. 288

289

290 *3.2 PUAs (dPUAs) and nutrients in seawater*

The total average concentrations of dPUAs detected in the seawater in the proximity of the two 291 macroalgae (Fig. S2) highlighted high values (ranging from 99.0 to 287.8 µg L⁻¹, corresponding to 292 293 0.9 and 2.7 µM) and temporal changes, with concentrations significantly lower at times T2 and T3, and higher (about 200 µg L⁻¹) in the subsequent times (T4-T6) (Table S1). A maximum value was 294 recorded at time T4 for DP and at time T6 for CC (287.8 ± 34.2 and $261.7 \pm 33.6 \mu g L^{-1}$, respectively). 295 ANOVA carried out on dPUAs concentrations in seawater showed highly significant differences for 296 the factor time (F=11.8, P<0.001), while no significant differences were found between the two algal 297 298 species (Table S1). Seawater temperatures during the sampling times ranged between 23 and 26°C, with salinity values of about 34-39 which were higher at T4-T5 (August-September). Concerning 299 nutrients, the concentration of nitrates was generally low, with a peak of 3.26 µM in T2 (June), while 300 phosphates were about 0.1-0.2 μ M, with a peak at T5 (0.24 μ M) (Table S2). 301

302

303 *3.3 Macroalgal complexity*

304 Surface area (S) and fractal dimension (D) were calculated to evaluate the different morphological 305 complexity of the apical parts of the two macroalgae that were analysed in the present work (Fig. S3).

306 Overall S and D were higher in DP than in CC. In CC the surface remained relatively constant along

the sampling period (20-25 cm² g⁻¹), while in DP S was higher at T5 and T6 than in the previous times (max value 127 cm² g⁻¹). As for D, more marked differences between the two algae were found at times T5 and T6, with higher values in DP than in CC (max values 1.84 and 1.73, respectively).

310

311 *3.4 Microphytobenthic community*

In total 25 genera were identified, belonging to 18 orders, and mostly to diatoms and dinoflagellates.
Sporadically, cyanobacteria (Cyanophyceae Schaffner, 1909) and juvenile stages of green algae
(Chlorophyta Pascher, 1914) were also observed.

Total density was significantly higher in DP (2608441 \pm 425527 cells g⁻¹ fw) than in CC (486309 \pm 77877 cells g⁻¹ fw) (Fig. 1; Table S3). Results of ANOVA showed a significant interaction between algae and times (F=5.75, P<0.001) (Table S3) due to a different temporal trend of density between the two macroalgae. Generally, diatoms represented the most abundant component of the microphytobenthic community, accounting for average values of 96% and 87% of the organisms in DP and CC, respectively.

Microalgal species belonging to the diatom group were classified into 22 genera and 15 orders. Some 321 species remained unidentified, namely undetermined pennate and centric diatoms (Fig. 2). On both 322 macroalgae the most represented order was Naviculales Bessey, 1907, that in DP showed an 323 increasing trend from T1, with an average density of 198287 ± 81945 cells g⁻¹ fw, to T6, with an 324 average density of 3489032 ± 1159137 cells g⁻¹ fw (Table S4), which represented 58% of the 325 epiphytic diatoms (Fig. 2A). For Bacillariales Hendey, 1937, a constant amount of about 60,000 cells 326 g^{-1} fw was observed in DP from T1 to T4, and a maximum at time T5 (average density of 621096 ± 327 213575 cells g⁻¹ fw) was recorded, which corresponded to 15% of the total diatoms (Fig. 2A). In 328 terms of average abundance, Lyrellales D.G. Mann and Cocconeidales E.J. Cox, 2015 represented 329 less than 10% of the diatoms (ranging from 25,8831 to 10,865 cells g⁻¹ fw), while species belonging 330 to other orders (i.e. Surirellales D.G. Mann, 1990, Thalassiophysales D.G. Mann, 1990, Mastogloiales 331

D.G. Mann, Licmophorales Round, 1990, Striatellales Round, 1990, Tabellariales Round, 1990)
 occurred with average densities below 100,000 cells g⁻¹ fw.

On CC almost all genera identified had a peak at T5, while a constant average density was observed 334 from T1 to T4, except for Licmophorales that reached a maximum at T2 (229178 \pm 91470 cells g⁻¹ 335 fw), accounting for 33% of the diatoms (Fig. 2B). Specifically, Naviculales, whose concentration 336 ranged from 25 to 56% of the entire diatom community, occurred with the lowest density of 70.457 337 \pm 31,328 cells g⁻¹ fw at T1 and a maximum average density at T5 (27,5598 \pm 89,219 cells g⁻¹ fw). 338 Bacillariales were about 10% of the total diatoms and increased consistently from T1 ($8,869 \pm 2.935$ 339 cells g^{-1} fw) to T5 (79,071 ± 18,396 cells g^{-1} fw), then they decreased at T6 (13,850 ± 4,154 cells g^{-1} 340 fw). Cocconeidales had a peak at T1 (33,615 \pm 16,335 cells g⁻¹ fw), accounting for the 16% of the 341 total diatoms. All other orders (i.e. Surirellales, Thalassiophysales, Mastogloiales, Licmophorales, 342 Striatellales, Tabellariales) represented less than 5% of the entire diatom assemblage. 343

