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MODIS time series contribution for the estimation of nutritional properties of alpine grassland

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Abstract

Despite the Normalised Difference Vegetation Index (NDVI) has been used to make predictions on forage quality, its relationship with bromatological field data has not been widely tested. This relationship was investigated in alpine grasslands of the Gran Paradiso National Park (Italian Alps). Predictive models were built using remotely sensed derived variables (NDVI and phenological information computed from MODIS) in combination with geo-morphometric data as predictors of measured biomass, crude protein, fibre and fibre digestibility, obtained from 142 grass samples collected within 19 experimental plots every two weeks during the whole 2012 growing season. The models were both cross-validated and validated on an independent dataset (112 samples collected during 2013). A good predictability ability was found for the estimation of most of the bromatological measures, with a considerable relative importance of remotely sensed derived predictors; instead, a direct use of NDVI values as a proxy of bromatological variables appeared not to be supported.

Keywords: NDVI, MODIS, time series, nutritional quality, grassland, Alps.

Introduction

The variability in nutrient content of pastures over time is of key importance in understanding the population dynamics of herbivores relying on grasslands as their only food resource. Normalised Difference Vegetation Index (NDVI) [Rouse et al., 1974] is widely used to monitor vegetation and plant responses to environmental change [Pettorelli et al., 2011].



NDVI is directly related to the fraction of absorbed photosynthetic active radiation intercepted (fAPAR) both theoretically [Sellers et al., 1992; Myneni and Williams, 1994] as well in the field [e.g. Di Bella et al., 2004; Viña and Gitelson, 2005; Richardson et al., 2007; Gitelson et al., 2014], in particular within herbaceous environments such as grasslands [e.g. Moreau et al., 2003; Fensholt et al., 2004; Boschetti et al., 2007; Li et al., 2010]. The correlation of NDVI with other characteristics of the canopy is only indirect, and needs to be empirically tested for different sensors and cover types [Roberts, 2001]. Biomass and Leaf Area Index (LAI) are probably the physical parameters which have been tested more in a large number of different environments: forests [e.g. González-Alonso et al., 2006; Tan et al., 2007; Madugundu et al., 2008], rice fields [e.g. Casanova et al., 1998; Wang et al., 2007; Gnyp et al., 2014], dry crops [e.g. Quarmby et al., 1993; Thenkabail et al., 2000; Marti et al., 2007; Boschetti et al., 2009] and extensive herbaceous lands [e.g. Moreau et al., 2003; Wang et al., 2005; Beeri et al., 2007; Darvishzadeh et al., 2008]. The relationship between NDVI and chemical parameters such as nitrogen content or digestibility has been tested only in few cases instead, and the lack of field validations of this relationship casts doubts on the usability of NDVI as a standalone estimator of forage quality [Pettorelli et al., 2011]. Islam et al. [2011] reported a high correlation between NDVI and nitrogen content of maize plants, but only at one specific phenological stage (8-leaf stage). Other studies linked NDVI with faecal crude protein (FCP) of large herbivores, which is considered a reliable indicator of the quality of vegetation [Cordova et al., 1978; Leslie and Starkey, 1985; Hodgman et al., 1996]. For example, Ryan et al. [2012] showed a positive relationship between log transformed NDVI and nitrogen faecal content of African Buffalo (Syncerus caffer) and Hamel et al. [2009] found a negative relationship between integrated NDVI in June and the date of the peak in FCP of mountain goats (Oreamnos americanus) and bighorn sheep (Ovis canadensis). Other studies used NDVI as a proxy of nutritional content of grasslands without validating with field measures [Griffith et al., 2002; Boone et al., 2006; Pettorelli et al., 2007; Mueller et al., 2008].

Another requirement in using remotely sensed data, besides field validation, is that spatial resolution should be adequate to recognise the surveyed surface, in order to reduce the presence of mixed pixels, which could lead to wrong estimations of parameters. For example, this may be an issue if estimations of grass biomass are done for pixels which contain also woodland [Elvidge and Lyon, 1985; Huete et al., 1985; Huete and Tucker, 1991].

Moderate Resolution Imaging Spectroradiometer (MODIS) data is probably the most widely used satellite data for phenological analyses, since its daily temporal resolution allows to reconstruct temporal profiles of vegetation indices, and its fifteen years long time series permits to explore changes of phenological parameters over time [Pettorelli et al., 2005; Boschetti et al., 2009]. The spatial resolution of MODIS (250 m for red and NIR bands; 500 m or 1 km for other bands) is sufficiently detailed for extensive high altitude grasslands such as those found on the Tibetan plateau and other mountain areas of China, characterised by large extensions and high homogeneity; indeed, most of the studies which use MODIS data to estimate forage quality in alpine grassland environments have been conducted in these sites [e.g. Zha et al., 2005; Gao et al., 2009; Zhang et al., 2011; Mao et al., 2012; Cong et al., 2013; Liang et al., 2013; Zhang et al., 2013; Xu et al., 2014]. Instead, in the European Alps, grasslands are typically located in the alpine zone, which is generally

fragmented between lower woody and higher rocky habitats, with a consequent higher spatial heterogeneity (at these elevations large uplands are uncommon, especially in the Western Alps). The high heterogeneity of grasslands in the European Alps entails that pure grassland 250 m pixels tend to be enclosed by mixed pixels, adding noise to the estimates when using NDVI as the only predictor of the nutritional content of pastures.

Being able to reliably estimate changes in the nutritional value of alpine grasslands over the last decades might be critical for better understanding the population dynamics of alpine herbivores. The case of the population of Alpine ibex (Capra ibex) in the Gran Paradiso National Park (GPNP, Northwester Italian Alps) is particularly emblematic: the dynamics of the population, estimated from total counts performed by park wardens since 1956 [von Hardenberg et al., 2000], was well explained up to 20 years ago by the interaction between winter snow depth and population density [Jacobson et al., 2004]. However, deterministic models fail to predict the decline in the population from nearly 5000 individuals counted in 1993 to little more than the 2300 counted in 2009 even when including in the models information on the age structure of the population [Mignatti et al., 2012]. Pettorelli et al. [2007] suggested that changes in the phenology and in forage quality due to climate change are negatively related to Alpine ibex kid survival and thus to the decline in the total population due to lack of recruitment. However, Pettorelli et al. [2007] based this hypothesis on the simple correlation between a crude NDVI-based indicator (NDVImax, estimated on 1×1 km squares and averaged over the whole territory of the GPNP) and yearly estimates of Alpine ibex kids survival. An estimation of the nutritional content of Alpine pastures in the GPNP made by calibrating a model with ground references would permit to more thoroughly investigate the interaction between climate change, grassland phenology, forage quality and the dynamics of the Alpine ibex population. The aim of this study is thus to assess the contribution of remotely sensed predictors from MODIS NDVI time series combined with geo-morphometric information to estimate the nutritional content of forage in Alpine grassland characterised by high environmental heterogeneity.

Methods

Study area

All field data have been collected within the territory of the GPNP (45.40°-45.65° N, 7.07°-7.58° E). The morphology of this area is typical of western Italian Alps, with high altitudinal ranges: valley bottoms are located between 1000 and 2000 m, while ridges rarely are at less than 3000 m (the highest altitude is reached at 4061 m, at the top of Gran Paradiso). Grasslands are located in the middle part, in the alpine zone (2200-2700 m), between woodlands and snow habitats (rocks, screes and glaciers): for this reason grassland surface appears very fragmented. Also the lithological composition of the area is quite heterogeneous, due to the presence of both calcareous and siliceous formations (calcschist ophiolite units - which are present mainly in the western part of the GPNP - define a mosaic of acid and basic soils, while granite gneisses - found mostly in the eastern part - are mainly acid). This condition, together with the heterogeneity in the exposure of mountainsides (mainly South in the southern part and East or West in the northern, but very variable due to the presence of a lot of deep small valleys), contributes to the presence of diversified plant communities.

Field data

Field data were collected during the 2012 growing season within 19 experimental plots, the maximum number of plots which one single operator can sample within the given time interval (see Figure 1 and Figure 2 for details). These plots have been chosen within GPNP alpine grasslands taking into account the following restrictions: (i) a minimum distance of 500 m from woodland, to avoid the possible interaction of this surface in the determination of the NDVI pixel value; (ii) a minimum distance of 1 km between plots, to minimise the autocorrelation between MODIS values; (iii) a balanced coverage of the altitudinal range of the grasslands and of the different exposures.

