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**Brewer's spent grains as biofuels in combustion-based energy recovery processes:
Evaluation of thermo-oxidative decomposition**

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Brewer's spent grains as biofuels in combustion-based energy recovery processes: evaluation of thermo-oxidative decomposition

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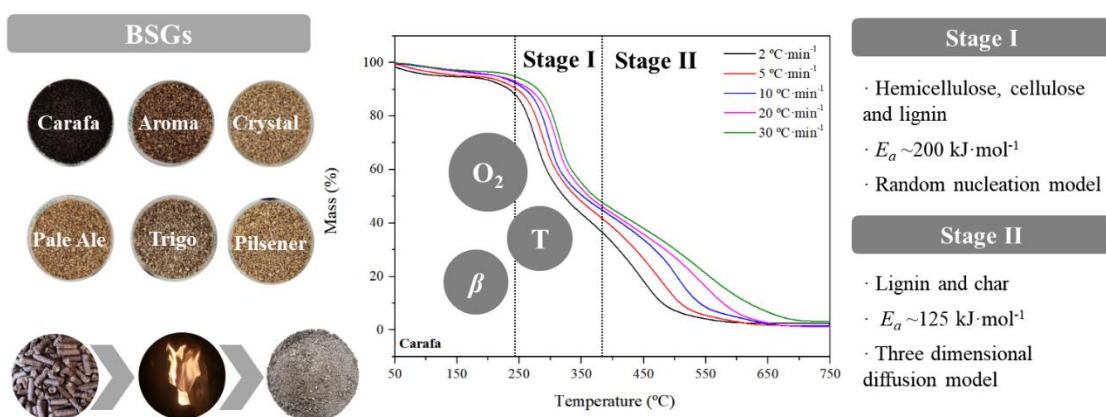
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Graphical abstract



Abstract

The high global generation of wastes and side streams from agri-food production has led to environmental impact and causes nature degradation due to their uncontrolled management. These wastes are profitable materials with significant economic value that could otherwise be exploited as new sources in the feed industry or the production of bioenergy. Among them, brewer's spent grain (BSG) is a solid by-product generated in the beer-brewing process that consists of the barley grain husks together with parts of the pericarp and seed coat layer. Although it is rich in fibres and proteins, its main use is currently limited to animal feed or simply deposition to landfills. This study pursues the evaluation of BSGs as biofuels in energy recovery processes. For this purpose, the elemental composition, the higher heating value, the content of volatiles, fixed carbon, moisture, and ash yield along with the kinetic analysis of its decomposition during thermo-oxidative combustion were assessed. The thermo-oxidative decomposition of hemicellulose, cellulose, lignin, and char occurred in two main stages. The average apparent activation energy for Stage I was 190-200 kJ·mol⁻¹, significantly higher than that of Stage II (100-150 kJ·mol⁻¹). Stage I revealed a random nucleation kinetic model (F_n) and involved the volatilisation of hemicellulose and cellulose and partially lignin. Stage II, described by a three-dimensional diffusion kinetic model (D_4), comprised the completion of lignin and the decomposition of char. The environmental advantage of using such residues was demonstrated with the low ash yield (in the range of 1.7-5.4%) and high calorific value between 17.8 and 19.1 MJ·kg⁻¹. The obtained results serve as the baseline for describing the energy recovery process in a reactor that uses BSG residues as feedstock under oxidative conditions. Altogether it is pointed out the feasibility of using pelletised BSGs regardless of the malt type or mixture as a renewable source of energy in incineration plants above 500 °C.

Keywords

Biomass, valorisation, renewable energy, kinetic analysis, activation energy

1. Introduction

Biomass can be defined as any hydrocarbon material source, including natural and derived structures, such as woody and herbaceous species, woody wastes, agricultural and industrial residues, waste paper, municipal solid waste, sawdust, grass, waste from food processing, animal wastes, aquatic plants and industrial and energy crops [1]. It has a complex chemical composition, of which organic and inorganic constituents are important for further conversion processes. The leading compounds of vegetal biomass are cellulose, hemicellulose, and lignin. Moreover, it also includes extractives, that are non-structural compounds, mostly soluble in water and/or numerous organic solvents such as fatty acids, lipids, terpenoids, phenolic compounds, glycosides, proteins, triglycerides, terpenes, waxes, cutin, suberin, flavonoids, betalains, or alkaloids [2–4].

Among the main uses of biomass, special interest is given to the production of energy [5,6]. Most of the energy produced from biomass comes from wood and wood wastes (64%), followed by solid waste (24%), agricultural waste (6%), and landfill gases (6%). There are several benefits of employing biomass for energy generation and biofuels production such as environmental, economic, social, or energetic. The use of biomass for the production of bioenergy can help to reduce greenhouse gases, contribute to the growth of healthier forests, enhance rural economies, and decrease the dependency on foreign oil [7].

In the last years, agri-food and agricultural residues have emerged as suitable feedstock for energy recovery processes due to their capacity in energy production [8]. Their high production rates and the necessity of suitable management have pointed them out as a feasible alternative to traditional fuels [9]. For being suitable for energetic valorisation,

calorific values above $15 \text{ MJ}\cdot\text{kg}^{-1}$ and a maximum degree of humidity from 10 to 15% is required [10].

The brewery industry gives a huge contribution to agri-food biomass production. Beer is one of the oldest and most widely consumed alcoholic drinks in the world, and the third most popular drink overall after water and tea. The beer production industry generates four main solid wastes: the brewer's spent grain (BSG) derived from the grain processing, hot trub (protein coagulation that occurs during the boiling of the wort), residual brewer's yeast (microbial fermentative activity), and the diatomaceous earth (the result of the beer clarification process), according to **Figure 1**.

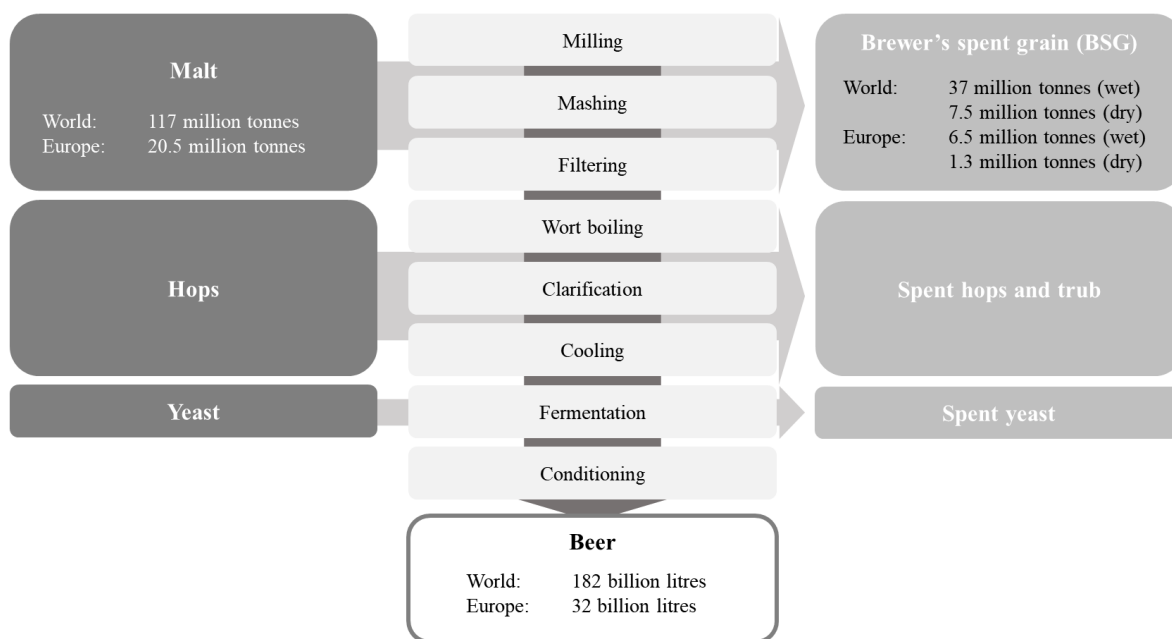


Figure 1. Schematic representation of the industrial brewery process and the by-products involved. Data for the use of feedstock, generation of wet BSG and beer production data is from 2020 [11–15].

BSG is a readily available by-product that consists of the barley grain husks together with parts of the pericarp and seed coat layer that is obtained as solid residue after the production of beer wort. Even if progress in technologies has permitted improvements in the process, in terms of quantity of by-products and efficiency, the quantity of residual BSG is difficult to

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reduce. In terms of BSGs generation, considering the global manufacture of beer in 2020 of 182 billion litres [11], of which around 32 billion litres were produced in Europe [12], a global amount of 37 million tonnes of wet BSG is estimated (7.5 million tonnes on dry basis) [13–15]. BSG production yield is around 20 kg of wet BSG per 100 L of brewed beer, and the wet BSG contains between 70 and 80% of water [16].