The average density of dinoflagellates was significantly higher on DP than on CC. This group was recorded especially from T4 to T6 (Table S4) due to the presence of a bloom of the toxic dinoflagellate *Ostreopsis* cf. *ovata*, whose concentration reached a maximum average density at T6 (364,167 \pm 77,126 cells g⁻¹ fw).

The nMDS plot carried out on the microphytobenthic communities showed a clear separation of the 348 assemblages associated with the two macroalgae (Fig. 3A), but with a different temporal trend. 349 PERMANOVA supported this pattern resulting in a significant interaction between algae and times 350 (pseudo-F= 2,2779, P \leq 0.001; Table S5), and the post-hoc analysis carried out between the two algae 351 at the various sampling times confirmed the different structure of the two communities at each time. 352 SIMPER analysis revealed that average similarities for DP was 36% and for CC was 40%, whereas 353 dissimilarities between the two macroalgae at each time ranged from 66% and 82% (Table S6). 354 Results showed that the average dissimilarities between the two algae at each time were largely due 355 to the higher abundances of five orders, namely Naviculales, Lyrellales, Gonyaulacales Taylor, 1980 356

(i.e. *Ostreopsis* cf. *ovata*), Bacillariales, and Licmophorales, that were always more abundant on DP
than on CC.

359

360 *3.5 Meiobenthic community*

A total of 12 taxa belonging to the meiobenthos were identified, one represented by larval stage 361 Copepoda nauplii (referred from now on as copepod nauplii) defined as the larvae of the meiobenthic 362 copepod species (mainly harpacticoids). Copepods and their nauplii were counted separately in view 363 of their different ecology (Hicks and Coull 1983). Harpacticoida (33.1%) and Nematoda (31.2%) 364 were the dominant taxa, followed by copepod nauplii (25.2%), Gastropoda Cuvier, 1795 (5.4%), 365 366 Isopoda Latreille, 1817 (1.6%), Polychaeta Grube, 1850 (1.6%), Amphipoda Latreille, 1816 (0.7%) Halacaroidea Cunliffe, 1954 (0.5%), Ostracoda Latreille, 1802 (0.3%), Kinorhyncha Reinhard, 1885 367 (0.3%), Chironomidae Newman, 1834 (0.2%) and Bivalvia Linnaeus, 1758 (0.1%). Overall, total 368 density (N) and number of taxa (S) at high taxonomic resolution resulted higher on DP than on CC 369 (Figs 4 and S4). Total density was in mean 371 ± 46 and 90 ± 15 N g⁻¹ fw in DP and CC, respectively 370 (Fig. 4), while the number of taxa was on average 5.6 ± 0.2 in DP and 4.4 ± 0.2 in CC (Fig. S4). 371 ANOVA results showed a significant interaction between the factors alga and time (F=7.06, P<0.001; 372 F=5.9, P<0.001) for both total density and number of taxa (Table S7), suggesting that the differences 373 374 between macroalgae were not consistent over time. For total density (Fig. 5), significantly higher values occurred on DP than on CC at corresponding times, except for T1 (Table S7). Moreover, for 375 DP densities increased with a peak at T6 (828 \pm 126 N g⁻¹ fw), while for CC the highest density 376 occurred at T5 (217 ± 63 N g⁻¹ fw) and then decreased at T6. 377

The number of taxa showed the highest value on DP at T2 (7 ± 1) and the lowest on CC at T6 (3 ± 0) . It resulted significantly higher on DP than on CC at times T2, T4, and T6, while was lower at T5 (5 ± 1) (Fig. S4; Table S7).

The nMDS analysis of meiofauna communities showed a clear separation between samples belonging
to the two macroalgae but, for each alga, the composition of the meiofaunal assemblage changed

following a different temporal pattern (Fig. 3B). PERMANOVA supported this pattern with a significant interaction between algae and times (pseudo-F= 2,606, P \leq 0.001; Table S8). The assemblages on the two algae resulted significantly different at corresponding sampling times, with the exception of T5 and, within each alga community, the structure changed among times as shown by post hoc comparisons (Table S8).

SIMPER analysis revealed that average similarities for each macroalga were 42.8% for DP and 43.7%
for CC. Moreover, the dissimilarities between macroalgae at each time ranged from 52% to 82%.
Both similarities and dissimilarities were largely due to the variations in abundance of the three
dominant taxa: nematodes, copepods, and copepod nauplii (Table S9).