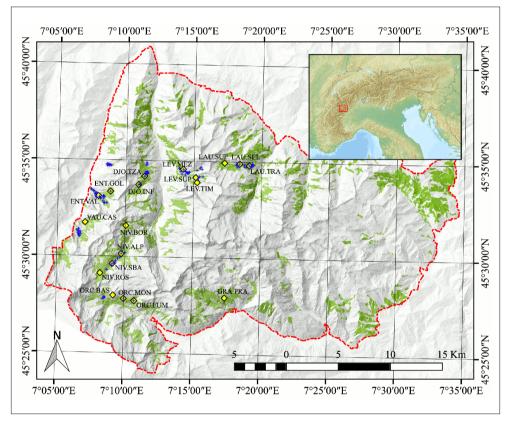


Figure 1 - Position of the 2012 (yellow squares) and 2013 (blue crosses) experimental plots within the Gran Paradiso National Park: red line represents park boundary; green area represents grassland surface; coordinates of the grid are in geographical units (WGS84). Little map shows the position of the Gran Paradiso National Park within the Alps (Background by Lencer via Wikimedia Commons, licence CC BY-SA 3.0).

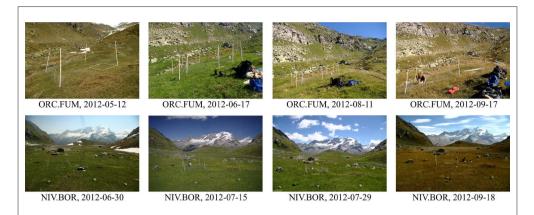


Figure 2 - Photos of two experimental plots, Alpe Fumetta (ORC.FUM, 2145 m) and Plan Borgnoz (NIV.BOR, 2688 m), in four different seasonal growing stages.

Each plot is composed of a squared surface of 3×3 m, enclosed to prevent grazing from domestic and wild herbivores. Enclosures were located on a homogeneous surface, enough representative of the surrounding buffer of 300 m (in terms of exposure, slope and microhabitat). Data collection was done from the beginning of the growing season (second half of May - end of June, depending on the plot) to the end of September (when all the vegetation communities in the plots reached senescence), with an interval between consecutive samples of two weeks. With this experimental design a total of 142 samples was obtained (from 4 to 10 samples for each plot). Each plot was divided into sectors of 50×50 cm; during each sampling session, all the grass present in one of these sectors was cut (with the exception of the eventual dry grass remaining from the year before). Also, a 1×1 m surface was reserved to take, each time, measures about grass height (taking 16 measures into each square with the help of a regular grid).

A supplementary set of samples used for validation purposes (see paragraph *Model validation*) was collected in 2013 in the early part of the season (June and July) with a different data design: in this case, each sample was cut within a different plot, to improve spatial distribution of points. Points were randomly generated on a 250×250 m grid with stratification (5 points in each cell) within alpine grasslands. To grant that validation would not be performed in different environmental conditions from the ones considered during model calibration, point selection was done only within the sectors in which 2012 sampling plots are present, and considering only points included in the altitudinal range of the 2012 dataset (2026-2804 m).

Collected samples were weighed (wet weight), dried in a ventilated oven at 60°C for 48 hours and weighed again (dry weight). Samples were analysed as described by Palmonari et al. [2014] to obtain, for each sample, the following nutritional information: (i) the relative content of crude protein (CP), obtained using the official semiautomated method AOAC 976.06 [AOAC, 1990]; (ii) the fraction of neutral detergent fibre (NDF), obtained using the method exposed by Goering and Van Soest [1970] with the modifications of Mertens et al. [2002]; (iii) the fraction of acid detergent fibre (ADF) and lignin (ADL), obtained with the official method AOAC 973.18 [AOAC, 2000]; (iv) the NDF digestibility after 24

hours (dNDF24) and 240 hours (dNDF240), measured by incubating forage in buffer fluids, obtained from rumen fluid of live cows, at body temperature under anaerobic conditions. The dNDF24 is estimated as the fraction of digested NDF after 24 hours, while dNDF240 is estimated from the fraction of indigestible fibre remaining after 240 hours incubation [Goering and Van Soest, 1970].

All the data used in this paper are publicly available on "figshare" repository [Ranghetti and Palmonari, 2015; Ranghetti et al., 2015].

Satellite data

Data taken from the MODIS TERRA MOD09Q1 250 m dataset, validated version V005 [NASA LP DAAC, 2014] were used, selecting all the images relatives to years 2012-2013 for the tile h18v04. This product provides composite images with a period of 8 days (in which the value of each pixel is the best value considering all 8 days) and two spectral bands, RED (red band, 620-670 nm) and NIR (near infrared band, 841-846 nm). Information about the day of acquisition of each pixel (DOY: Day of the Year, the progressive integer starting from the beginning of the solar year) was taken from the MODIS MOD09A1 product, since the MOD09Q1 dataset does not provide them.

Preprocessing of MODIS images has been done using MODIStsp package [Busetto and Ranghetti, 2016] and GDAL 1.11.0 [Warmerdam, 2008]. The script used is publicly available as part of the "LR EstGrass" project on GitHub [Ranghetti, 2015].

Images were maintained in the original sinusoidal projection to avoid reshaping noise. NDVI maps were obtained by applying, for each pixel, the formula (NIR-RED)/(NIR+RED).

Nineteen seasonal series from 2012 NDVI maps were considered, extracting values at the plot positions. Since spatial resolution (250 m) is quite coarse in a heterogeneous environment, each pixel value can vary consistently from its neighbours; for this reason, associating to each plot the exact value of the pixel in which the plot lies appeared rough estimated. Therefore, each value was computed from its image with a bidimensional 3×3 kernel gaussian lowpass filter (σ =1), taking into account the distance from each plot to the centres of the nearest pixels. To filter unrealistic extreme variability from our data, values identifiable as isolated peaks or pits (NDVI values whose difference from both the preceding and the succeeding value was more than 0.1) were first erased; then the series were smoothed by fitting a local polynomial regression [Cleveland et al., 1992] between NDVI and time. With these methods seasonal series resulted more homogeneous, but they are not continuous, since only one value for each period of eight days is available. To obtain daily values for each time series, the predictions were extended using a spline interpolation [Press et al., 1992].

The NDVI values were extracted from this preprocessed daily dataset in order to relate to each field record (using values which refer to the exact days of acquisition of each field sample); time series of experimental plots are shown in Figure 3. This work-frame has been used also for the field data collected in 2013.

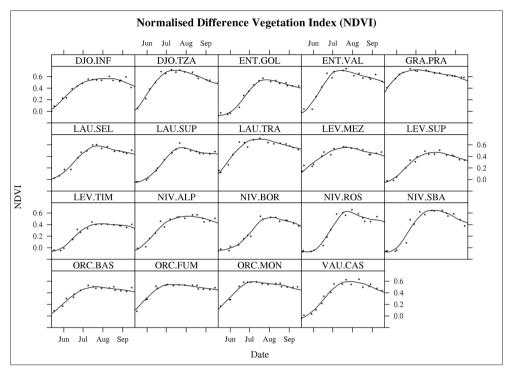


Figure 3 - Seasonal trend (2012) of NDVI values. Points are NDVI values computed from MOD09Q1 before smoothing (see paragraph Satellite data), lines represent daily smoothed values.

Data analysis

Dependent variables and predictors

Data analysis was performed with the open source statistical environment R version 3.1.1 [R Core Team, 2016] using packages sp [Pebesma and Bivand, 2005], rgdal [Bivand et al., 2016], raster [Hijmans, 2014], relaimpo [Grömping, 2006] and car [Fox and Weisberg, 2011]. The script used is publicly available as part of the "LR EstGrass" project on GitHub [Ranghetti, 2015] (release 1.0).

The relationships between predictors and the following nutritional parameters as dependent variables (Y) were modelled: Aboveground biomass (AB), crude protein (CP), neutral detergent fibre (NDF), acid detergent fibre (ADF), lignin (ADL), NDF digestibility at 24 hours (dNDF24) and NDF digestibility at 240 hours (dNDF240). AB is an absolute measure (expressed in grams), while the others are relative measures on total content (range 0-1). Each nutritional parameter was modelled separately as a dependent variable (Y).

In order to estimate these nutritional parameters, both morphometric and phenological parameters were used. While morphometric predictors can be easily obtained from a Digital Elevation Model (DEM), phenological parameters depends on seasonality; thus remotely sensed information is necessary to estimate them. Details about the used predictors and the methods to quantify or estimate them are described below.

Elevation

Elevation was extracted from the 10 m TINITALY digital elevation model [Tarquini et al., 2007; Tarquini et al., 2012]. Elevation values have been computed with a 3×3 kernel gaussian lowpass filter, as done for the NDVI values.

North-South aspect

An aspect map was derived from the elevation map and values were computed using the nearest neighbour method. From this map the North-South component of the aspect (aspectNS) was derived as the additive inverse of cosine of the aspect, which can vary from -1 (North) to 1 (South).