Although deposition of such residue may be a reasonable option for management, new uses of this waste may be explored from an economical perspective [17,18]. Among the different alternatives, involving the production of activated carbon, biomethane production, extraction of high-value compounds from BSG, animal or human nutrition, or fertilisers, energetic valorisation stands out as a promising environmental option [19–21]. Even if this biomass presents a suitable composition for energy valorisation processes, they are usually wet and involve a significant obstacle for being used as a solid biofuel and, therefore, a drying stage is required to moisture content lower than 15% [22]. Moreover, they have morphological heterogeneity, and thus, a study of their decomposition kinetics during thermochemical conversion becomes essential to understand their degradation mechanism before its use as energy vectors [23].

The evaluation of the decomposition kinetics has been widely applied for developing a model for the energy valorisation of biomass in full-scale systems [24,25]. Most studies in the literature are related to the kinetics of biomass decomposition during pyrolysis under inert atmospheres [26–28]. However, the evaluation of the oxidative decomposition for the simulation of combustion reactions has been scarcely explored, especially in BSGs. Reactions underwent in the oxidative atmosphere are usually heterogeneous and complex and commonly result in the generation of intermediate species due to both synergistic pyrolytic and oxidation phenomena [29,30]. According to the literature, gasification is recognized as the best technique for thermochemical conversions of biomass [31]. Different reactor

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configurations could be used for this purpose, such as fixed bed reactors, fluidized bed reactors, or spouted bed reactors among others [32]. For a suitable implementation of biomass in the industrial-scale application, a deep understanding of the oxidative kinetics when submitted to high temperature is necessary to find the best operational conditions in terms of design, scale-up, and feasibility [3].

A useful method for the evaluation of parameters for the design and optimization of the valorisation setup can be obtained using thermogravimetric analysis (TGA) when non-isothermal multiple heating rates are applied [33]. From these analyses, the main kinetic parameters can be calculated and serve as the baseline for the energy valorisation of BSGs in large-scale gasification processes that allow for product yield control [34,35]. For this purpose, a combination of iso-conversional methods, along with the use of Master-Plots and the use of the Perez-Maqueda *et al.* criterion permitted to assess the so-called kinetic triplet of BSGs: activation energy (E_a), the pre-exponential factor (A), and the kinetic function $f(\alpha)$. Finally, the feasibility of BSGs for being used as fuels in a pilot spouted bed reactor was verified.

2. Materials and methods

2.1. Materials and sample preparation

Brewer's spent grains (BSG) of pure Carafa, Aroma, Crystal, Pale Ale, Trigo, and Pilsener malt varieties were purchased from *Tu Cerveza Casera Homebrew* and processed in a 20 L pilot-scale brewing installation. Firstly, the malt was crushed in a mill. Then the boiler was filled with water and heated up to 65 °C with a heating rate of 1 °C·min⁻¹. Then, the mashed malt was added and thoroughly mixed, so that the malt mass was obtained, and allowed for fermentation for 1 h. During the fermentation process, the pH was kept at 5.50 with the aid of a few mL of lactic acid. Finally, the temperature was increased up to 75 °C for 5 min, before

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filtering and rinsing. For filtering, a Brewferm filter bowl was used, equipped with a perforated filter plate and a tap. At this point, the malt mass was rinsed with water at 78 °C. Afterwards, rinsed malt mass was pressed to remove water and subsequently dried at 50 °C on a large surface until a constant mass was reached, approximately 48 h. Lastly, BSGs were ground using a Moulinex Super Junior “S” grinder, and the obtained powders were conserved in a desiccator. The microscopic appearance of the powders was evaluated using a Leica MZ APO system.

2.2. Proximate analysis

Ash (*A*), volatile matter (*VM*), and moisture (*M*) contents of the BSGs were measured according to the standardized procedures described in ISO 18122:2016 (Solid Biofuels – Determination of ash content) [36], ISO 18123:2016 (Solid biofuels – Determination of the content of volatile matter) [37], and ISO 18134-1:2016 (Solid biofuels – Determination of moisture content – Oven dry method – Part 1: Total moisture – Reference method) [38], respectively. The fixed carbon (*FC*) was determined by the difference of these values, according to **Equation 1**.

$$FC(\%) = 100 - M(\%) - VM(\%) - A(\%) \quad (1)$$

Moisture content was determined in a Heraeus Vacutherm oven while volatile matter and ashes were determined in a Thermo Heraeus M-104 muffle furnace.

2.3. Ultimate analysis

The ultimate analysis of BSGs in terms of Carbon (*C*), Hydrogen (*H*), Nitrogen (*N*), Sulphur (*S*), and Oxygen (*O*) was determined using a CHNS-O Thermo Flash 2000 elementary analyser, operating at 900 °C in an atmosphere of pure oxygen. The oxygen content was calculated by the difference on a dry and ash-free basis according to **Equation 2**.

$$O(\%) = 100 - C(\%) - H(\%) - N(\%) - S(\%) \quad (2)$$

2.4. Higher heating value

The higher heating value was calculated using the equation proposed by Channiwala and Parikh, using carbon (C), hydrogen (H), nitrogen (N), sulphur (S), oxygen (O), and ash (A) weight percentages, according to **Equation 3**.

$$\begin{aligned} HHV (MJ/kg) = & 0.3491 \cdot C (\%) + 1.1783 \cdot H (\%) + 0.1005 \\ & \cdot S (\%) - 0.1034 \cdot O (\%) - 0.0151 \cdot N (\%) \\ & - 0.0211 \cdot A (\%) \end{aligned} \quad (3)$$

2.5. Thermogravimetric analysis (TGA)

The thermogravimetric analysis (TGA) was carried out with a Mettler Toledo TGA/STDA 851^e setup. The samples, with a mass between 2 and 5 mg, were placed in triplicates into 70 μ L alumina capsules. The analyses were performed with multi-rate non-isothermal segments ($\beta = 2, 5, 10, 20$ and $30 \text{ }^\circ\text{C}\cdot\text{min}^{-1}$) from 25 to $800 \text{ }^\circ\text{C}$ under an oxidative atmosphere with a flux of $50 \text{ mL}\cdot\text{min}^{-1}$ of pure oxygen. The obtained results were evaluated with the aid of the software STARe 9.10 from Mettler Toledo.

2.6. Kinetic analysis of the decomposition

2.6.1. Apparent activation energy

The most effective methods for the determination of the kinetic parameters are based on the analysis of dynamic thermogravimetric measurements obtained at different heating rates (β). Given that the kinetic parameters vary with the progress of the decomposition, iso-conversional methods, serve to calculate the evolution of activation energy during conversion. These methods use data from different multi-linear non-isothermal experiments

and avoid model assumptions for the analysis, which are the main source of error of model-fitting methods. In this study, multiple- α integral methods proposed by Flynn-Wall-Ozawa (FWO) and Kissinger-Akahira-Sunose (KAS) along with differential method given by Friedman were considered.

Flynn, Wall and Ozawa proposed the solution described in **Equation 4**, assuming the Doyle approximation, based on the Schlömilch expansion. The activation energy for different conversions can be calculated from the slope of the plots of $\log \beta$ versus the reciprocal of temperature.

$$\log \beta = \log \frac{Ea \cdot A}{R \cdot g(\alpha)} - 2.315 - \frac{0.457 \cdot Ea}{R \cdot T} \quad (4)$$

Kissinger-Akahira-Sunose (KAS) proposed a similar approach, by following **Equation 5**, where $g(\alpha)$ is the inverse integral kinetic model $g(\alpha) = \int_0^\alpha (f(\alpha))^{-1} d\alpha$. The activation energy may be taken from the slope of $\ln(\beta/T^2)$ versus the reciprocal of temperature.

$$\ln \left(\frac{\beta}{T^2} \right) = \ln \left(\frac{A \cdot R}{Ea \cdot g(\alpha)} \right) - \frac{Ea}{R} \cdot \frac{1}{T} \quad (5)$$

Friedman proposed **Equation 6**, in which the activation energy at a specific conversion degree can be obtained from the slope of $\ln(d\alpha/dt)$ versus the reciprocal of temperature.

$$\ln \left(\frac{d\alpha}{dt} \right) = \ln(A \cdot f(\alpha)) - \frac{Ea}{R} \cdot \frac{1}{T} \quad (6)$$

2.6.2. Analysis of the kinetic function

Thermal and thermo-oxidative decompositions in full-scale systems are described in terms of intrinsic kinetics, in which heat and mass transfer limitations are not included. These models circumvent the rigorous and exhaustive description of the chemistry of thermal

decomposition and describe the process using a simplified reaction pathway. It is widely known that each single reaction step is representative of a complex network of reactions. Among the different solid-state mechanisms, a sigmoidal function or a deceleration function are frequently selected to describe the kinetic mechanism.