In DP average densities of nematodes (Fig. 5A; Table S10) increased from T1 ($57 \pm 22 \text{ N g}^{-1} \text{ fw}$) to a maximum in T6 ($287 \pm 46 \text{ N g}^{-1} \text{ fw}$), as well as the average values of the density of the copepods (Fig. 5B) that showed a similar trend to that of nematodes, with a pick at T6 ($285 \pm 69 \text{ N g}^{-1} \text{ fw}$). The densities of copepod nauplii (Fig. 5C), defined as the larvae of the meiobenthic Copepod species (harpacticoids), showed two peaks at times T2 ($215 \pm 52 \text{ N g}^{-1} \text{ fw}$) and T6 ($217 \pm 34 \text{ N g}^{-1} \text{ fw}$).

Conversely, in CC nematodes, copepods and copepod nauplii had lower mean densities at each time when compared with those found in DP (Fig. 5). Both nematodes and copepods densities were very low from time T1 (14 ± 6 and 8 ± 2 N g⁻¹ fw) to T4 (10 ± 2 and 14 ± 3 N g⁻¹ fw); only at T5 an increase occurred (105 ± 32 and 77 ± 27 N g⁻¹ fw) and then a decrease took place at T6 (19 ± 4 and $15 \pm 7 \text{ N g}^{-1}$ fw). Instead, the mean abundance of copepod nauplii slowly increased until T6 (43 ± 15 N g⁻¹ fw).

403

404 *3.6 Harpacticoid community*

Twelve harpacticoid species were identified from the two macroalgae; they belonged to 11 genera
and 11 families. The dominant species was *Heterolaophonte minuta* Boeck, 1873 (32.5%), followed
by *Amphiascus parvulus* Claus, 1866 (11.9%), *Harpacticus gracilis* Claus, 1863 (11.6%), *Paradactylopodia brevicornis* Claus, 1866 (10.8%), *Parastenhelia spinosa* Fischer, 1860 (10.2%),

and *Ectinosoma melaniceps* Boeck, 1865 (9.6%). The other species ranged from 6.8% (*Porcellidium viride* Philippi, 1840) to 0.2% (*Scutelledium longicaudum longicaudum* Philippi, 1840). Results of
ANOVA for total density (N), and number of species (S) are shown in table S11. Generally, total
densities (N) were significantly higher in DP than in CC at T2, T3, T4, and T6 (Fig. 5B), while the
number of harpacticoid species (S) was higher on DP compared to CC at T2 (7 vs 3) and T6 (7 vs 3)
(Table S12).

The nMDS plot of harpacticoid species showed a clear separation between algae and a different temporal pattern of assemblages on each macroalga (Fig. 3C). These results were supported by the significant interaction between the factors alga and time (PERMANOVA pseudo-F=19.084; P<0.01). In particular, harpacticoid assemblages resulted significantly different between the two algae at each time, with the exception of T1 and T5 (Table S13). A significant gradual change in the community structure took place in DP, whereas on CC significant community structure changes were evident at T5. The details of pairwise comparisons among times for each alga are shown in Table S13.

SIMPER results revealed that average dissimilarity between the two algae at each time was largely
due to the high abundances of five species, namely *Heterolaophonte minuta, Parastenhelia spinosa, Paradactylopodia brevicornis, Harpacticus gracilis,* and *Porcellidium viride,* that resulted more
abundant in DP than in CC (Table S14).

426 The mean abundance over time of these six species is shown in figure 6. In DP, the densities of *Heterolaophonte minuta* showed two peaks at times T4 (68 ± 17 N g⁻¹ fw) and T6 (82 ± 20 N g⁻¹ fw), 427 whereas on CC the densities were always low, with a peak at T3 (17 ± 7 N g⁻¹ fw). In DP the density 428 of Parasthenelia spinosa increased at T5 (17 ± 4 N g⁻¹ fw) and T6 (27 ± 9 N g⁻¹ fw), while in CC 429 there was an increase at T5 (12 ± 5 N g⁻¹ fw), then a decrease at T6 (4 ± 1 N g⁻¹ fw). *Paradactylopodia* 430 *brevicornis* showed higher densities at times T3 (11 ± 6 N g⁻¹ fw), T5 (16 ± 6 N g⁻¹ fw), and T6 (31431 \pm 11 N g⁻¹ fw) in DP, while on CC this species was almost always absent, except at T5 (16 \pm 6 N g⁻¹ 432 fw). The average densities of Harpacticus gracilis resulted always higher on DP. The densities of 433 *Porcellidium viride* showed the highest values on DP at T3 (31 ± 11 N g⁻¹ fw), then decreased and 434

finally disappeared at T6; conversely, the same pattern occurred for CC, but with lower densities. On DP the density of *Ectinosoma melaniceps* increased at T6 (19 ± 6 N g⁻¹ fw), while on CC this species was almost always absent, except at T5 (7 ± 3 N g⁻¹ fw).