Beginning of growing season

The date of beginning of the growing season (BGS) represents a temporal metric useful to synthesise the variability between phenology of different stations [Chen et al., 2000]. BGS was estimated as the day when each seasonal series reached a threshold relative value t (a threshold is a fraction of the maximal annual NDVI range). Figure 4 represents how this metric (along with Dmax and NDVImax, described in the following paragraphs) were computed. This approach was proposed by Fontana et al. [2008] who used a value of $t \approx 0.75$ to estimate the BGS. However, this value of t cannot be used with our data because of the different spatial resolution (250 m in our case and 500 m in Fontana et al. [2008]). Thus, t was recomputed using NDVI time series compared with our 2012 field data, obtaining t = 0.51 as best threshold value. The methodology of this estimation is described in appendix A1, along with the analysis of the efficiency. BGS dates were used also to recompute, for each sample, the date from solar progressive Day Of the Year (DOY) to phenological day (DOS, Day Of the Season): DOS = DOY - BGS.

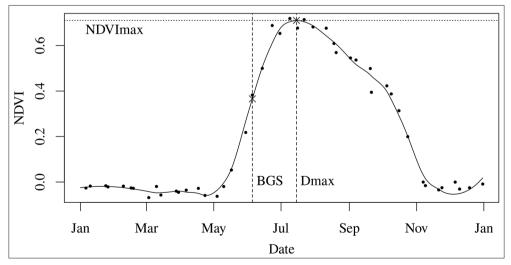


Figure 4 - Yearly 2012 NDVI profile of a sampling field plot (DJO.TZA): points are the values taken from MOD09Q1 products as described in section "Satellite data", line is the daily interpolation. Points marked with "×" are the NDVI values corresponding respectively with BGS (51% of the maximum NDVI range) and Dmax (maximum NDVI value: NDVImax).

NDVI maximum value.

NDVI is a measure of reflectivity, and its relation with biomass is known [Pettorelli et al., 2005; Boschetti et al., 2009]. So the maximum NDVI value (NDVImax) can provide information about the maximum seasonal productivity. This metric was simply computed as the maximum value among all the daily ones within each plot.

Day of NDVI maximum value.

The date in which this maximum NDVI value was reached (Dmax) was also considered as a proxy for the day of peak of the season: in fact, the NDVI decrease corresponds to a senescence of vegetation, so the starting day of this decrease may represent the moment in which grass stops to increment its biomass and begins to dry.

Daily NDVI.

Daily values of NDVI were used in combination with DOY (Day Of the Year) or DOS (Day Of the season) in order to allow the generation of daily measures of the Y dependent variables (see paragraph *Satellite data* for the preprocessing methods used to obtain daily NDVI values).

Model development

Here are described the models used to estimate Y dependent variables. First, the correlation matrix between predictors was computed in order to avoid collinearity (using Spearman's rank correlation ρ coefficients). Three couple of variables resulted autocorrelated: BGS-Dmax ($\rho = 0.75$, p-value = 0.0011), BGS-elevation ($\rho = 0.90$, p-value < 10⁻⁴) and Dmax-elevation ($\rho = 0.81$, p-value = 0.0002): for this reason, elevation and Dmax were excluded. Then, for each dependent variable, the following possible predictive linear models were built (using here and in the rest of the paper the modified Wilkinson-Rogers notation for linear models [Wilkinson and Rogers, 1973], widely used in statistical languages such as R):

$$Y \sim BGS + NDVImax + aspectNS + NDVI \times DOY$$
[1]

$$Y \sim BGS + NDVImax + aspectNS + NDVI \times (DOY + DOY^{2})$$
[2]

$$Y \sim BGS + NDVImax + aspectNS + NDVI \times (DOS + DOS^{2})$$
[3]

Model selection was performed based on the Akaike Information Criterion [Akaike, 1974; Burnham and Anderson, 2002]. A linear model with DOS (Day Of the Season) at first order

only was not built, since DOS is a linear combination of DOY and BGS, so this model would present the same AIC as the first model.

In order to evaluate the relative impact of each predictor, which is fundamental to understand the contribution of each predictor to the estimation of Y dependent variables, the R^2 contribution averaged over orderings among regressors [Lindeman et al., 1980; Chevan and Sutherland, 1991] were computed. The adjusted R^2 (adj- R^2) of the final models were also computed to measure the proportion of predicted variance. Moreover, the R^2 of the univariate models between each response variable and NDVI were computed to better check the proportion of variance which is explained by NDVI oneself.

Model validation

To validate these models two different methods were used: a leave-one-out cross-validation [Picard and Cook, 1984] and a validation with data collected during season 2013.

Since the season 2013 was quite different from the precedent, with a delayed snow melting (about 3 weeks) and summer temperatures generally lower, the use of the 2013 dataset allows to check the robustness and generalisation power of the defined models; also, it allows to improve the spatial interpolation, since the samples were collected in a greater number of plots (114 instead of 19; see paragraph *Field data*).

Validation was performed using RMSE (Root-Mean-Square Error) and MAE (Mean-Average Error).

Normalised values of RMSE and MAE (respectively NRMSE and NMAE) were also computed as the ratio between RMSE or MAE and the ranges of the Y in the 2012 datasets, in order to obtain values which are comparable being on the same scale. The residual distributions (mean and standard error) were also considered to detect the presence of bias in the predictions.

Results and discussion

Seasonal trend of variables

Figure 5 (solid lines) shows the trend of variables during the season. While protein content (CP) and NDF digestibility at 24 hours (dNDF24) present a clear decreasing trend, fibre variables (NDF, ADF and ADL) increase during the season. In particular, crude protein shows a very regular trend into each plot seasonal series: once the day of beginning of growing season (DOS) has been reached, the protein content shows a regular decrease, in agreement with results of other studies [Pérez Corona et al., 1998; Mountousis et al., 2011]. Aboveground biomass (AB) and NDF digestibility at 240 hours (dNDF240) do not show a clear trend (but it is possible to notice an increase of biomass in the first part of the season).

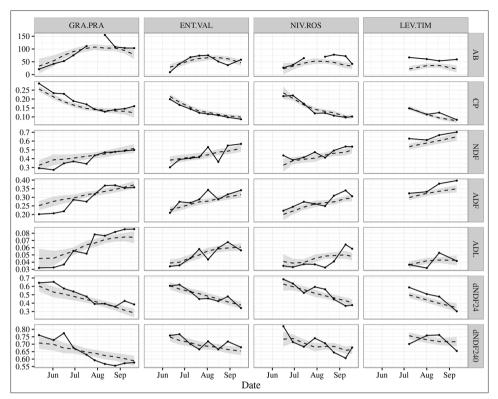


Figure 5 - Seasonal trends (2012) of field variables (each row represents one variable) for 4 sample field sites (one per column; see Appendix A2 for the graphics of all the 19 field plots). Solid lines represent the field measures, dashed lines represent predicted values, dotted lines border 95% confidence intervals.

Model characteristics and role of variables

The relative importance of each predictor in the models was summarised in Table 2. It is possible to notice that the predictor with the greater R^2 contribution is time (Day of the Year - DOY - or Day of the Season - DOS - at first and second order); only in the case of aboveground biomass (AB) DOY is not the predictor with the greater R^2 contribution. This evidences are in agreement with the seasonal trend of variables discussed in the paragraph above (the presence of a regular seasonal trend is reflected by a high relative importance of DOY).

The remaining portion of variance is mainly explained by remote sensing derived predictors: the date of beginning of growing season (BGS), the maximum seasonal NDVI value (NDVImax) and the daily NDVI value (NDVI). It is interesting to notice that daily NDVI solely explains a limited portion of variance (generally between 0.01 and 0.06, and greater than 0.10 only in the case of AB), while the contribution of seasonal parameters derived from NDVI (BGS + NDVImax) is greater (between 0.07 and 0.24). The univariate relation between these two quantities (see Tab. 1): indeed, with the exception of AB, for which a correlation with NDVI is shown, although low ($R^2 = 0.25$), in all the other cases R^2 are lower than 0.10. These results show that, in accordance with the literature [Mueller

et al., 2008; Pettorelli et al., 2011], NDVI used as a single proxy is a very poor predictor of the nutritional properties of Alpine grasslands, explaining not more than 10% of the total variance in most of the examined nutritional properties of Alpine grasslands.