Master Plots (MP) involve referencing theoretical curves depending on the kinetic model, which is generally independent of the kinetic parameters of the process. In many cases, the experimental kinetic data can easily be transformed into experimental curves and the comparison with the MPs allows for the selection of the appropriate kinetic model of the process under investigation. In this study, MPs were based on the integral form of the generalized kinetic equation (MP_g), reduced at $\alpha = 0.5$ for better visualization, as described by Criado [39]. This method also allows the determination of the kinetic model, when the activation energy is already known [39]. The kinetic analysis of the non-isothermal process is normally carried out employing a single-step kinetic equation (**Equation 7**), which integration leads to the inverse integral kinetic function $g(\alpha)$, as shown in **Equation 8**.

$$\frac{d\alpha}{dt} \equiv \beta \cdot \frac{d\alpha}{dT} = A \cdot f(\alpha) \cdot k(T) = A \cdot f(\alpha) \cdot e^{\frac{Ea}{R \cdot T}} \quad (7)$$

$$g(\alpha) = \int_0^\alpha \frac{d\alpha}{f(\alpha)} = \frac{A \cdot Ea}{\beta \cdot R} \cdot \int_0^\infty \frac{e^{-x}}{x^2} = \frac{A \cdot Ea}{R \cdot T} \cdot p(x) \quad (8)$$

where $x = \frac{E}{R \cdot T}$ and $p(x) = \frac{e^{-x}}{x^2} \cdot \sum_n \frac{n \cdot (1-n)}{x+2 \cdot (n+1)}$ is the result of the Senum and Yang approximation. The Senum and Yang's approximation was truncated at the fifth term since it permits obtaining a good approximation for a given temperature integral for $x > 10$ [40].

Once were evaluated the activation energies, the calculated $g(\alpha)$ were compared with the master plots and the mechanism was ascertained.

2.6.3. Pre-exponential factor and reaction order calculation

The kinetic triplet was completed with the pre-exponential factor (A) and reaction order (n). For this purpose, the Perez-Maqueda *et al.* criterion ($P-Mc$) was applied. This criterion relies on the independence of the activation energy (Ea) and the pre-exponential factor (A) on the heating rate (β), which must be assessed with the aid of the Coats-Redfern expression shown in **Equation 9**.

$$\ln \frac{\beta \cdot g(\alpha)}{T^2} = \ln \frac{A \cdot R}{Ea} + \frac{Ea}{R} \cdot \frac{1}{T} \quad (9)$$

If so, the minimization of ξ in **Equation 10** will give the proper reaction order (n) for the model and, subsequently, the most precise pre-exponential factor (A), by averaging the values from the intercept at $y = 0$ of the Coats-Redfern equation.

$$\xi(n, \alpha) = \sum_i^h \left| (-R) \cdot \frac{d}{dt} \left(\frac{\ln(\beta_i \cdot T^{-2} \cdot g(\alpha))}{T^{-1}} \right) - Ea_{iso} \right| \quad (10)$$

2.7. Combustion of BSGs in a pilot incineration plant

As a validation approach, the dry beer spent grains (BSGs) of the different malt varieties were pelletised together in a Smart Wood PLT-50 setup, with a production capacity of 20-30 kg·h⁻¹. Pellets with a diameter of 6 mm and a length of 30 mm were obtained, which morphology is shown in **Figure 2**. Afterwards, they were subjected to isothermal combustion for 1 h at temperatures in the range from 400 to 700 °C in steps of 50 °C in a Thermo Heraeus M-104 muffle furnace, to validate the ash yields and the appearance after combustion.

Finally, 1 kg of the pelletised BSGs was applied in an own-designed pilot incineration plant, such as that shown in **Figure 2**. The setup was equipped with a hopper with a base square 0.5 m length and inclined walls together with a cylindrical bar to promote vibrations to prevent

blocks. The feeding mass flow was controlled by a rotary valve with a diameter of 4.86 cm, previously calibrated. Then, the conveyor belt with dimensions of 4.55 m length, 0.15 m wide, and spaced partitions every 0.175 m, allows for the transport of the pellets to the feeding tube regulated by a rotary valve and subsequently to the incineration unit. The incineration unit consists of a combustion chamber with a moving grill in a circular motion to prevent potential slagging. Inside this chamber, two ignition glow plugs with lengths 140 mm and 250 W each are installed, along with four temperature sensors monitor the temperature of the top, the gas outlet, the heart of the combustion chamber, and the collector of ashes. Air is conveyed from the downstream of the combustion chamber to ensure combustion and extract fumes from the chamber. Ashes are recovered into a tray located at the bottom of the chamber, with dimensions of 19×30.4×9.4 cm³. The tightness of the chamber was ensured with an isolated door that included a viewing port. Overall, the operating conditions were feeding rate of ~1 kg·h⁻¹ and airflow of 50 L·min⁻¹ (21% O₂), to assure combustion temperatures above ~500 °C during operation. Once the mentioned amount of pellet was consumed, the remnant ash was collected and the yield and apparent density were calculated.



Figure 2. Pellets of dry brewer's spent grains (BSG), an overview of the pilot incineration plant and detail of the incineration unit.

3. Results and discussion

3.1. Initial characterization of the biomass

The brewer's spent grains (BSG) of pure Carafa, Aroma, Crystal, Pale Ale, Trigo, and Pilsener malt varieties were initially characterised. For this purpose, the macroscopic appearance, the microscopic morphology, and the composition –both by proximate and ultimate analyses– were evaluated and the obtained results are shown in **Table 1**.

In terms of macroscopic appearance, milled and dried BSGs revealed different colours, which are due to the particular malt variety and preparation conditions [41]. These colours are strictly correlated to the flavours that malt brings to the beer [42]. Indeed, malt flavours are used to bring balance to beer and are the source of limitless combinations for differentiation [43]. Dark roasted grains such as those of Carafa are used to impart a signature rustiness reminiscent of coffee or chocolate and contribute to significant levels of colour to beer. However, they are frequently used in lower percentages below 5% to avoid intense modifications. Typically, these roasted malts infer a beer colour of 800 to 1000 EBC (European Brewing Convention). Conversely, amber malts such as the Aroma variety, bring hints of toasted walnut, dark caramel, and dried fruit with a colour of 350 to 450 EBC and can be used in percentages up to 15%. Then, the Crystal, Pale Ale, Trigo or Pilsener malts can bring smother flavours and colours varying from 100 to 400 EBC, according to the stewing phase and the pre-roasting temperatures.













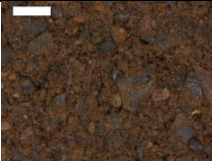
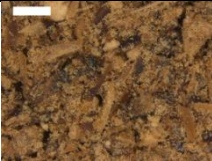

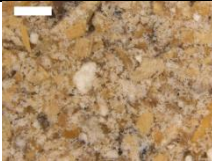
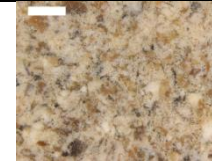

The microscopic morphology of the powder after grinding revealed a relatively homogeneous structure, with the size of grain lower than 1 mm. The preparation of uniform powder samples is crucial for a representative evaluation of the biomass, especially in terms of homogeneity during further thermo-oxidative stability and kinetic analysis of decomposition.

Subsequently, the moisture (*M*), volatile matter (*VM*), fixed carbon (*FC*), and ash (*A*) along with the weight proportion of the constituting elements (C, H, O, N and S) were evaluated to validate BSGs as solid biofuels [44]. Overall, the proximate analysis showed similar percentages of moisture, volatiles, ash yield, and fixed carbon for all the BSGs regardless of the malt variety. The applied drying procedure permitted to obtain a very low retained moisture percentage, below 1.5%. Volatile content was comparable among all the BSGs and ranged between 86 and 95%. The perceived differences may be ascribed to the remnant presence of low molar mass molecules such as extractives or polysaccharides from the maceration stage. Ash yield was around 2% in all the BSGs except in the Carafa malt variety, in which it was significantly higher, reaching 5.4%. The specific roasting treatment of this malt before maceration may be responsible for this high ash yield.

In terms of ultimate analysis, Carbon, Hydrogen, Oxygen, and Nitrogen percentages were virtually equal in the BSGs of all the malt varieties. A close view of the obtained values highlights the higher Carbon percentage for Carafa's BSGs. Again, the different roasting nature of this malt would be responsible for this behaviour. This higher percentage of C implies a higher calorific value and, from the point of view of energy production, it enhances its suitability in thermo-chemical conversion processes. A null content of Sulphur and low Nitrogen content was found for all analysed BSGs, which suggests the improbable production of unwanted SO_x and NO_x compounds during combustion.

Finally, the higher heating value was calculated applying the equation of Channiwala and Parikh [45], and values between 17.8 and 19.1 MJ·kg⁻¹ were achieved, higher than those described for agri-food and agricultural residues, with reported HHVs around 15 MJ·kg⁻¹ [8,24,25]. This observation together with the null Sulphur content and the low percentages of ash yield, approve these materials for the production of bioenergy.

Table 1. Macroscopic appearance of the milled and ground dry BSGs and microscopic images of the obtained powders. Scale bar in micrographs represents 1 mm.

| | | Carafa | Aroma | Crystal | Pale Ale | Trigo | Pilsener |
|--------------------------------------|--|---|---|---|---|---|---|
| | Macroscopic appearance of the milled BSGs |  |  |  |  |  |  |
| | Macroscopic appearance of the ground BSGs |  |  |  |  |  |  |
| | Microscopic morphology of the powders |  |  |  |  |  |  |
| Proximate analysis (% w/w) | Moisture (M) | 0.7 | 0.6 | 1.2 | 0.4 | 0.8 | 1.1 |
| | Volatile matter (VM)^c | 88.3 | 91.8 | 86.6 | 95.3 | 95.6 | 92.2 |
| | Fixed carbon (FC)^{a,c} | 5.5 | 5.9 | 10.4 | 2.0 | 1.2 | 4.6 |
| | Ash (A)^c | 5.4 | 1.7 | 1.7 | 2.3 | 2.5 | 2.1 |
| Ultimate analysis (% w/w) | C^b | 46.7 | 44.7 | 44.4 | 44.6 | 44.4 | 44.0 |
| | H^b | 6.4 | 6.4 | 6.4 | 6.6 | 6.6 | 6.4 |
| | N^b | 44.3 | 47.3 | 47.8 | 46.9 | 46.9 | 48.2 |
| | S^b | 2.6 | 1.6 | 1.4 | 1.9 | 2.1 | 1.4 |
| | O^a | - | - | - | - | - | - |
| HHV (MJ·kg⁻¹) | | 19.1 | 18.2 | 18.1 | 18.4 | 18.3 | 17.8 |

(^a)Obtained by difference; (^b)Weight percentage on dry ash-free basis; (^c)Weight percentage on dry basis.