438

3.7 Relationship between the meiofauna and harpacticoid assemblages with macroalgae complexity,
PUAs, and microphytobenthos

The results of DistLM (Table S15) carried out to analyze the relationships between all the three sets of variables and the meiofauna assemblage showed that for the marginal test the three macroalgae complexity descriptors alone accounted for 26% of the variation in the meiofauna abundance, PUAs for 17% and microphytobenthic taxa for 38%.

After fitting complexity descriptors, PUAs and microphytobenthos taxa explained an additional 7%
and 13% respectively, resulting in a total variation of 47%. These additional amounts were significant
according to the sequential test.

The same analysis carried out on the harpacticoid assemblages showed that the three complexity descriptors alone accounted for 19% of the variation in the harpacticoid abundance, PUAs for 13%, and microphytobenthic taxa for 23%.

451 After fitting complexity, PUAs and microphytobenthos taxa explained an additional 9% and 10% 452 respectively, resulting in a total variation of 38%. Only the additional amount of PUAs resulted 453 significant according to the sequential test.

454

455 **4. Discussion**

456 Qualitative and quantitative differences were observed in the PUAs produced by the two macroalgae 457 considered (*Dictyopteris polypodioides*, DP, and *Cystoseira compressa*, CC), in agreement with 458 previous results obtained for the same species (Pezzolesi et al., 2021). DP produced, in fact, higher 459 average concentrations than CC, but also a variety of long-chain compounds, such as hexadecatrienal 460 (C16:3), hexadecatetraenal (C16:4) and tetradecapentaenal (C14:5), while *C. compressa* produced

the short-chain hexadienal (C6:2) as main compound, demonstrating the ability of macroalgae to 461 462 produce short, middle and/or long-chain aldehydes (Kajiwara et al., 1996; Akakabe et al., 2003; Pezzolesi et al., 2021). These results highlighted that PUAs profile could be a fingerprint for each 463 algal species, since the same compounds were consistently detected regardless of the sampling period, 464 while their relative and total amount may vary depending on environmental conditions or 465 morphotypes (Alsufyani et al., 2014; Pezzolesi et al., 2021). In addition, since the apical parts of both 466 467 algae are those in which growth actively takes place, they are the most metabolically active parts, and therefore presumably also those in which the greatest production of PUAs takes place. Dissolved 468 PUAs (dPUAs) concentrations in proximity of the macroalgae was high (in the order of µM), 469 470 especially when compared with results of previous studies carried out in an Adriatic planktonic community (Ribalet et al., 2014), but are in accordance with values previously recorded in the same 471 site in other studies (Bartual et al., 2020; Pezzolesi et al., 2021) or hypothesized to occur in proximity 472 473 of the PUA producers (Ribalet et al., 2008; Bartual et al., 2018). These high concentrations are reasonable considering that the sampling area is colonized by a well structured phytobenthic 474 475 community (macro- and microalgal), thus dPUAs derive from the contribution of the different algal species. Additionally, the reduced hydrodynamism of this benthic site causes a low dispersion of any 476 477 secondary metabolites produced by the various organisms.

478 Seawater parameters (temperature, salinity and nutrients) were within the range previously reported for the Piscinetta site (e.g. Accoroni et al., 2012), and confirmed the seasonal variability associated 479 to this shallow inlet, that is subjected to a moderate anthropic impact (mainly in the form of summer 480 481 tourism, as it is a popular site for swimming in the summer months). Results showed a predominace of diatoms in the microphytobenthic community, as usually reported in the Adriatic Sea, even in 482 planktonic communities (e.g. Accoroni et al., 2016). Dinoflagellates were present at low densities, 483 except during the bloom of Ostreopsis cf. ovata, which was the main dinoflagellate and showed its 484 typical blooming trend, with maximum abundances recorded in late summer (September) and within 485 ranges previously observed (e.g. Gémin et al., 2020). 486

The main diatom genera found in the present study (e.g. Navicula, Cylindrotheca, Lyrella, Cocconeis, 487 488 Gyrosigma A.H. Hassall, 1845, Licmophora C. Agardh, 1827, Nitzschia, Mastogloia Thwaites ex W.Smith, 1856, Striatella C. Agardh, 1832, Coscinodiscus Ehrenberg, 1839) are among the most 489 common on Mediterranean macroalgae and in the microphytobenthos, particularly in the Adriatic Sea 490 (Carnicer et al., 2015; Accoroni et al., 2016; Rogelja et al., 2016; Pennesi and Danovaro, 2017; 491 Ternon et al., 2020). Species belonging to the orders Naviculates and Lyrellates on DP, to the 492 493 Licmophorales on CC and, generally, to centric diatoms on both algae, showed an inverse trend (low density at T1 and high at subsequent times) in relation to the production of PUAs by macroalgae. 494 These results could be ascribed to a negative effect of these compounds, as also demonstrated by 495 496 Ribalet et al. (2007) for some planktonic species. Long-chain aldehydes, such as C14-C16, can induce a stronger growth inhibition than short-chain PUAs, probably due to longer alkyl chains that increase 497 the reactivity of the molecules (Adolph et al., 2003). 498