Telef to ND v1 predictor; K- values are also given.									
Y	t value	Pr (<i>t</i>)	R ²						
AB	6.7	< 0.001	0.25						
СР	-2.6	0.010	0.04						
NDF	-1.2	0.225	0.00						
ADF	1.5	0.124	0.01						
ADL	4.3	< 0.001	0.11						
dNDF24	-1.6	0.106	0.01						
dNDF240	-4.0	< 0.001	0.09						

Table 1 - Univariate analysis of variance between each experimental variable and pure NDVI values. *t* test coefficients and their *p* values refer to NDVI predictor; R² values are also given.

This fact is evident also from Figure 5: the more variance is explained from the temporal predictor, the more the temporal trend of variable is regular. So in the case of biomass, in which the temporal predictor has a lower effect, the predicted values do not present a regular trend, since they are affected by each single NDVI value. At the opposite, when the day of the year has a higher importance there is a clear regular trend. For example, the predictions of CP are driven by a clear second order decreasing trend, that fits well the measured values. Also the predictions of NDF digestibility at 24 hours are linearly decreasing during the season. Predictions of neutral detergent and acid detergent fibre are instead linearly increasing. Globally, adjusted R² values ≈ 0.5 were found for most of the models, with the exception of crude protein (adj-R² = 0.8) and of NDF digestibility at 240 hours (adj-R² = 0.33): these values demonstrate that models fit quite well field data, in consideration of the high environmental heterogeneity of the study area.

Model validation

The use of both cross-validation and validation on an independent dataset allowed to analyse the robustness of each model to temporal extrapolation.

Table 3 shows validation metrics computed on both 2012 and 2013 validation sets, while Figure 6 compared measured and predicted values for 2012 and 2013. Using the 2012 dataset allows to obtain results that confirmed the calibration metrics (RMSE are generally acceptable, with NRMSE values never greater than 0.20, and mean Y differences are close to zero).

The model validation using the 2013 dataset led to different results. With the exception of biomass (AB), which showed results comparable with 2012 cross-validation ones, RMSE and MAE values result greater than with the 2012 data, providing thus less support to the models. Validation with the 2013 dataset provided worse results, which in some cases allowed to recognise the predictive power of models (CP, NDF and dNDF24), while in some others forced to reject them (ADF, ADL and dNDF240). Results about each single variable are discussed below.

Table 2 – Relative importance of models' predictors and adjusted R² (adj-R²) of models. Column "model" indicates which of the calibration models has been used (with the numeration used in paragraph "Model development"). Next columns indicate the R² contribution among each predictor [Lindeman et al., 1980; Chevan and Sutherland, 1991] (signs between brackets refer to the sign of each coefficient in the regression). In columns "DOY/DOS" values refer to the predictor used in each model ("DOY" if model [1] or [2], "DOS" if model [3]).

Y	Model	BGS	NDVImax	aspectNS	IAQN	DOY/DOS	DOY/DOS ²	SOQ/YOQ:IVQN	SOQ/AOQ:IAQN	adj-R ²
AB	2	(–) 0.172	(+) 0.068	(+) 0.023	(–) 0.164	(+) 0.048	(-) 0.043	(+) 0.009	(+) 0.009	0.507
СР	3	(–) 0.060	(+) 0.070	(+) 0.017	(–) 0.046	(–) 0.378	(+) 0.233	(-) 0.005	(+) 0.007	0.804
NDF	1	(-) 0.021	(–) 0.171	(-) 0.012	(+) 0.015	(+) 0.249		(-) 0.011		0.457
ADF	1	(–) 0.143	(–) 0.075	(-) 0.003	(-) 0.017	(+) 0.314		(+) 0.003		0.535
ADL	2	(–) 0.154	(+) 0.052	(+) 0.002	(-) 0.061	(+) 0.121	(–) 0.117	(-) 0.019	(+) 0.019	0.519
dNDF24	1	(+) 0.061	(+) 0.016	(+) 0.001	(-) 0.013	(-) 0.442		(+) 0.001		0.513
dNDF240	1	(+) 0.041	(–) 0.047	(-) 0.022	(+) 0.057	(–) 0.178		(-) 0.012		0.329

Table 3 - Validation metrics (RMSE, MAE, NRMSE and NMAE, see paragraph Model validation) computed for each model and for each validation dataset (2012 with cross-validation and 2013). ΔY values summarise the distributions of the differences between measured and predicted values (as mean \pm standard deviation), in order to estimate the presence of biases.

X	RMSE 2012	RMSE 2013	MAE 2012	MAE 2013	NRMSE 2012	NRMSE 2013	NMAE 2012	NMAE 2013	ΔY 2012	ΔY 2013
AB	19	16	15	13	0.13	0.11	0.10	0.09	0±19	-4±16
СР	0.02	0.03	0.02	0.03	0.09	0.14	0.07	0.11	$0.00{\pm}0.02$	0.00±0.03
NDF	0.08	0.07	0.06	0.06	0.18	0.16	0.14	0.13	$0.00{\pm}0.08$	0.03± 0.06
ADF	0.03	0.09	0.03	0.08	0.16	0.4	0.12	0.4	0.00±0.03	0.08± 0.03
ADL	0.009	0.03	0.007	0.03	0.15	0.5	0.12	0.4	0.000±0.009	0.026±0.015
dNDF24	0.08	0.09	0.06	0.07	0.14	0.16	0.11	0.13	$0.00{\pm}0.08$	-0.07 ± 0.05
dNDF240	0.05	0.08	0.04	0.06	0.16	0.2	0.13	0.19	0.00 ± 0.05	-0.04 ± 0.06

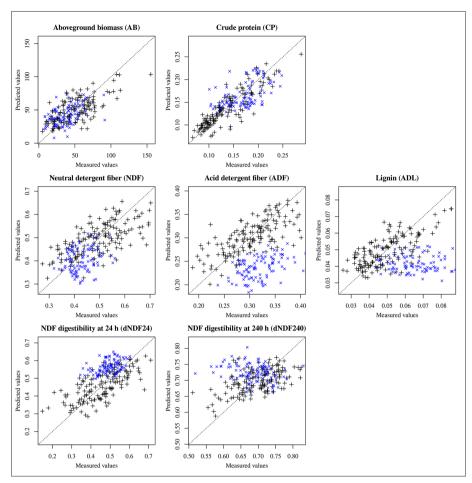


Figure 6 - Predicted values of each variable compared to the measured ones. Values of 2012 dataset are marked with black "+", 2013 ones with blue "×".

Aboveground biomass

The only variable for which 2013 validation metrics are better than 2012 ones is biomass (AB). This model is well supported by the results of the validation model (NRMSE = 0.11 and NMAE = 0.09, values which are acceptable and lower than in 2012) and its predictions are not biased ($\Delta Y = -4\pm 16$ g). The reason of that could be attributed to the low portion of variance explained by temporal predictor (DOY and DOY²): thanks to this, it was possible to better make temporal extrapolations.

Crude protein

The predictive power of crude protein (CP) decreased using 2013 dataset (NRMSE = 0.14 and NMAE = 0.11, while with cross-validation these values are respectively 0.09 and 0.07), but these values are comparable with the values obtained for other well-predicted variables, and predictions are not biased ($\Delta Y = 0.00\pm0.03$ g): for this reason, CP can still be considered predictable.

Fibre digestibility

Regarding the digestibility, is it possible to notice a different behaviour between NDF digestibility at 24 (dNDF24) and 240 hours (dNDF240). In the first case models are able to explain about half of the total variance, and validation metrics for 2013 dataset (NRMSE = 0.16, MAE = 0.13) are not different from the ones in 2012 (NRMSE = 0.14, NMAE = 0.11); however a little bias in the 2013 predictions is present ($\Delta Y = -0.07 \pm 0.05$; see also a graphical overestimations in Figure 6). The reason of this bias could be similar to the one hypothesised for crude protein: since the temporal predictors are more important in the model than NDVI, predicting values in a different season could produce less accurate results. NDF digestibility at 240 hours cannot be predicted, since its model explains too little variance (adj-R² = 0.329).

Fibre

Also in the predictions for fibres a difference similar as in the case of digestibility was noticed. In the case of neutral detergent fibre (NDF) the model is quite good and supported by validation results (NRMSE = 0.16, NMAE = 0.14 for 2013); a bias is not evident from Δ Y distribution (Δ Y = 0.03±0.06), but a little underestimation is graphically suggested. Acid detergent fibre (ADF) and lignin (ADL) models have to be rejected because validation with 2013 data clearly fails (NRMSE = 0.4, NMAE = 0.4 for ADF; NRMSE = 0.5, NMAE = 0.4 for ADL). There is a possible reason for these difference within both fibre and digestibility (predictability of dNDF24 and NDF but not of dNDF240, ADF and ADL) considering the nature of these variables: cellulose and lignin are the vegetal component which less manifest themselves into visible effects (like greenness), and NDF digestibility at 240 hours is linked to their relative content more than in the case of NDF digestibility at 24 hours [Clipes et al., 2006; Casali et al., 2008].