3.2. Thermo-oxidative stability

The thermo-oxidative stability of the BSGs was assessed at different heating rates ($\beta = 2, 5, 10, 20$ and $30 \text{ }^\circ\text{C}\cdot\text{min}^{-1}$) and the thermogravimetric (TG) and first derivative curves (DTG) were obtained. For demonstration purposes, **Figure 3** shows the TG and DTG curves for the BSGs analysed at $\beta = 10 \text{ }^\circ\text{C}\cdot\text{min}^{-1}$. The TG and DTG curves were furtherly analysed, and the onset and endset decomposition temperatures (T_o, T_e), the peak temperature (T_p), and the mass loss percentages for the different stages (Δm) were calculated for this heating rate, which values are gathered in **Table 2**.

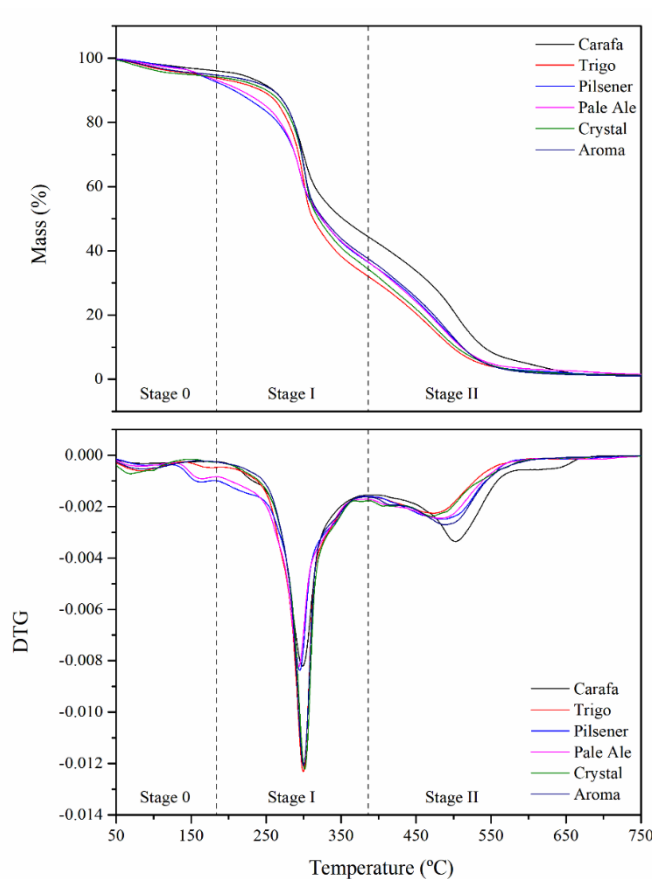


Figure 3. Thermogravimetric (TG) and first derivative (DTG) curves of the dry BSGs measured at $\beta = 10 \text{ }^\circ\text{C}\cdot\text{min}^{-1}$.

Table 2. Onset and endset decomposition temperatures (T_o , T_e), peak temperature (T_p) and mass loss percentages for the different thermo-oxidative degradation stages (Δm) obtained at $\beta = 10 \text{ }^\circ\text{C}\cdot\text{min}^{-1}$ for the dry BSGs. A standard deviation below 2% was omitted for the sake of clarity.

| | Stage 0 (Humidity + volatiles) | | | Stage I (Cellulose + hemicellulose + lignin) | | Stage II (Lignin + char) | | Endset T_e ($^\circ\text{C}$) | $\Delta m_1 /$ ($\Delta m_2 -$ 25%*) |
|----------|--------------------------------------|----------------------------------|---------------------|---|---------------------|----------------------------------|---------------------|---|---|
| | T_o ($^\circ\text{C}$) | T_{p0} ($^\circ\text{C}$) | Δm_0 (%) | T_{p1} ($^\circ\text{C}$) | Δm_1 (%) | T_{p2} ($^\circ\text{C}$) | Δm_2 (%) | | |
| Carafa | 265.1 | 91.1 | 3.2 | 298.7 | 56.1 | 505.6 | 42.5 | 547.1 | 3.2 |
| Aroma | 274.8 | 86.2 | 4.3 | 300.3 | 60.6 | 491.6 | 38.5 | 533.5 | 4.5 |
| Crystal | 275.4 | 77.4 | 5.1 | 300.0 | 61.8 | 476.9 | 37.0 | 529.6 | 5.2 |
| Pale Ale | 264.7 | 91.1 | 3.6 | 293.9 | 62.2 | 481.0 | 37.7 | 527.8 | 4.9 |
| Trigo | 273.0 | 82.2 | 4.5 | 298.6 | 66.5 | 472.7 | 32.5 | 523.3 | 9.0 |
| Pilsener | 258.6 | 96.0 | 2.8 | 294.1 | 61.3 | 483.9 | 37.5 | 537.2 | 4.9 |

*According to char content reported in [46].

The initial mass loss with a percentage below ~5% occurred from 25 to 170 $^\circ\text{C}$ with a peak (T_{p0}) located between 80 and 90 $^\circ\text{C}$. This stage is mainly due to water vaporization but also caused by proteins and extractives decomposition, especially above 150 $^\circ\text{C}$ perceivable for BSGs from Pilsener, Pale Ale and Trigo malts.

Afterwards, two main decomposition stages were found, according to expectations for lignocellulosic biomass [34]. Stage I involved the decomposition between 170 and 380 $^\circ\text{C}$ and Stage II from 380 $^\circ\text{C}$ onwards. Although this two-stage behaviour has been widely reported in the literature, different approaches have been proposed for the identification of the components decomposing in each stage. It has been proved that the synergistic and simultaneous purely pyrolytic thermal degradation and heterogeneous oxidation occur during decomposition in the oxidative atmosphere, giving, as a result, a complex thermogram. Therefore, the mass loss profile during thermogravimetric analysis may be the sum of the pyrolysis of biomass, the heterogeneous oxidation reactions, and the combustion of the generated char. If the biomass composition is considered as a combination of hemicellulose, cellulose, and lignin, the peaks in the DTG curve would be assigned according to the reported

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stability of these components. However, the reaction mechanism corresponding to the different peaks has been the focus of discussion and two hypotheses have been proposed. The first hypothesis describes that during Stage I, the complete volatilization of the three main constituents (hemicellulose, cellulose, and lignin) occurs, followed by the combustion of char residue in Stage II [47]. Conversely, the second hypothesis considers that in Stage I the hemicellulose and cellulose are completely degraded, while lignin is partially decomposed in Stage I and also in Stage II. Therefore, Stage II would involve the completion of the volatilisation of the lignin together with the decomposition of the char. This second hypothesis has been recently proved by Ding *et al.*, demonstrating that the decomposition of lignin is a long-lasting process that contributes to both Stages I and II [34].

In the current study, the decomposition of hemicellulose, cellulose, and partial lignin was intuited during Stage I. Subsequently, lignin and char volatilization were supposed from 380 °C onwards during Stage II. This behaviour was validated through the evaluation of the mass loss for both stages. Coronado *et al.* studied the composition of brewer's spent grains (BSGs) in terms of hemicellulose, cellulose, and lignin and compared their results with those previously reported in the literature [48]. If their reported percentages are averaged, the general composition of the BSGs would result as 28.4% ($\pm 5.3\%$) hemicellulose, 20.6% ($\pm 3.4\%$) cellulose, and 13.6% ($\pm 1.9\%$) lignin. This composition would be equivalent to proportions of hemicellulose+cellulose to lignin close to 3.5/1. Considering the char yield for BSGs in purely pyrolytic processes of around 25% [46], Stage II in this study would involve between 7.5 and 17.5% of lignin decomposition contribution. Therefore, the proportion of the mass loss percentage for Stage I versus the remnant lignin in Stage II, *i.e.* $\Delta m_1/(\Delta m_2 - 25\%)$ shown in **Table 2**, corroborates that the decomposition of lignin started during Stage I and was completed during Stage II, demonstrating the second proposed hypothesis.