499 Since many diatom species recorded in the present study, as those belonging to the genus Cylindrotheca spp. and Nitzschia spp., are themselves among the main PUAs producers (Wichard et 500 501 al., 2005a; Lavrentyev et al., 2015; Pezzolesi et al., 2017; Cózar et al., 2018), they may have developed different sensitivities and/or tolerances to these compounds in a species-specific way. 502 Studies have also shown a different susceptibility based on the life cycle, with more resistant juvenile 503 504 cells and more sensitive stationary phase cells (Ribalet et al., 2007; Leflaive and Ten-Hage, 2009) and based also on other factors, such as cell size, wall properties and lipid content. In particular, 505 species with a well-structured and mineralized cell wall, a low surface-volume ratio and a certain 506 507 lipid content can limit the ability of these compounds to penetrate the cell. Centric diatoms could potentially be more sensitive to PUAs, as those found in the present study (Chaetoceros spp. 508 Ehrenberg, 1844, Coscinodiscus spp., Guinardia spp. H. Peragallo, 1892 and Rhizosolenia spp. 509 Brightwell, 1858) are not listed among the species able to produce these compounds (Wichard et al., 510 2005). 511

Ostreopsis cf. ovata does not seem to be negatively influenced, in terms of abundance, by the 512 513 production of PUAs by macroalgae. Conversely, macroalgal compexity seems to explain the different abundances found on CC and DP. The presence of a rigid cell wall and the high biovolume could 514 partially explain the apparent lower sensitivity of Ostreopsis cells to PUAs compared to other species, 515 in addition to the protective role that can be offered by the mucilaginous layer produced by O. cf. 516 ovata, which provides an additional barrier against the substances dissolved in the water column 517 518 (Allen et al., 2016). Similarly, benthic dinoflagellates have shown a higher resistance than planktonic ones to potential allelochemicals (Ben Gharbia et al., 2017). 519

Regarding the meiofauna, 12 major taxa associated with the two macroalgae were found, with a total 520 521 density recorded in some samples of over 1300 individuals per gram of alga and within the range found in previous studies (Jarvis and Seed, 1996). Harpacticoid copepods, together with their nauplii, 522 were the most abundant taxon, comprising 58% of the total meiofauna, followed by Nematodes that 523 524 were 33%, as reported in previous studies (Carlo Heip, Magda Vincx, 1985; De Troch et al., 2005). By comparing the results obtained for the two macroalgae, the average abundances of total meiofauna 525 526 and the number of taxa resulted higher on DP than CC, as well as the abundances of copepods and nematodes. As a result, the community structure resulted different between the two macroalgae and 527 528 showed a different temporal pattern, although they were subjected to the same environmental 529 characteristics in terms, for instance, of hydrodynamism and tides. DP and CC are different for life

cycle, thallus structure and production of compounds. In particular, DP is formed by ribbon-like 530 fronds, with very irregular and proliferating edges, on which numerous meiobenthic organisms can 531 532 settle. In the area where sampling was conducted, the species persists in its fully developed habit for most of spring and summer. CC is a highly branched, leathery macrophyte; when fully grown, in the 533 study area it may reach 1-1.5 m in height. For a large-sized, habitat-forming species, its growth is 534 relatively fast, especially if compared with other Mediterranean fucalean brown algae. However, in 535 the area of the Passetto its full development is limited to a guite restricted period of the year (from 536 May to mid July); by mid summer this species loses most of its branches and persists in a more 537

reduced habit, with a few short branches. This means that most of its fronds occur in the field for a shorter time and are not available long enough to allow the settlement of a very diverse epiphytic community. The shorter temporal availability of this substrate also means that there will be a lower accumulation of sediment and detritus, a potential source of food for meiofaunal organisms (Hicks, 1980; Gibbons, 1988; Frame et al., 2007). Both species have apical growth, determined by divisions of a group of meristematic cells in DP and a single apical cell in CC. Therefore, the apical parts are the youngest parts of the thallus, on which sediment and epiphytes had less time to settle.

Taking into consideration all these aspects, our results suggest that the morphological complexity of 545 DP may affect total number of individuals, but also the associated species that have evolved 546 547 morphological adaptations necessary for the adhesion to a thallus of this type (Taylor and Cole, 1994; Chemello and Milazzo, 2002). Although macrophytes are not true fractal objects, estimates of 548 complexity using tools of fractal geometry have proved to be a useful approach for quantifying and 549 550 separating effects of habitat architecture from those of habitat quantity (Gee and Warwick, 1994; McAbendroth et al., 2005; Hooper and Davenport, 2006). However, fractal measures performed for 551 not truly fractal objects are just an estimate of complexity for a given scale. 552

The putative role of PUAs on structuring meiofauna may be reflected by the low density of nauplii, copepods, and nematodes at time T1 recorded on both macroalgae, which could be related to the high concentration of PUAs in this period, as observed also for microphytobenthos. Moreover, host specificity may be supposed, according to what determined by Bates and DeWreede (2007), that found specific chemical, structural and morphological characteristics of the algal species.