Models' improvement

All these results evidence how remotely sensed information can contribute to predict the daily values of nutritional content of grassland, since the relative importance of remote sensing derived predictors (measured as R^2 relative contribution) is quite relevant. Nevertheless, the experimental design used for this work did not allow to take advantage of inter-annual variability for the calibration of models, since they are built basing on field data collected in a single growing season; consequently, temporal extrapolation made on 2013 dataset (which cannot be used also for calibration, since samples are not collected during the whole growing season) evidenced a loss in predictive performance. For these reasons, a recalibration of models performed on a field dataset collected on a longer time window appear to be convenient before an operational use of models.

Furthermore, the use of different vegetation indices could also be exploited [Barati et al., 2011]. However, in this paper only NDVI was considered because additional MODIS spectral bands, required for the computation of most of these indices (typically blue or green bands), are not available at 250 m resolution. Making a spatial disaggregation of these bands or using data at 500 m or 1 km spatial resolution was avoided, since it could be potentially misleading in the context of high environmental heterogeneity of our study area. Equally, the use of different remotely sensed data (multi-spectral or hyper-spectral) appeared inconsistent with the perspective of an analysis of the vegetation dynamics in the

past, since a temporal resolution of few days is required in the context of our study area (in which cloud cover is frequent), and in our knowledge no other data sets with a sufficiently good temporal resolution are freely available for the past fifteen years.

Conclusions

This study showed how satellite data can be useful to estimate some important bromatological properties of alpine grasslands.

First, the correlation between single NDVI values and nutritional variables resulted low (in the case of biomass, a variable for which a direct link with the vegetation index is known) or absent (in the other cases): this result evidenced how the use of this remotely sensed index as a proxy for them can lead to errors. Nevertheless, MODIS data resulted useful for the estimation of temporal dynamics of pastures, which concur to explain a substantial portion of variance in terms of R^2 contribution. In particular, dates of beginning of growing season (BGS) predicted from satellite data showed a high correlation with ground truth (adj- $R^2 = 0.84$), confirming the validity of our method.

The possibility to estimate phenological predictors from MODIS data allowed to make predictive models of nutritional variables; anyway, it is evident that not all these variables can be predicted in the same way and with the same accuracy. Our results showed that the defined models allow to make predictions on biomass, crude protein, NDF digestibility at 24 hours and neutral detergent fibre, but not on NDF digestibility at 240 hours, acid detergent fibre and lignin. In the case of biomass, protein and neutral detergent fibre (and partially with NDF digestibility at 24 hours) the goodness of these models for doing temporal extrapolations was confirmed by validating them with 2013 data.

Our findings pave the road for a multi-annual analysis of changes in nutritional content of Alpine grasslands in terms of crude protein, NDF digestibility at 24 hours, neutral detergent fibre and biomass, which could be potentially useful to better understand the complex effects of climate change on the population dynamics of Alpine herbivores. However, in order to predict variable depending on specific seasonal condition with empirical multivariate model it is convenient to have a calibration dataset that includes multi annual data; so, to improve temporal robustness the recalibration of the models with field data collected in other seasons appears convenient before operationally use them.

Acknowledgments

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Appendices Appendix 1: estimation of BGS from NDVI time series

The procedure used to estimate the date of beginning of growing season (BGS) is here described.

The aim is to identify a threshold relative value *t* of the maximal annual NDVI range which can be considered as the value corresponding to the start of grass growth. The method proposed by Fontana et al. [2008] was used; *t* was recomputed considering BGS as the day when the mean grass height (measured as described in paragraph) exceeds the value of 5 cm. To compute this value, our 2012 field time series were interpolated with a spline to obtain daily values, as done for NDVI. Only 11 of 19 plots were used (in the 8 rejected plots, mean grass height was already greater than 5 cm in the first sampling). For each of the possible *t* thresholds (the 99 percentiles of each NDVI range), the days *d_t* when seasonal series reached *t* were computed. To avoid the influence of false detection in not meaningful period (winter), the first 100 days of the solar year were not considered. Then the difference between BGS and each *d_t* was analysed in terms of days of offset (OD) and adjusted R², calculated as showed in equations [S1-1] and [S1-2], where BGS_j and *d_{ij}* are the values of BGS and *d_t* in each plot *j*, and RSS_t and TSS_t are respectively the residual sum of squares and the total sum of squares of the linear regression between *d_t* and BGS.

$$OD_t = \frac{1}{n_{plot}} \sum_{j=1}^{n_{plot}} \left(d_{ij} - BGS_j \right) \quad [A-1]$$

$$adj - R^{2} = 1 - \frac{n_{plot} - 1}{n_{plot} - 2} \frac{RSS_{t}}{TSS_{t}} \quad [A - 2]$$

The method resulted efficient, allowing both to select the best threshold value (t = 0.51) and to well estimate the field measured BGS date (adj-R² = 0.84).

Figure S1-1 shows, for each possible value of *t*, the offset days (OD_t) and the adjusted R² values $(adj-R_t^2)$. The lowest absolute OD has been obtained with t = 0.51 (with this value of $t OD_t = 0$), which was then selected as best threshold value.

To check that this threshold is not only unbiased but also a good estimator of BGS, the adjusted R² was calculated as described in paragraph. With t = 0.51 we obtained adj-R² = 0.84, which is one of the highest obtained with all the possible threshold values (Fig. S1-1), justifying the use of t = 0.51 in the subsequent analyses.

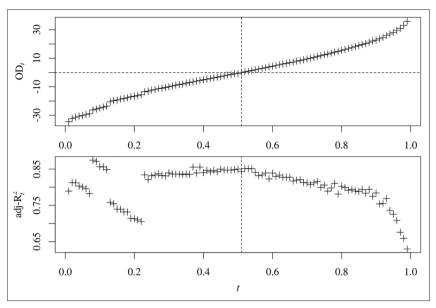


Figure A1-1 - Estimation of the NDVI threshold to use as BGS: OD (offset days) and adj-R² computed for each threshold value.

Appendix 2: seasonal trends (2012) of field variables

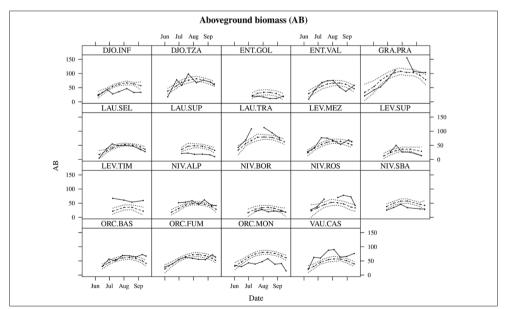


Figure A2 (Continued on the next page) - Seasonal trends (2012) of field variables. Solid lines represent the field measures, dashed lines represent predicted values, dotted lines border 95% confidence intervals.

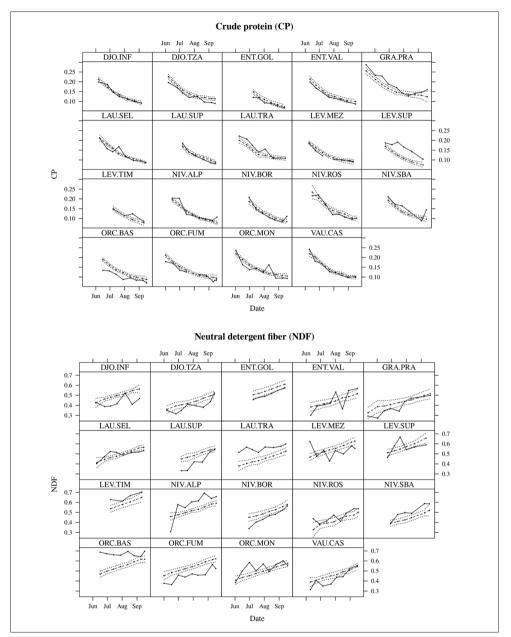


Figure A2 (Continued from preceding page and on the next page) - Seasonal trends (2012) of field variables. Solid lines represent the field measures, dashed lines represent predicted values, dotted lines border 95% confidence intervals.