In terms of onset temperature (T_o), non-significant differences were found, and it can only be highlighted the lowest onset temperature for the Pilsener's BSGs, which started substantial mass-loss from 258 °C onwards. The peak temperatures of the main decomposition stage (T_{pl}) remained unaltered and mass loss contributions were quite similar for all the varieties. The only difference to be noted was in the Carafa's BSGs, in which the mass loss was a little lower than the average and the temperature of the peak was higher than that of all the other BSGs. Given that hemicellulose and cellulose are decomposed in this stage, it may be supposed a lower percentage of these components in such BSGs. In Stage II, non-significant changes among the analysed varieties could be identified, again except for the BSGs of Carafa malt. The peak temperature was relatively higher than the other malts and the mass loss in this stage was also bigger, suggesting a greater char production during combustion. This observation was also confirmed by the highest ash yield and lower volatiles content obtained in the proximate analysis section. The shift to higher temperatures of the Carafa's BSGs was also visible in the endset temperature (T_e).

At this point, the contribution of the heating rate to the thermo-oxidative stability of the BSGs was evaluated. As an illustrating representation, the thermogravimetric (TG) and derivative thermogravimetric curves (DTG) for Carafa's BSGs as a function of the heating rate are shown in **Figure 4**. Results for the other BSGs with an analogous pattern are shown in **Figures S1** and **S2** in the Supplementary Material. According to expectations, higher heating rates displaced both decomposition processes (Stage I and Stage II) towards higher temperatures during thermo-oxidative decomposition for all the BSGs regardless of the malt variety. Heat transfer during faster heating may be limited, and therefore, higher peak temperatures were found [25,49]. To avoid the influence of the dynamic heating program during thermogravimetric analyses, the thermo-oxidative decomposition behaviour (TDB)

when the heating rate tends to zero ($\beta \rightarrow 0$) was calculated, adjusting the results to the exponential relationship shown in **Equation 11**, where a , b and k are parameters of the fitting.

$$TDB(\beta) = \frac{a}{1 + b \cdot e^{-k \cdot \beta}} \quad (11)$$

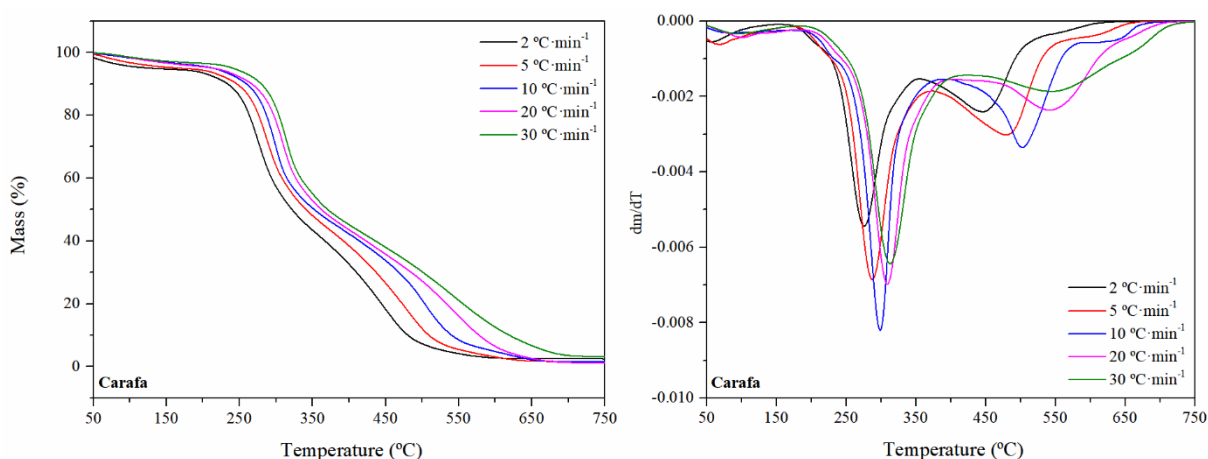


Figure 4. Thermogravimetric (TG) and first derivative (DTG) curves of Carafa's dry BSGs measured at $\beta = 2, 5, 10, 20,$ and $30 \text{ °C} \cdot \text{min}^{-1}$.

In particular, the zero-decomposition temperature (ZDT) for the onset (T_o), Stage I (T_{p1}), Stage II (T_{p2}), and endset (T_e) temperatures were assessed. For demonstrative purposes, the fitting of the model as a function of the heating rate for Carafa's BSGs is plotted in **Figure 5**. Results for the other BSGs with analogous patterns are shown in **Figure S3** in the Supplementary Material. **Table 3** gathers the obtained values along with the fitting parameters and regression coefficient for all the BSGs. Although most of the fitting lines showed a high degree of correlation (R^2) generally above 0.95, the BSGs should be compared with caution. While all the calculated values of ZDT give valuable information, the technological utility of the beginning and ending of the decomposition given by the onset and endset temperatures (T_o and T_e) is of great importance. Particularly, the values of the ZDT for the T_o were in the range from 212 to 247 °C and decomposition finished at ZDT for the T_e in

the span from 462 to 495 °C. Therefore, it can be concluded that the decomposition occurred in a temperature span of 250 °C regardless of the BSGs variety.

Table 3. Values of Zero Decomposition Temperatures (*ZDT*) along with the fitting parameters and regression coefficient for the thermo-oxidative decomposition of dry BSGs.

| | | <i>ZDT</i> (°C) | <i>a</i> | <i>b</i> | <i>k</i> | <i>R</i> ² |
|----------|----------------------|--------------------|----------|-----------|-----------|-----------------------|
| Carafa | <i>T_o</i> | 232 | 280±4 | 0.21±0.02 | 0.12±0.03 | 0.962 |
| | <i>T₁</i> | 271 | 314±1 | 0.16±0.01 | 0.11±0.01 | 0.996 |
| | <i>T₂</i> | 433 | 550±7 | 0.27±0.02 | 0.12±0.03 | 0.978 |
| | <i>T_e</i> | 495 | 700±63 | 0.49±0.26 | 0.04±0.03 | 0.940 |
| Aroma | <i>T_o</i> | 243 | 292±1 | 0.20±0.01 | 0.12±0.01 | 0.997 |
| | <i>T₁</i> | 264 | 316±2 | 0.20±0.01 | 0.13±0.01 | 0.993 |
| | <i>T₂</i> | 409 | 528±8 | 0.29±0.03 | 0.14±0.03 | 0.972 |
| | <i>T_e</i> | 464 | 612±27 | 0.32±0.04 | 0.08±0.03 | 0.948 |
| Crystal | <i>T_o</i> | 247 | 297±1 | 0.20±0.00 | 0.10±0.00 | 0.998 |
| | <i>T₁</i> | 269 | 320±2 | 0.19±0.01 | 0.10±0.01 | 0.992 |
| | <i>T₂</i> | 419 | 541±12 | 0.29±0.02 | 0.08±0.02 | 0.984 |
| | <i>T_e</i> | 462 | 624±10 | 0.35±0.02 | 0.06±0.01 | 0.997 |
| Pale Ale | <i>T_o</i> | 229 | 274±1 | 0.20±0.00 | 0.16±0.01 | 0.999 |
| | <i>T₁</i> | 263 | 313±4 | 0.19±0.02 | 0.11±0.03 | 0.976 |
| | <i>T₂</i> | 410 | 529±12 | 0.29±0.04 | 0.13±0.05 | 0.939 |
| | <i>T_e</i> | 463 | 644±40 | 0.39±0.07 | 0.06±0.03 | 0.964 |
| Trigo | <i>T_o</i> | 247 | 289±1 | 0.17±0.01 | 0.11±0.01 | 0.993 |
| | <i>T₁</i> | 265 | 315±3 | 0.19±0.02 | 0.13±0.03 | 0.976 |
| | <i>T₂</i> | 422 | 523±20 | 0.24±0.04 | 0.09±0.05 | 0.920 |
| | <i>T_e</i> | 479 | 628±75 | 0.31±0.13 | 0.05±0.04 | 0.901 |
| Pilsener | <i>T_o</i> | 212 | 267±2 | 0.26±0.04 | 0.22±0.05 | 0.968 |
| | <i>T₁</i> | 263 | 310±2 | 0.18±0.01 | 0.12±0.02 | 0.989 |
| | <i>T₂</i> | 428 | 539±26 | 0.26±0.05 | 0.07±0.04 | 0.935 |
| | <i>T_e</i> | 467 | 622±29 | 0.33±0.05 | 0.07±0.03 | 0.967 |

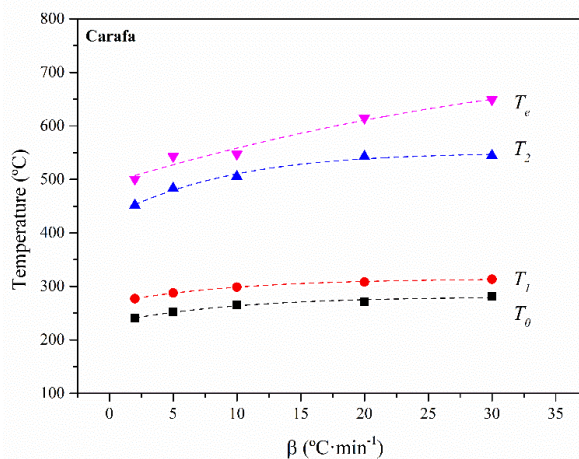


Figure 5. Evolution of the characteristic decomposition temperatures for Stages I and II (T_{p1} and T_{p2}) at all heating rates.