Harpacticoids have a number of features that make them an attractive group of benthic organisms in which PUA toxicity responses could be investigated, such as abundance, ecological importance, and short generation cycles (Raisuddin et al., 2007). In this study, harpacticoids associated with the two macroalgae consisted of 12 species. Even if the number of species was relatively low compared to results of previous studies (Hicks, 1977b; Arroyo et al., 2006), it is interesting to note that the 12 species belonged to 12 genera and 11 families, so showing a high taxonomic distinctness. Moreover,

no species was present exclusively on a single macroalga; only the relative abundance of individual 564 565 species was different between the two macroalgae, with Heterolaophonte minuta as the dominant species on both. Almost all species showed low abundance at T1, potentially suggesting a role of 566 PUAs, as previously postulated for meiofauna and microphytobethos. To our knowledge, only one 567 study was carried out in laboratory to analyze the effects of these compounds on the harpacticoid 568 *Tisbe holothuriae* (Taylor et al., 2007); while, other studies carried out on planktonic ecosystem have 569 570 shown deleterious effects of PUAs produced by diatoms on the reproduction of calanoid copepods (e.g. Temora stylifera, Calanus helgolandicus), which feed on them (Ianora et al., 2003, 2012; Miralto 571 et al., 1999), as well as apoptosis in maturing oocytes (Poulet et al., 2007a) during embryo 572 573 development (Romano et al., 2003) and in newly hatched nauplii (Ianora et al., 2004b).

574 It has to be considered that the present study was carried out in the field, so the link between PUAs effects on the community structure is more difficult to evaluate, due to the variability of PUAs 575 576 production both by diatoms (Wichard et al., 2005b) and macroalgae (Pezzolesi et al., 2021), of the copepod sensitivity (Ianora et al., 2003; Sommer, 2009) and to the detoxification ability developed 577 578 by certain species of copepods (e.g. Taylor et al., 2007; Wichard et al., 2008). As attested by Taylor et al. (2007), the benthos tends to be a more stressful environment compared to planktonic, where 579 rapid fluctuations in physical conditions occur and both natural and anthropogenic toxins can 580 581 accumulate at high levels within the sediments between the algal fronds. Since benthic organisms must be highly adapted to survive in such a harsh environment, it is not unreasonable to speculate 582 that harpacticoids may have a more developed detoxification system than planktonic calanoid 583 584 copepods, thus being better equipped to resist to the toxic effect of oxylipins. Indeed, a number of candidate detox genes were found in an analysis of 686 sequence tags expressed by Tigriopus 585 japonicus (Lee et al., 2005, 2008). Therefore, it is possible that the species found during the present 586 study have developed effective and efficient detoxification strategies to survive in an environment 587 such as the Passetto area. The different community structure between the two macroalgae and the 588 temporal changes could be also explained by the ecological and trophic role of the various species, 589

and by the different morphological evolution of the two macroalgae during the sampling time. This 590 591 is an important aspect to consider in the case of this study, as we sampled apical parts of thalli of CC and DP, that may be not representative of the entire thalli, as are the youngest parts on which epiphytes 592 had less time to settle and also those most exposed to light. In particular, four species (Porcellidium 593 594 viride, Parastenhelia spinosa, Heterolaophonte minuta, and Harpacticus gracilis) belonged to the phytal group sensu strictu, and two (*Paradactylopodia brevicornis*, and *Ectinosoma melaniceps*) 595 596 belonged to migrator and cosmopolitan group (Mascart et al., 2015). E. melaniceps is a tolerant eurytopic species, which presumably is not affected by the biochemical compounds produced by 597 macroalgae, as for example reported for the green alga Ulva lactuca (Hicks, 1980). P. brevicornis is 598 599 cosmopolitan species that have a wide distribution range and was found in different habitats, therefore able to adapt to a large number of different environmental conditions (Hicks, 1980). Conversely, P. 600 viride, H. minuta and H. gracilis are endemic species of the Mediterranean Sea with certain 601 602 morphological characteristics that allow them to live adhering to macroalgal surfaces; therefore, they could either be affected by the effect of the various compounds produced by both macroalgae and 603 604 microphytobenthos, or adapt to the various environmental conditions.

The DistLm procedure highlighted that the observed differences in the meiofauna and harpacticoid community structure between the two macroalgal species could be mainly explained by microphytobenthic main taxa, after fitting macroalgal complexity.