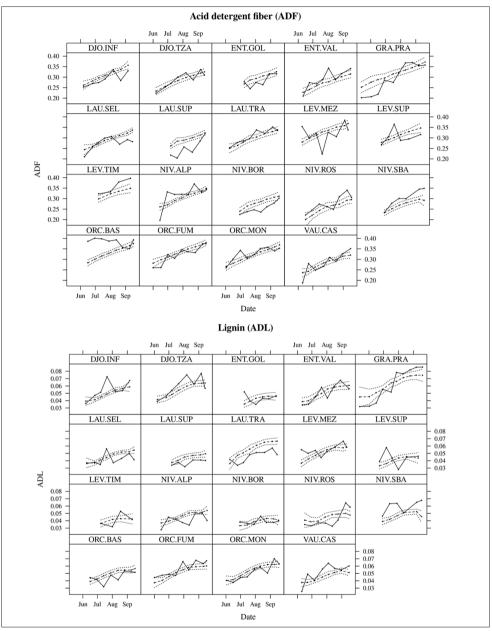


Figure A2 (Continued from preceding page and on the next page) - Seasonal trends (2012) of field variables. Solid lines represent the field measures, dashed lines represent predicted values, dotted lines border 95% confidence intervals.

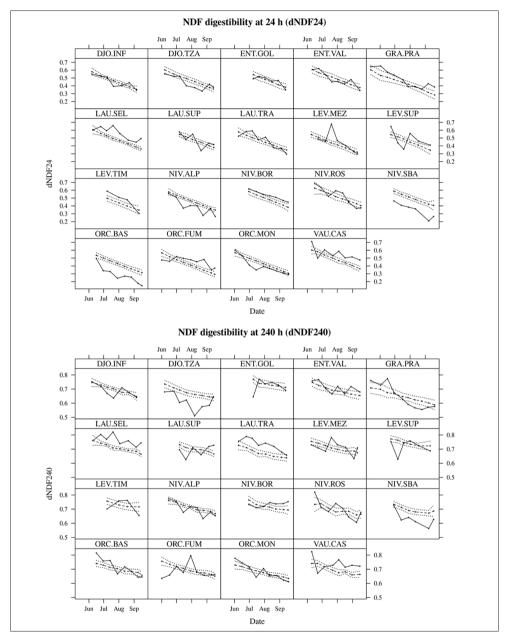


Figure A2 (Continued from preceding page) - Seasonal trends (2012) of field variables. Solid lines represent the field measures, dashed lines represent predicted values, dotted lines border 95% confidence intervals.

References

- Akaike H. (1974) A New Look at the Statistical Model Identification. IEEE Transactions on Automatic Control, AC-19, 6: 716-723. doi: http://dx.doi.org/10.1109/ TAC.1974.1100705.
- AOAC (1990) *Protein (crude) in animal feed (976.06)*. Official Methods of Analysis of the Association of Official Analytical Chemists. 15th Ed., p. 72.
- AOAC (2000) Fiber (Acid Detergent) and Lignin in Animal Feed (973.18). Official Methods of Analysis of the Association of Official Analytical Chemists. 17th Ed., p. 72.
- Barati S., Rayegani B., Saati M., Sharifi A., Nasri M. (2011) Comparison the accuracies of different spectral indices for estimation of vegetation cover fraction in sparse vegetated areas. Egyptian Journal of Remote Sensing and Space Science, 14 (1): 49-56. doi: http:// dx.doi.org/10.1016/j.ejrs.2011.06.001.
- Beeri O., Phillips R., Hendrickson J., Frank A.B., Kronberg S. (2007) Estimating forage quantity and quality using aerial hyperspectral imagery for northern mixedgrass prairie. Remote Sensing of Environment, 110 (2): 216-225. doi: http://dx.doi. org/10.1016/j.rse.2007.02.027.
- Bivand R., Keitt T., Rowlingson B. (2016) *rgdal: Bindings for the Geospatial Data Abstraction Library*. Available online at: http://CRAN.R-project.org/package=rgdal.
- Boone R.B., Thirgood S., Hopcraft J. (2006) Serengeti wildebeest migratory patterns modeled from rainfall and new vegetation growth. Ecology, 87 (8): 1987-1994. doi: http://dx.doi.org/10.1890/0012-9658(2006)87[1987:SWMPMF]2.0.CO;2.
- Boschetti M., Bocchi S., Brivio P.A. (2007) Assessment of pasture production in the Italian Alps using spectrometric and remote sensing information. Agriculture, Ecosystems & Environment, 118: 267-272. doi: http://dx.doi.org/10.1016/j.agee.2006.05.024.
- Boschetti M., Stroppiana D., Brivio P.A., Bocchi S. (2009) *Multi-year monitoring of rice crop phenology through time series analysis of MODIS images*. International Journal of Remote Sensing, 30: 4643-4662. doi: http://dx.doi.org/10.1080/01431160802632249.
- Burnham K.P., Anderson D.R. (2002) *Model selection and multimodel inference: a practical information-theoretic approach*. Springer Science & Business Media. doi: http://dx.doi.org/10.1007/b97636.
- Busetto L., Ranghetti L. (2016) *MODIStsp: an R package for preprocessing of MODIS Land Products time series*. Computers and Geosciences, 97: 40-48. doi: http://dx.doi. org/10.1016/j.cageo.2016.08.020.
- Casali A., Detmann E., Valadares Filho S., Pereira J., Henriques L., De Freitas S., Paulino M. (2008) Influência do tempo de incubação e do tamanho de partículas sobre os teores de compostos indigestíveis em alimentos e fezes bovinas obtidos por procedimentos in situ. Revista Brasileira de Zootecnia, 37 (2): 335-342. doi: http://dx.doi.org/10.1590/S1516-35982008000200021.
- Casanova D., Epema G., Goudriaan J. (1998) *Monitoring rice reflectance at field level for estimating biomass and LAI*. Field Crops Research, 55 (1-2): 83-92. doi: http://dx.doi. org/10.1016/S0378-4290(97)00064-6.
- Chen X., Tan Z., Schwartz M., Xu C. (2000) Determining the growing season of land vegetation on the basis of plant phenology and satellite data in Northern China. International Journal of Biometeorology, 44 (2): 97-101. doi: http://dx.doi.org/10.1007/ s004840000056.

- Chevan A., Sutherland M. (1991) *Hierarchical Partitioning*. The American Statistician 45: 90-96. doi: http://dx.doi.org/10.1080/00031305.1991.10475776.
- Cleveland W.S., Grosse E., Shyu W.M. (1992) *Local regression models*. In Statistical Models in S, Chambers J.M., Hastie T.J. (Eds.), Taylor Francis, chapter 8.
- Clipes R., Detmann E., Silva J., Vieira R., Nunes L., Lista F., Ponciano N. (2006) Evaluation of acid detergent insoluble protein as an estimator of rumen non-degradable protein in tropical grass forages. Arquivo Brasileiro de Medicina Veterinaria e Zootecnia, 58 (4): 694-697. doi: http://dx.doi.org/10.1590/S0102-09352006000400041.
- Cong N., Wang T., Nan H., Ma Y., Wang X., Myneni R., Piao S. (2013) Changes in satellite-derived spring vegetation green-up date and its linkage to climate in China from 1982 to 2010: A multimethod analysis. Global Change Biology, 19 (3): 881-891. doi: http://dx.doi.org/10.1111/gcb.12077.
- Cordova F., Wallace J.D., Pieper R.D. (1978) *Forage intake by grazing livestock: a review*. Journal of Range Management, 31 (6): 430-438. doi: http://dx.doi.org/10.2307/3897201.
- Darvishzadeh R., Skidmore A., Schlerf M., Atzberger C., Corsi F., Cho M. (2008) -LAI and chlorophyll estimation for a heterogeneous grassland using hyperspectral measurements. ISPRS Journal of Photogrammetry and Remote Sensing, 63 (4): 409-426. doi: http://dx.doi.org/10.1016/j.isprsjprs.2008.01.001.
- Di Bella C., Paruelo J., Becerra J., Bacour C., Baret F. (2004) Effect of senescent leaves on NDVI-based estimates of fAPAR: Experimental and modelling evidences. International Journal of Remote Sensing, 25 (23): 5415-5427. doi: http://dx.doi.org/10.1080/014311 60412331269724.
- Elvidge C., Lyon R. (1985) *Influence of rock-soil spectral variation on the assessment of green biomass*. Remote Sensing of Environment, 17 (3): 265-279. doi: http://dx.doi. org/10.1016/0034-4257(85)90099-9.
- Fensholt R., Sandholt I., Rasmussen M. (2004) Evaluation of MODIS LAI, fAPAR and the relation between fAPAR and NDVI in a semi-arid environment using in situ measurements. Remote Sensing of Environment, 91 (3-4): 490-507. doi: http://dx.doi. org/10.1016/j.rse.2004.04.009.
- Fontana F., Rixen C., Jonas T., Aberegg G., Wunderle S. (2008) Alpine grassland phenology as seen in AVHRR, VEGETATION, and MODIS NDVI time series - A comparison with in situ measurements. Sensors, 8 (4): 2833-2853. doi: http://dx.doi.org/10.3390/s8042833.
- Fox J., Weisberg S. (2011) An R Companion to Applied Regression. Second, Sage, Thousand Oaks CA.
- Gao Q., Li Y., Wan Y., Qin X., Jiangcun W., Liu Y. (2009) Dynamics of alpine grassland NPP and its response to climate change in Northern Tibet. Climatic Change, 97 (3): 515-528. doi: http://dx.doi.org/10.1007/s10584-009-9617-z.
- Gitelson A., Peng Y., Huemmrich K. (2014) *Relationship between fraction of radiation absorbed by photosynthesizing maize and soybean canopies and NDVI from remotely sensed data taken at close range and from MODIS 250m resolution data*. Remote Sensing of Environment, 147: 108-120. doi: http://dx.doi.org/10.1016/j.rse.2014.02.014.
- Gnyp M., Miao Y., Yuan F., Ustin S., Yu K., Yao Y., Huang S., Bareth G. (2014) *Hyperspectral* canopy sensing of paddy rice aboveground biomass at different growth stages. Field Crops Research, 155: 42-55. doi: http://dx.doi.org/10.1016/j.fcr.2013.09.023.
- Goering H., Van Soest P. (1970) Forage Fiber Analysis. USDA Agricultural Research