3.3. Kinetic analysis of the thermo-oxidative decomposition

The kinetic study of a given chemical reaction such as that of thermo-oxidative decomposition involves the (i) evaluation of the phenomenological kinetic laws that govern the reaction rate as a function of the concentrations of reactants and products or temperature; and (ii) the assessment of the mechanism of reactions (depolymerization or chain scission), or whether there is a series of stages that lead to the formation of intermediate species and products. Indeed, understanding the oxidative kinetics of biomass is essential for modelling and developing appropriate combustion methods. Although most of the approaches for the kinetic analysis proposed in the literature do not consider heat and mass transfer limitations and circumvent an exhaustive description of chemical reactions during decomposition, they offer a simplified reaction pathway valuable for the design of energy valorisation setups. In this regard, the theoretical decomposition mechanisms and their mathematical models are correlated by the kinetic functions.

The solid-state decomposition kinetics involves the calculation of activation energy (Ea), the pre-exponential factor (A), and the kinetic function $f(\alpha)$ evaluation, conforming to the so-called kinetic triplet. For this purpose, iso-conversional methods, Master-Plots, and Perez-

Maqueda *et al.* criterion were applied. In this regard, dynamic thermogravimetric analyses at different heating rates (2, 5, 10, 20, and 30 °C·min⁻¹) were considered.

3.3.1. Apparent activation energy

The apparent activation energy (E_a) was calculated using iso-conversional model-free approaches considering integral multiple- α methods, as described by Kissinger-Akahira-Sunose (KAS) and Flynn-Wall-Ozawa (FWO), and differential multiple- α models as the one proposed by Friedman. These methods were used for the calculation of the apparent activation energy as a function of the conversion degree (α) for the previously defined and main decomposition processes, *i.e.* Stage I (170-380 °C) and Stage II (380-700 °C). As an example, **Figure 6** shows the application of the FWO, KAS, and Friedman methods to the BSGs of the Carafa malt variety. All the fitting lines showed a high degree of correlation (R^2), especially for α in the range 0.2-0.6, where initiation and termination reactions are omitted. The results for the other BSGs with a comparable performance are shown in **Figures S4** and **S5** in the Supplementary Material for Stage I and Stage II, respectively.

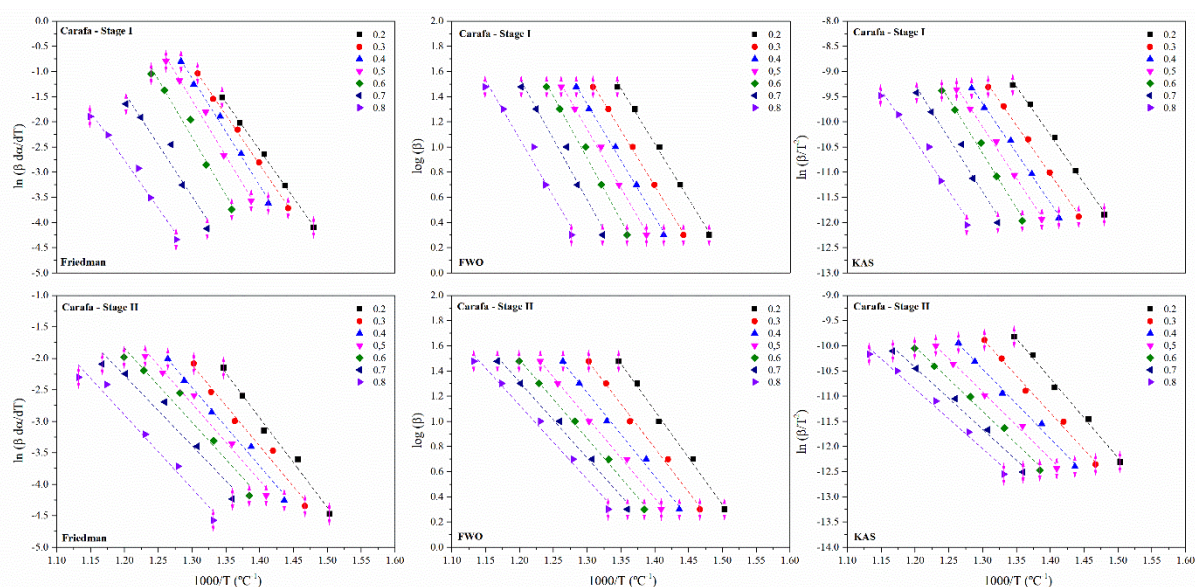


Figure 6. Application of the Friedman, FWO, and KAS and methods to the decomposition during Stage I and II for the Carafa's dry BSGs.

The obtained apparent activation energy (Ea) for the BSGs of all the malt varieties are shown in **Figure 7** as a function of the conversion degree, calculated as an average of the results obtained by the Friedman, FWO, and KAS methods. Moreover, the Ea was furtherly averaged in the α range from 0.2 to 0.8 and the obtained values are gathered in **Table 4** for the Stages I and II, separately. The apparent activation energy was fairly similar for the FWO and KAS methods, while were to some extent different for the Friedman method, also with higher average deviation. This could be due to the differential nature of this method that, after the required numerical differentiation, could give this slightly diverse result [25]. Overall, results were in line with other studies dealing with different types of lignocellulosic biomasses, in which activation energy is reported in the range 140-200 $\text{kJ}\cdot\text{mol}^{-1}$ for the given α range [50,51].

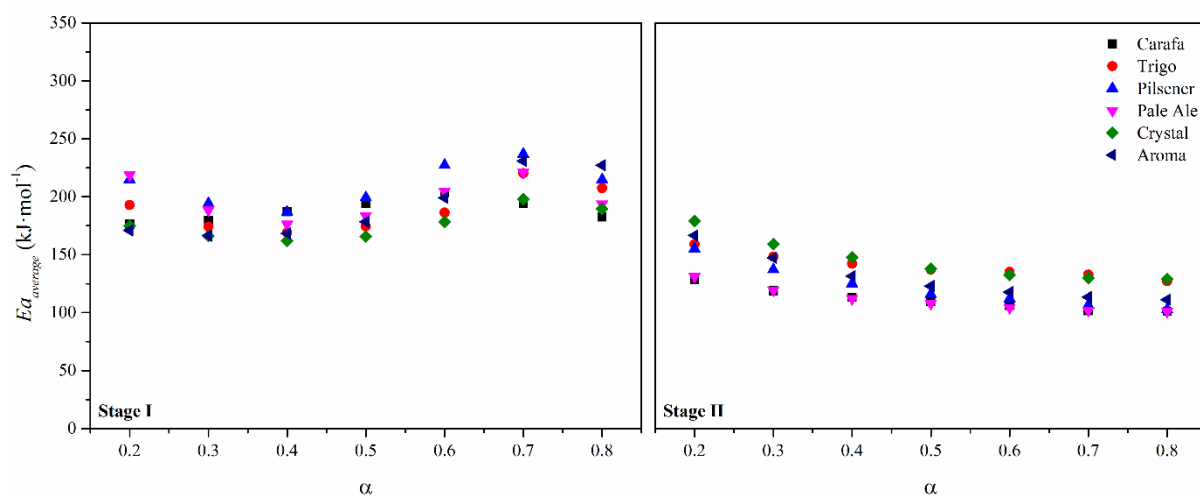


Figure 7. Average of the apparent activation energy (Ea) as a function of the conversion degree calculated through Flynn-Wall-Ozawa (FWO), Kissinger-Akahira-Sunose (KAS), and Friedman for all the dry BSGs during Stages I and II.

Table 4. Average of the apparent activation energy calculated in the range from $\alpha = 0.2-0.8$ through Flynn-Wall-Ozawa (FWO), Kissinger-Akahira-Sunose (KAS), and Friedman for all the dry BSGs during Stages I and II.

| | Stage I | | | Stage II | | |
|----------|---------------------------------------|---------------------------------------|--|---------------------------------------|---------------------------------------|--|
| | Ea_{FWO} (kJ·mol ⁻¹) | Ea_{KAS} (kJ·mol ⁻¹) | $Ea_{Friedman}$ (kJ·mol ⁻¹) | Ea_{FWO} (kJ·mol ⁻¹) | Ea_{KAS} (kJ·mol ⁻¹) | $Ea_{Friedman}$ (kJ·mol ⁻¹) |
| Carafa | 187±7 | 187±7 | 191±12 | 119±8 | 112±9 | 103±6 |
| Aroma | 186±19 | 187±20 | 202±34 | 139±17 | 134±18 | 117±14 |
| Crystal | 176±9 | 176±9 | 178±20 | 153±15 | 148±17 | 135±12 |
| Pale Ale | 198±12 | 199±12 | 197±24 | 119±9 | 113±10 | 102±6 |
| Trigo | 188±14 | 189±15 | 191±25 | 147±8 | 142±9 | 133±7 |
| Pilsener | 208±13 | 210±13 | 215±20 | 132±15 | 126±16 | 109±13 |

As expected, the activation energy for Stage I was significantly higher (190-200 kJ·mol⁻¹) than that of Stage II (100-150 kJ·mol⁻¹) due to the subsequent decomposition of pure components and char, respectively. Although slight differences were found for a given stage, the calculation of the apparent activation energy after the separation of the process into two main stages was demonstrated to be suitable for such biomass, which suffers heterogeneous reactions when subjected to thermo-oxidative decomposition [52,53].