Microalgae, in fact, are at the base of the food web and provide energy for all the trophic levels above 608 them; thus, these results confirm their important role as primary food source of essentially all marine 609 610 food chains producers. In fact, most of the epifaunal taxa directly graze upon the macroalgae or the epiphytes for their food source, thus variations in the epiphytic algal density and composition between 611 612 two macroalgal species may influence the abundance of associated epifauna (Gestoso et al., 2010). Additionally, the macroalgal complexity, rather than PUAs production alone, resulted an important 613 component influencing the community structure. The role of the shape and structural complexity of 614 macroalgae in determining the abundance patterns and size structure of the epiphytic organisms is 615

supported in literature, either for epifaunal and epiphytic assemblages (e.g. Chemello and Milazzo, 616 617 2002; McAbendroth et al., 2005; Cacabelos et al., 2010). To better investigate the effect of macroalgal complexity and PUAs composition on variations of the microphytobenthos at some representative 618 genera levels, on main meiofauna taxa, and on harpacticoid species Pearson correlation analysis was 619 used (Fig. 7). Among PUAs, long-chain compounds (i.e. C14-C16), when compared with the short 620 one (i.e. C6:2), showed higher effects on the abundances of some representative microalgae (i.e. 621 622 Navicula spp., Lyrella spp., Cocconeis spp., Cylindrotheca spp./Nitzschia spp., and Ostreopsis cf. ovata), on the main meiobenthos taxa (i.e. Nematoda, Copepoda, copepod nauplii) and on 623 harpacticoid species (i.e. H. minuta, H. gracilis, P. viride). Results are in agreement with laboratory 624 625 studies performed to investigate the responses of allelochemicals, such as PUAs, on target organisms, including copepods and algae (Ianora et al., 2004a; Taylor et al., 2007; Adolph et al., 2004; Caldwell 626 et al., 2004; Ribalet et al., 2007; Pichierri et al., 2016). Macroalgal complexity, either in terms of 627 628 fractal dimension and area, is highly correlated and showed significant relationships with almost all the taxa considered, except with P. viride, protentially due to its ecological role, being a phytal species 629 630 able to colonize a variety of macroalgae thanks to morphological adaptations that has evolved to attach to morphologically diverse thalli (Hicks, 1980). 631

632 It is important to point out that several metabolites, as well as PUAs, are known to be produced by 633 macroalgae, including some (e.g. diterpenoids, polyphenols) that inhibit the growth of microalgae, bryozoan or other benthic species (Ternon et al., 2020), thus their contribution to the observed 634 dynamics could not be excluded. Additionally, the epiphytic community that colonize the macroalgal 635 636 surface may contribute either to the surface metabolome, adding inhibitory effects on the co-occurring species (Monti and Cecchin, 2012) or to the higher trophic levels, and directly interacting with the 637 epifaunal, thus adding complexity to the understanding of the relationships among these organisms 638 and to the role of PUAs. Finally, toxic microalgae such as O. cf. ovata, that was found to bloom 639 during the sampling period, have been reported to affect copepods, and particularly nauplii (Guidi-640 Guilvard et al., 2012), with different sensitivities among species (Pavaux et al., 2019). 641

Taking into consideration all these aspects, PUAs production by macroalgae, together with their
complexity, resulted one of the main factors involved in the benthic community structure dynamics,
but it is not enough to explain the differences in the microphytobenthos and meiofana assemblages.

646 **5.** Conclusion

Epiphytic communities on the two macroalgae highlighted a clear separation of the meiofauna and 647 648 microphytobenthos assemblages with different temporal trends. The average dissimilarities were due to several microalgal orders, namely Naviculales, Lyrellales, Gonyaulacales (i.e. Ostreopsis cf. 649 ovata), Bacillariales, and Licmophorales, and to the three meiofauna dominant taxa (nematodes, 650 651 copepods, and copepod nauplii) that were always more abundant on DP than on CC. Particularly, average dissimilarities of harpacticoid copepods were largely due to the abundances of five species, 652 namely Heterolaophonte minuta, Parastenhelia spinosa, Paradactylopodia brevicornis, Harpacticus 653 654 gracilis, Porcellidium viride, and Ectinosoma melaniceps. Generally, variations in the meiofauna and harpacticoid abundances were mainly due to macroalgal complexity variables 655 and 656 microphytobenthos, while a minor contribution was due to PUAs. Results documented that i) microphytobenthos resulted to affect the meiofauna population dynamics, in particular the 657 harpacticoid assemblages, attesting the role of these organisms as primary food source of essentially 658 659 all marine food chains producers, being at the base of the food web and providing energy for all the trophic levels above them, ii) the macroalgal complexity rather than PUAs production alone could be 660 a major trigger of the community structure. PUAs effects, in fact, resulted species-specific, thus 661 662 affecting some grazers instead of the entire community structure, as demonstrated also by Pearson's ρ correlations between taxa abundances and several macroalgal parameters. Among PUAs, long-663 chain compounds (i.e. C14-C16), with respect to the short one (i.e. C6:2), showed higher effects on 664 the abundances of some representative microalgal genera, harpacticoid species and on the main 665 666 meiobenthos taxa.