Service. Handbook number 379. US Department of Agriculture. Superintendent of Documents, US Government Printing Office, Washington, DC.

- González-Alonso F., Merino-De-Miguel S., Roldán-Zamarrón A., García-Gigorro S., Cuevas J. (2006) - Forest biomass estimation through NDVI composites. The role of remotely sensed data to assess Spanish forests as carbon sinks. International Journal of Remote Sensing, 27 (24): 5409-5415. doi: http://dx.doi.org/10.1080/01431160600830748.
- Griffith B., Douglas D., Walsh N., Young D., McCabe T., Russell D., White R., Cameron R., Whitten K. (2002) *The Porcupine caribou herd*. In: Arctic Refuge coastal plain terrestrial wildlife research summaries, Douglas D.C., Reynolds P.E., Rhode E.B. (Eds.), U. S. Geological Survey, Biological Resources Division, chapter 3, pp. 8-37.
- Grömping U (2006) *Relative Importance for Linear Regression in R: The Package relaimpo.* Journal of Statistical Software, 17 (1): 1-27. doi: http://dx.doi.org/10.18637/jss.v017.i01.
- Hamel S., Garel M., Festa-Bianchet M., Gaillard J.-M., Côté S.D. (2009) Spring Normalized Difference Vegetation Index (NDVI) predicts annual variation in timing of peak faecal crude protein in mountain ungulates. Journal of Applied Ecology, 46 (3): 582-589. doi: http://dx.doi.org/10.1111/j.1365-2664.2009.01643.x.
- Hijmans R.J. (2016) *raster: Geographic data analysis and modeling*. Available online at: http://R-Forge.R-project.org/projects/raster/.
- Hodgman T., Davitt B., Nelson J. (1996) *Monitoring mule deer diet quality and intake with fecal indices*. Journal of Range Management, 49 (3): 215-222. doi: http://dx.doi. org/10.2307/4002881.
- Huete A., Jackson R., Post D. (1985) Spectral response of a plant canopy with different soil backgrounds. Remote Sensing of Environment, 17 (1): 37-53. doi: http://dx.doi. org/10.1016/0034-4257(85)90111-7.
- Huete A., Tucker C. (1991) Investigation of soil influences in AVHRR red and nearinfrared vegetation index imagery. International Journal of Remote Sensing, 12 (6): 1223-1242. doi: http://dx.doi.org/10.1080/01431169108929723.
- Islam M., Garcia S., Henry D. (2011) Use of normalised difference vegetation index, nitrogen concentration, and total nitrogen content of whole maize plant and plant fractions to estimate yield and nutritive value of hybrid forage maize. Crop and Pasture Science, 62 (5): 374-382. doi: http://dx.doi.org/10.1071/CP10244.
- Jacobson A., Provenzale A., von Hardenberg A., Bassano B., Festa-Bianchet M. (2004) -*Climate forcing and density dependence in a mountain ungulate population*. Ecology, 85 (6): 1598-1610. doi: http://dx.doi.org/10.1890/02-0753.
- Leslie D., Starkey E. (1985) *Fecal indices to dietary quality of cervids in old-growth forests*. Journal of Wildlife Management, 49 (1): 142-146. doi: http://dx.doi. org/10.2307/3801860.
- Li G., Wang D., Liu S., Fan W., Zhang H., Xin X., Zhang H. (2010) Validation of MODIS FAPAR products in hulunber grassland of China. IEEE International Simposium on Geoscience and Remote Sensing Symposium (IGARSS), Honolulu, HI, pp. 1047-1050. doi: http://dx.doi.org/10.1109/IGARSS.2010.5650113.
- Liang S., Peng S., Lin X., Cong N. (2013) NDVI-based spatial-temporal change in grassland growth of China from 1982 to 2010. Beijing Daxue Xuebao (Ziran Kexue Ban)/Acta Scientiarum Naturalium Universitatis Pekinensis, 49 (2): 311-320.

- Lindeman R.H., Merenda P.F., Gold R.Z. (1980) *Introduction to Bivariate and Multivariate Analysis*. Glenview IL: Scott, Foresman.
- Madugundu R., Nizalapur V., Jha C. (2008) Estimation of LAI and above-ground biomass in deciduous forests: Western Ghats of Karnataka, India. International Journal of Applied Earth Observation and Geoinformation, 10 (2): 211-219. doi: http://dx.doi. org/10.1016/j.jag.2007.11.004.
- Mao D.-H., Wang Z.-M., Luo L., Han J.-X. (2012) *Dynamic changes of vegetation net primary productivity in permafrost zone of Northeast China in 1982-2009 in response to global change*. Chinese Journal of Applied Ecology, 23 (6): 1511-1519.
- Marti J., Bort J., Slafer G., Araus J. (2007) Can wheat yield be assessed by early measurements of Normalized Difference Vegetation Index? Annals of Applied Biology, 150 (2): 253-257. doi: http://dx.doi.org/10.1111/j.1744-7348.2007.00126.x.
- Mertens D., Allen M., Carmany J., Clegg J., Davidowicz A., Drouches M., Frank K., Gambin D., Garkie M., Gildemeister B., Jeffress D., Jeon C.-S., Jones D., Kaplan D., Kim G.-N., Kobata S., Main D., Moua X., Paul B., Robertson J., Taysom D., Thiex N., Williams J., Wolf M. (2002) - *Gravimetric determination of amylase-treated neutral detergent fiber in feeds with refluxing in beakers or crucibles: Collaborative study.* Journal of AOAC International, 85 (6): 1217-1240.
- Mignatti A., Casagrandi R., Provenzale A., von Hardenberg A., Gatto M. (2012) *Sex- and age-structured models for Alpine ibex Capra ibex ibex population dynamics*. Wildlife Biology, 18 (3): 318-332. doi: http://dx.doi.org/10.2981/11-084.
- Moreau S., Bosseno R., Gu X., Baret F. (2003) Assessing the biomass dynamics of Andean bofedal and totora high-protein wetland grasses from NOAA/AVHRR. Remote Sensing of Environment, 85 (4): 516-529. doi: http://dx.doi.org/10.1016/S0034-4257(03)00053-1.
- Mountousis I., Dotas V., Stanogias G., Papanikolaou K., Roukos C., Liamadis D. (2011) -Altitudinal and seasonal variation in herbage composition and energy and protein content of grasslands on Mt Varnoudas, NW Greece. Animal Feed Science and Technology, 164 (3-4): 174-183. doi: http://dx.doi.org/10.1016/j.anifeedsci.2011.01.007.
- Mueller T., Olson K., Fuller T., Schaller G., Murray M., Leimgruber P. (2008) *In search* of forage: Predicting dynamic habitats of Mongolian gazelles using satellite-based estimates of vegetation productivity. Journal of Applied Ecology, 45 (2): 649-658. doi: http://dx.doi.org/10.1111/j.1365-2664.2007.01371.x.
- Myneni R., Williams D. (1994) On the relationship between FAPAR and NDVI. Remote Sensing of Environment, 49 (3): 200-211. doi: http://dx.doi.org/10.1016/0034-4257(94)90016-7.
- NASA LP DAAC (2014) *MOD09Q1*. USGS/Earth Resources Observation and Science (EROS) Center, Sioux Falls, South Dakota, Available online at: https://lpdaac.usgs.gov/products/modis_products_table/mod09q1.
- Palmonari A., Fustini M., Canestrari G., Grilli E., Formigoni A. (2014). Influence of maturity on alfalfa hay nutritional fractions and indigestible fiber content. Journal of Dairy Science, 97 (12): 7729-7734. doi: http://dx.doi.org/10.3168/jds.2014-8123.
- Pebesma E.J., Bivand R. (2005) Classes and methods for spatial data in R. R News, 5 (2): 9-13.
- Pérez Corona M., Vâzquez De Aldana B., Criado B., Ciudad A. (1998) Variations in

nutritional quality and biomass production of semiarid grasslands. Journal of Range Management, 51 (5): 570-576. doi: http://dx.doi.org/10.2307/4003378.