3.3.2. Evaluation of the kinetic model

The use of reduced Master Plots (MP) to ascertain the mathematical decomposition model allow for obtaining the reaction mechanism during thermo-oxidative decomposition [25,54,55]. MP offer theoretical curves which are superposed to experimental data to select the adequate kinetic model for the given reaction. The normalized $MP_{g(\alpha)/g(0.5)}$ along with the experimental curves obtained with a heating rate of 10 °C·min⁻¹ for the different BSGs during Stage I and Stage II are represented in **Figure 8**.

When MP and experimental curves were compared, it was found that all the reaction mechanisms for Stage I were comparable to curve F_n type, with n equal to 2, 3, 4, and 5.

Physically speaking, F mechanisms represent solid-state processes, in which random nucleation with one, two, or three nuclei on the individual particle takes place. However, Stage II revealed a mechanism more difficult to assign. Although experimental curves were always between F_n and D_n models, the mechanism selected as the most appropriate to describe all the experimental results of Stage II was the D_4 type. This mechanism was furtherly corroborated in the next section and describes a solid-state process in which three-dimensional diffusion occurs. These identified mechanisms, which algebraic expressions for $f(\alpha)$ and $g(\alpha)$ are presented in **Table 5**, are in accordance to results reported in the literature, in which the mechanism of decomposition of lignocellulosic materials under oxidative conditions for temperatures in the range of those selected for Stage I (mostly attributed to hemicellulose and cellulose) is often identified as an F_n mechanism, while the decomposition of Stage II (attributed to remaining lignin and char) is sometimes described according to F_n , A_n or D_n mechanisms for different materials [24,25,56].

Table 5. Algebraic expressions for the differential $f(\alpha)$ and integral $g(\alpha)$ functions for the selected mechanisms of the solid-state processes.

| Mechanism | $f(\alpha)$ | $g(\alpha)$ | Solid-state process |
|-----------|---|--|--|
| F_1 | $1 - \alpha$ | $-\ln(1 - \alpha)$ | Random nucleation with one nucleus on the individual particle |
| F_2 | $(1 - \alpha)^2$ | $1/(1 - \alpha)$ | Random nucleation with two nuclei on the individual particle |
| F_3 | $\frac{1}{2}(1 - \alpha)^3$ | $1/(1 - \alpha)^2$ | Random nucleation with three nuclei on the individual particle |
| F_4 | $\frac{1}{3}(1 - \alpha)^4$ | $1/(1 - \alpha)^3$ | Random nucleation with four nuclei on the individual particle |
| D_4 | $\frac{3}{2}[(1 - \alpha)^{\frac{1}{3}} - 1]$ | $\left[1 - \left(\frac{2}{3}\alpha\right)\right] - (1 - \alpha)^{\frac{2}{3}}$ | Three-dimensional diffusion |

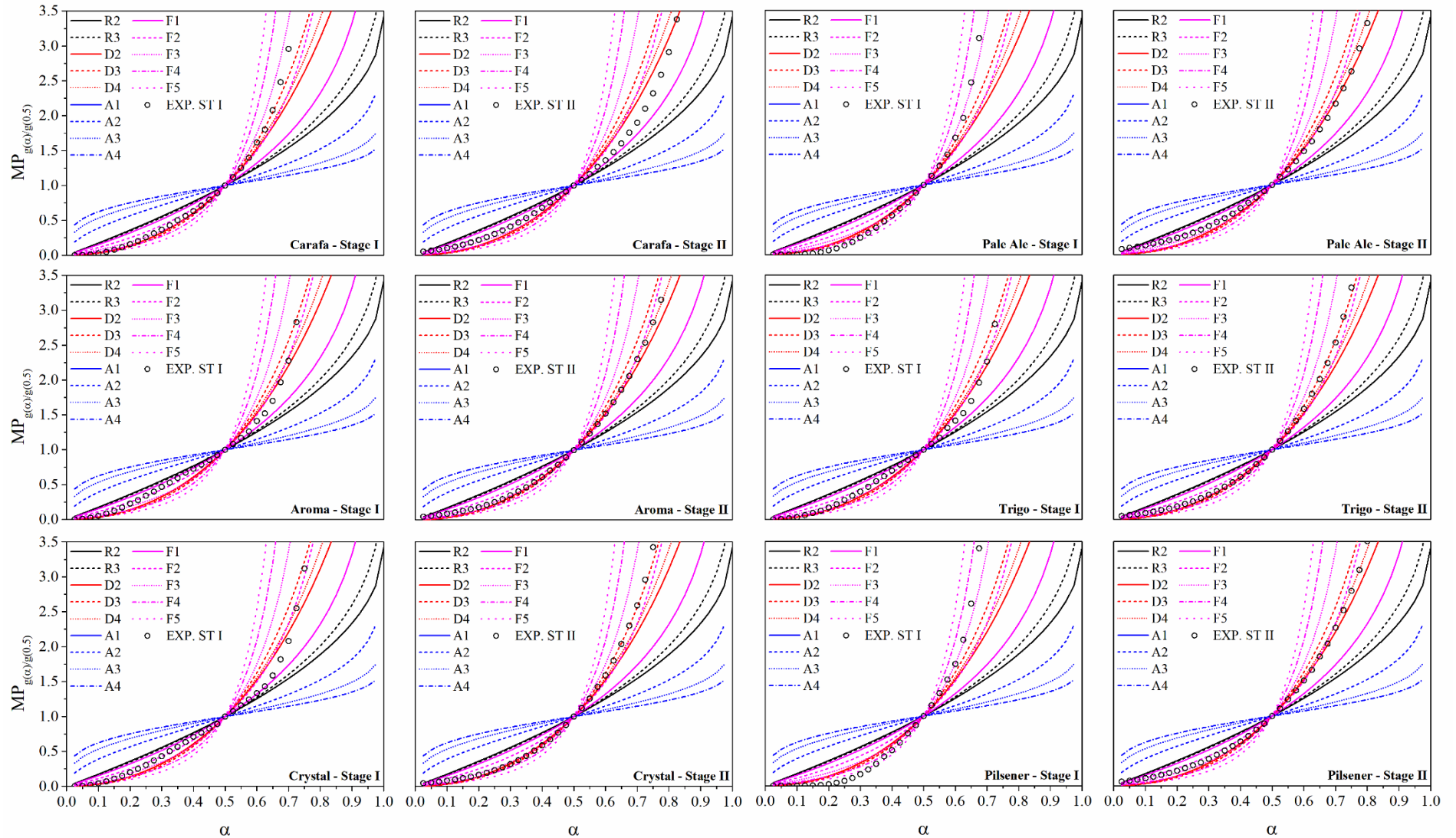


Figure 8. Comparison of the theoretical and experimental master plots (MP) for Stage I and Stage II of the different dry BSGs.

3.3.3. Determination of the pre-exponential factor and reaction order

Once the best theoretical mechanism was graphically evaluated, the pre-exponential factor (A) and the optimal analytical order of reaction (n) were assessed for the selected mechanism applying the Coast-Redfern equation (**Equation 9**). Using the average value calculated for the apparent activation energies and applying a regression method described in **Equation 10**, the order of reaction was obtained. Subsequently, $\ln A$ was gained from the intercept of plots of $\ln \left(\beta \frac{g(\alpha)}{T^2} \right)$ vs. $\frac{1}{T}$ for each heating rate. According to the Perez-Maqueda *et al.* criterion, the points should lie on the same straight line if the appropriate $g(\alpha)$ is selected. **Figure 9** illustrates these linearly coincident representations for the Carafa's BSGs in Stage I and Stage II, respectively. The Perez-Maqueda plots for the other BSGs are shown in **Figures S6** and **S7** in the Supplementary Material. The validity of the simplified kinetic triplet was therefore demonstrated, being its use advisable when the activation energy could be considered constant within a narrow confidence interval [55]. Moreover, the good fitting of the selected mechanisms (F_n and D_4) was confirmed. However, since the passages from Stage 0 to Stage I and, from Stage I to Stage II, are not always smooth due to complex initiation and termination reactions, respectively, a small deviation from linearity was noticeable at the edges. Moreover, it can be stated that the differences for a low degree of conversion during Stage I of the Pilsener and Pale Ale BSGs are explainable by the fact that at these temperatures some trace of remnant low molar mass compounds such as extractives may still be present at the beginning of this stage.