Since several of the epiphytic diatom species found, in addition to macroalgae, can produce PUAs, the understanding of the effects of these compounds on the community structure and on the relationships among taxa in field studies are complicated, thus opening to further in-depth investigations in simplified systems (i.e. microcosms).

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1018	Figure captions
1019	Figure 1 – Total density of microphytobenthos community (cells g-1 fw) in Dictyopteris
1020	polypodioides (DP) and Cystoseira compressa (CC) at the different sampling times (T1-T6).
1021	Figure 2 – Relative abundance (%) of species belonging to the Bacillariophyceae orders in A)
1022	Dictyopteris polypodioides (DP) and B) Cystoseira compressa (CC) at the different sampling times
1023	(T1-T6).
1024	Figure 3 - Two-dimensional nMDS of centroids for A) microphytobenthos orders, B) meiofauna,
1025	and C) harpacticoid community on the two algae Dictyopteris polypodioides (DP) and Cystoseira
1026	compressa (CC) at the different sampling times (T1-T6).
1027	Figure 4 – Total density (N g-1 fw) of meiobenthos in <i>Dictyopteris polypodioides</i> (DP) and
1028	Cystoseira compressa (CC) at the different sampling times (T1-T6).

- 1029 Figure 5 Total density (n° organisms g-1 fw) of A) nematodes, B) copepods and C) copepod
- nauplii in *Dictyopteris polypodioides* (DP) and *Cystoseira compressa* (CC) at the different sampling
 times (T1-T6).
- **Figure 6** Total density (N g-1 fw) of harpacticoids in A) *Dictyopteris polypodioides* (DP) and B)
- 1033 *Cystoseira compressa* (CC) at the different sampling times (T1-T6).
- 1034 Figure 7 Results of Pearson correlation analysis between substrate attribution (D and S) and PUA
- 1035 composition and densities of more representative genera of microphytobenthos, main meiofauna
- 1036 taxa and harpacticoid species, after square root transformation. Non-significant relationships are
- indicated by white squares. Signif. codes: *** p < 0.001, ** p < 0.01, * p < 0.05. D, fractal
- dimension; S, macroalgal surface; tot, total PUA; C14:5, tetradecapentaenal; C16:4,
- 1039 hexadecatetraenal; C16:3, hexadecatrienal; C6:2, hexadienal.
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1044 Table 1: Relative abundance (%) of identified and unknown polyunsaturated aldehydes (PUAs) and

total concentration of PUAs (µg g-1 fw) in *Dictyopteris polypodioides* (DP) and *Cystoseira*

1046 *compressa* (CC) at the different sampling times (T1-T6).

Time	Alga	C6:2	C16:3	C16:4	C14:5	Unknown	Tot (µg g⁻¹ fw)
T1	DP	10%	12%	15%	50%	13%	225.5
Т2	DP	0%	39%	20%	29%	14%	73.9
Т3	DP	7%	21%	29%	32%	11%	44.3
T4	DP	7%	8%	14%	60%	11%	100.3
T5	DP	0%	12%	19%	61%	7%	82.1
Т6	DP	6%	0%	8%	79%	7%	87.4
T1	CC	91%	0%	0%	0%	9%	17.1
T2	CC	64%	0%	0%	0%	36%	2.0
Т3	CC	81%	0%	0%	0%	19%	2.3
T4	CC	88%	0%	0%	0%	12%	3.4
T5	CC	83%	0%	0%	0%	17%	5.3
Т6	CC	68%	0%	0%	0%	32%	4.8

- 1068 Figure 1



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- 1090 Figure 2
- 1091 A)







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- 1125 Figure 5









1136 A)



	Pearson's p							
	0.75	0.50	0.25	-0.25	-0.50	-0.75		
	Comp	lexity	PUAs					
	D	S	tot	C14:5	C16:4	C16:3	C6:2	
Navicula spp.	***	***	*	**				
<i>Lyrella</i> spp.	* * *	***	*	**			*	
Cocconeis spp.		*				* * *		
Cylindrotheca spp. / Nitzschia spp.	* * *	* * *	*	*	*		*	
Other diatoms	***	* * *	*	*		*	**	
Ostreopsis cf. ovata	***					*		
Other dinoflagellates	***	**	**	**	*			
Nematoda	***	***	**	***				
Copepoda	***	***	*	**				
copepod nauplii	***	***					*	
Porcellidium viride					***	**		
Heterolaophonte minuta	***	***		*			*	
Parastenhelia spinosa	***	**						
Harpacticus gracilis	**	*	*	**				
Paradactylopodia brevicornis	***	**						