- Pettorelli N., Pelletier F., von Hardenberg A., Festa-Bianchet M., Côté S.D. (2007) Early onset of vegetation growth vs. rapid green-up: Impacts on juvenile mountain ungulates. Ecology, 88 (2): 381-390. doi: http://dx.doi.org/10.1890/06-0875.
- Pettorelli N., Ryan S., Mueller T., Bunnefeld N., Jedrzejewska B., Lima M., Kausrud K. (2011) - The Normalized Difference Vegetation Index (NDVI): Unforeseen successes in animal ecology. Climate Research, 46 (1): 15-27. doi: http://dx.doi.org/10.3354/ cr00936.
- Pettorelli N., Vik J., Mysterud A., Gaillard J.-M., Tucker C., Stenseth N. (2005) Using the satellite-derived NDVI to assess ecological responses to environmental change. Trends in Ecology and Evolution, 20 (9): 503-510. doi: http://dx.doi.org/10.1016/j. tree.2005.05.011.
- Picard R.R., Cook R.D. (1984) Cross-Validation of Regression Models. Journal of the American Statistical Association, 79 (387): 575-583. doi: http://dx.doi.org/10.1080/01 621459.1984.10478083.
- Press W.H., Teukolsky S.A., Vetterling W.T., Flannery B.P. (1992) *Numerical Recipes in C: The Art of Scientific Computing*, 2nd Ed. Cambridge University Press, New York, NY, USA.
- Quarmby N., Milnes M., Hindle T., Silleos N. (1993) The use of multi-temporal NDVI measurements from AVHRR data for crop yield estimation and prediction. International Journal of Remote Sensing, 14 (2): 199-210. doi: http://dx.doi. org/10.1080/01431169308904332.
- R Core Team (2016) *R: A Language and Environment for Statistical Computing.* R Foundation for Statistical Computing, Vienna, Austria. Available online at: http://www.R-project.org.
- Ranghetti L. (2015) Estimation of nutritional properties of alpine grassland from MODIS NDVI data: R code. GitHub project on figshare repository. doi: http://dx.doi.org/10.6084/ m9.figshare.1381836.
- Ranghetti L., Palmonari A. (2015) Nutritional content of alpine grassland in Gran Paradiso National Park, 2013 dataset. figshare repository. doi: http://dx.doi.org/10.6084/ m9.figshare.1375306.
- Ranghetti L., Stendardi L., Palmonari A. (2015) Nutritional content of alpine grassland in Gran Paradiso National Park, 2012 dataset. figshare repository. doi: http://dx.doi. org/10.6084/m9.figshare.1375305.
- Richardson A., Jenkins J., Braswell B., Hollinger D., Ollinger S., Smith M.-L. (2007) Use of digital webcam images to track spring green-up in a deciduous broadleaf forest. Oecologia, 152 (2): 323-334. doi: http://dx.doi.org/10.1007/s00442-006-0657-z.
- Roberts G. (2001) A review of the application of BRDF models to infer land cover parameters at regional and global scales. Progress in Physical Geography, 25 (4): 483-511. doi: http://dx.doi.org/10.1191/030913301701543154.
- Rouse, J.W., R.H. Haas, J.A. Schell, D.W. Deering (1973) *Monitoring Vegetation Systems in the Great Plains with ERTS*, 309-317. Third ERTS Symposium, NASA SP-351 I.
- Ryan S., Cross P., Winnie J., Hay C., Bowers J., Getz W. (2012) The utility of normalized difference vegetation index for predicting African buffalo forage quality. Journal of

Wildlife Management, 76 (7): 1499-1508. doi: http://dx.doi.org/10.1002/jwmg.407.

- Sellers P., Berry J., Collatz G., Field C., Hall F. (1992) Canopy reflectance, photosynthesis, and transpiration. III. A reanalysis using improved leaf models and a new canopy integration scheme. Remote Sensing of Environment, 42 (3): 187-216. doi: http://10.1016/0034-4257(92)90102-P.
- Tan K., Piao S., Peng C., Fang J. (2007) Satellite-based estimation of biomass carbon stocks for northeast China's forests between 1982 and 1999. Forest Ecology and Management, 240 (1-3): 114-121. doi: http://dx.doi.org/10.1016/j.foreco.2006.12.018.
- Tarquini S., Isola I., Favalli M., Mazzarini F., Bisson M., Pareschi M., Boschi E. (2007) -*TINITALY/01: A new Triangular Irregular Network of Italy*. Annals of Geophysics, 50 (3): 407-425. doi: http://dx.doi.org/10.4401/ag-4424.
- Tarquini S., Vinci S., Favalli M., Doumaz F., Fornaciai A., Nannipieri L. (2012) Release of a 10-m-resolution DEM for the Italian territory: Comparison with global-coverage DEMs and anaglyph-mode exploration via the web. Computers and Geosciences, 38 (1): 168-170. doi: http://dx.doi.org/10.1016/j.cageo.2011.04.018.
- Thenkabail P., Smith R., De Pauw E. (2000) Hyperspectral vegetation indices and their relationships with agricultural crop characteristics. Remote Sensing of Environment, 71 (2): 158-182. doi: http://dx.doi.org/10.1016/S0034-4257(99)00067-X.
- Viña A., Gitelson A. (2005) New developments in the remote estimation of the fraction of absorbed photosynthetically active radiation in crops. Geophysical Research Letters, 32 (17): 1-4. doi: http://dx.doi.org/10.1029/2005GL023647.
- von Hardenberg A., Bassano B., Peracino V., von Hardenberg J., Provenzale A. (2000) -Preliminary results on the temporal variability of the Alpine ibex population in the Gran Paradiso National Park. Journal of Mountain Ecology, 5: 201-210.
- Wang F.-M., Huang J.-F., Tang Y.-L., Wang X.-Z. (2007) Estimation of rice LAI by using NDVI at different spectral bandwidths. Chinese Journal of Applied Ecology, 18 (11): 2444-2450.
- Wang J., Rich P., Price K., Dean Kettle W. (2005) Relations between NDVI, grassland production, and crop yield in the central great plains. Geocarto International, 20 (3): 5-11. doi: http://dx.doi.org/10.1080/10106040508542350.
- Warmerdam F. (2008) The Geospatial Data Abstraction Library. In: Open Source Approaches in Spatial Data Handling, Brent Hall G., Leahy M.G. (Eds.), Springer Berlin Heidelberg, pp. 87-104. doi: http://dx.doi.org/10.1007/978-3-540-74831-1_5.
- Wilkinson G., Rogers C. (1973) Symbolic description of factorial models for analysis of variance. Journal of Applied Statistics, 22 (3): 392-399. doi: http://dx.doi. org/10.2307/2346786.
- Willmott C., Matsuura K. (2005) Advantages of the mean absolute error (MAE) over the root mean square error (RMSE) in assessing average model performance. Climate Research, 30 (1): 79-82. doi: http://dx.doi.org/10.3354/cr030079.
- Xu W., Xin Y., Zhang J., Xiao R., Wang X. (2014) Phenological variation of alpine grasses (Gramineae) in the northeastern Qinghai-Tibetan plateau, China during the last 20 years. Shengtai Xuebao/ Acta Ecologica Sinica, 34 (7): 1781-1793. doi: http:// dx.doi.org/10.5846/stxb201303270531.
- Zha Y., Gao J., Zhang Y. (2005) *Grassland productivity in an alpine environment in response to climate change*. Area, 37 (3): 332-340. doi: http://dx.doi.org/10.1111/j.1475-

4762.2005.00637.x.

- Zhang G., Xu X., Zhou C., Zhang H., Ouyang H. (2011) Responses of grassland vegetation to climatic variations on different temporal scales in Hulun Buir Grassland in the past 30 years. Journal of Geographical Sciences, 21 (4): 634-650. doi: http://dx.doi. org/10.1007/s11442-011-0869-y.
- Zhang G., Zhang Y., Dong J., Xiao X. (2013) Green-up dates in the Tibetan Plateau have continuously advanced from 1982 to 2011. Proceedings of the National Academy of Sciences of the United States of America, 110 (11): 4309-4314. doi: http://dx.doi. org/10.1073/pnas.1210423110.

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