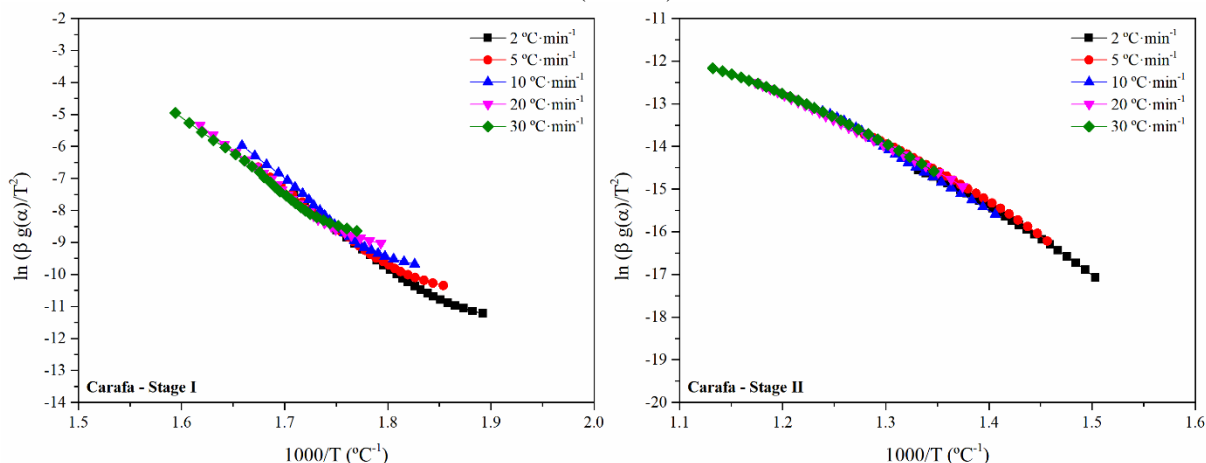


Figure 9. Application of Perez-Maqueda criterion for Carafa's dry BSGs.

Table 6 gathers the values of the Napierian logarithm of the pre-exponential factor ($\ln A$) calculated from the average of the A_β at the different heating rates, the most appropriate mechanism of reaction from the Master Plots (MP), and the calculated analytical order (n) for all BSGs during Stage I and Stage II. Moreover, e (%) represents the average of absolute deviations from the mean in each set of $\ln A_\beta$.

Table 6. Kinetic parameters for all the dry BSGs during thermo-oxidative decomposition in Stages I and II.

| | Stage I | | | | Stage II | | | |
|----------|---------|---------|-------|-----|----------|---------|----|-----|
| | $\ln A$ | e (%) | MP | n | $\ln A$ | e (%) | MP | n |
| Carafa | 40.0 | 6 | F3 | 2.8 | 13.0 | 17 | D4 | - |
| Aroma | 41.8 | 12 | F2-F3 | 2.2 | 16.4 | 21 | D4 | - |
| Crystal | 38.1 | 5 | F2 | 1.9 | 19.1 | 18 | D4 | - |
| Pale Ale | 40.9 | 7 | F3 | 3.1 | 13.4 | 17 | D4 | - |
| Trigo | 39.5 | 4 | F2-F3 | 2.2 | 18.9 | 10 | D4 | - |
| Pilsener | 45.6 | 5 | F3-F4 | 3.7 | 15.3 | 21 | D4 | - |

A random nucleation kinetic model (F_n) was determined for Stage I of all the BSGs, while Stage II was described by a three-dimensional diffusion kinetic model (D_4), similarly to other







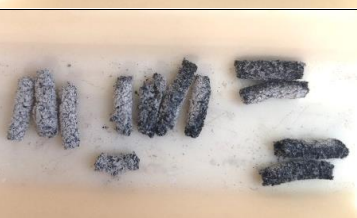







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lignocellulosic materials. In the order-based models, the reaction rate is proportional to the concentration of reactant raised to a particular power (integral or fractional) which is the reaction order. Conversely, in diffusion-controlled reactions, the rate of product formation decreases proportionally with the thickness of the product barrier layer [57]. Overall, the obtained results were comparable for all the analysed BSGs in the two stages, both in terms of pre-exponential factor and kinetic function, and suggest, along with the comparable calculated activation energy found in previous sections, that BSGs could be used as solid biofuels, both separately and combined, regardless of their variety or treatment.

3.4. Pelletised BSGs as combustion fuel

Given the minor perceived differences in terms of thermo-oxidative decomposition among the BSGs of the different malt varieties, the BSGs were all pelletised together. Subsequently, for validation purposes, the pellets were preliminarily subjected to isothermal combustion in a muffle furnace at different temperatures ranging from 400 to 700 °C. These temperatures were selected according to the results found in the thermo-oxidative stability analyses, which revealed complete combustion from ~500 °C onwards. Temperatures below 500 °C were selected for monitoring purposes. After combustion, the ash yields were calculated and their appearance was evaluated, both gathered in **Table 7**.

Table 7. Ash percentage and appearance of the pellets of BSGs after combustion at different temperatures.

| Temperature (°C) | Ash (%) | Appearance before agitation | Appearance after agitation |
|------------------|---------|---|--|
| 400 | 7.05 |  |  |
| 450 | 3.33 |  |  |
| 500 | 3.39 |  |  |
| 550 | 2.52 |  |  |
| 600 | 2.64 |  |  |
| 650 | 2.51 |  |  |
| 700 | 2.44 |  |  |

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In general, ash yields below 3.5% for combustion above 450 °C were found, which slightly decreased with higher temperatures up to 2.4% at 700 °C. The ash percentage was comparable to that found for the combustion of the grounded BSGs measured according to normalised conditions. However, the experiments carried out at 400 °C revealed partial combustion, with a 7.0% of residue. As a whole, the shape of the pellets remained unaltered after combustion, which turned into powder after slight agitation of the tray, especially at temperatures higher than 500 °C. It must be highlighted that the combustion at temperatures below 500 °C resulted in the generation of more carbonaceous ash, given the blackish colour of the powder and more consistent morphology. On the other hand, above 500 °C, pure ash with a greyish colour was obtained, typical for the complete combustion of biomass. In this line, the combustion of BSGs in gasification processes at temperatures above 500 °C would allow for the lowest ash yield and a powder morphology, manageable in automated setups [58] and useful for a variety of further applications [59].

This lab-scale behaviour was validated in an incineration pilot plant, in which 1 kg of BSG pellets was applied. A pellet feeding rate of $\sim 1 \text{ kg}\cdot\text{h}^{-1}$ and airflow of $50 \text{ L}\cdot\text{min}^{-1}$ (21% O₂) were considered to ensure that combustion was accomplished at temperatures above 500 °C. Although flames in these experimental tests experience large temporal and spatial gradients in local species and temperatures, they remained virtually constant during the whole operation, as observed through the viewing port and shown in **Figure 10** [60].

Once the pellets were consumed, the incineration unit was opened, and the remnant ash in the combustion chamber was collected into the ashtray in the bottom, as shown in **Figure 10**. The combustion ash revealed a greyish powder appearance, as found after muffle furnace combustion assays. Given the contribution of the moving grill in a circular motion of the incineration unit,

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powder together with particles lower than 1.5 mm was obtained, as can be observed in **Figure 10**.

For the considered amount of 1 kg of pellets, 24.12 g of ash were gained, which represents a yield of 2.4%. This result is in accordance with that found in the muffle furnace experiments at temperatures above 500 °C, and also with that determined in the proximate analysis. Moreover, given its powder morphology, the ash revealed an apparent density of 0.37 g·mL⁻¹. Therefore, the reliable combustion behaviour, along with the higher heating value of ~18 MJ·kg⁻¹ found in previous sections, makes this biomass a potential candidate for being used as a source of energy in combustion incineration devices. Although ash had not adhered in the walls of the incineration unit and could be properly managed, further characterisation of ash would allow for a deeper evaluation in terms of ash formation and deposition mechanisms, particle properties or sticking behaviour [61].



Figure 10. Flame during combustion of pellets of dry BSGs (left), residue into the incineration unit after combustion (middle), and collected ash (right).

4. Conclusions

The constituting elements of the analysed beer spent grains (BSGs) are mainly Carbon and Oxygen. This, along with a low percentage of Nitrogen and a null Sulphur content makes these materials attractive for combustion purposes, providing a major environmental advantage for the

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use of this biomass as a thermo-oxidation energy vector instead of fossil fuels. With an exhaustive control of the combustion temperature, the products will be mainly carbon dioxide and monoxide, while the production of unwanted nitrogenous compounds will be in non-significant proportions.

All the BSGs have similar volatiles and fixed carbon percentages, with low ash yield, that avoids operational difficulties during combustion. The acquired analogous superior higher heating value (*HHV*) around $18 \text{ MJ}\cdot\text{kg}^{-1}$ for all the BSGs highlights the benefits of using this biomass in thermo-oxidation reactions.

The global decomposition process has two main stages in the range of 212 to 495 °C for all the BSGs. Stage I corresponds to the thermo-oxidation of hemicellulose, cellulose, and partially lignin. It follows a random nucleation kinetic model (F_n) and has a mean apparent activation energy of $\sim 190 \text{ kJ}\cdot\text{mol}^{-1}$. Stage II is related to the thermo-oxidation of and remnant lignin and char. It is described by a three-dimensional diffusion kinetic model (D_4) with a mean activation energy of $\sim 120 \text{ kJ}\cdot\text{mol}^{-1}$. These parameters can be applied as input variables in a theoretical model to describe the thermal recovery processes in a reactor using BSGs as feedstock.

The appropriate combustion process of pelletised BSGs carried out in a pilot spouted bed reactor corroborate the viability of this reaction, along with the low ash formation when combustion was carried out above 500 °C. Overall, this BSG waste biomass may be considered a beneficial source of renewable energy in thermo-oxidation reactions, regardless of the malt variety.

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Data availability

The data that support the findings of this study are available from the corresponding author upon reasonable request.

Conflict of interest

The authors declare no conflicts of interest.

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