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***Helicrysum italicum* (roth) G. Don, a promising species for the phytostabilization of polluted mine sites: a case study in the Montevecchio mine (Sardinia, Italy).**

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Abstract

Mine exploitations worldwide have generated a great amount of tailings, which still contain large quantities of Potentially Harmful Elements (PHEs) able to contaminate soil, water, air, wildlife, and the food chain. Phytoremediation is an option to immobilize and/or extract PHEs from polluted mining areas. This study aims at assessing the phytoremediation properties of *Helicrysum italicum* (roth) G. Don, and in particular the capacity to absorb, transfer and accumulate some PHEs, such as Cd, Cu, Ni, Pb and Zn, in the plant tissues. A restricted literature review (7 papers) is also proposed in order to outline the *H. italicum*'s behaviour and verify its possible use in phytoremediation strategy of polluted mine soils. A number of 22 contaminated sites from Montevecchio mine area (Sardinia, Italy) were sampled and the results compared with 6 uncontaminated sites. In each site both rhizospheric soil materials and *H. italicum* plants were sampled. Total composition and bioavailable fractions were analysed on soil samples. *Helicrysum italicum* roots, stems and leaves were separately analysed to extract PHEs; root/soil and leaf/root ratios were calculated to elucidate plant behaviour. Results show that Cd and Zn are the most bioavailable PHEs in contaminated sites compared to uncontaminated ones (300 and 500 folds, respectively), while Cd, Cu, Pb and Zn exceed the thresholds of the Italian environmental legislation when aqua regia extraction is executed. *Helicrysum italicum* plants growing on contaminated soils accumulate significantly more Cd, Ni, Pb and Zn than plants growing on uncontaminated soils, while no significant differences are found for Cu. For all considered PHEs the root/soil ratios are > 1 in both contaminated/uncontaminated sites meaning that *H. italicum* can be defined as a tolerant species for remediation of metal-polluted soils. The leaf/root ratios weakly > 1 of Zn, Pb and Cu and < 1 of Ni and Cd indicate *H. italicum* not suitable for phytoextraction. Our results are aligned with the available literature indicating *H. italicum* a tolerant species, especially for Cd,

36 Pb and Zn. The low leaf/root ratios, along with its being a spontaneous and perennial species
37 able to propagate seeds directly on contaminated soils, recommended the use of *H. italicum*
38 as pioneering strategy for the phytostabilization.

39

40 **Keywords**

41 *Helicrysum italicum (roth) G. Don*, Potentially Harmful Elements, phytoremediation, mine soils,
42 literature review

43

44 **1. Introduction**

45 Every historical civilization' advancement has required mined resources and also future human
46 development and green technologies will depend on extensive Earth-extracted resources
47 (Mills, 2020). Mining activities are widespread in the world, locally modifying the original
48 environment and impacting biota even many years after their dismissal (Camizuli et al., 2018).
49 Estimations for Europe, China and the USA reveal that about 0.1% of the land is represented
50 by abandoned mining sites (Arbogast et al., 2000; EUROSTAT, 2012; Lin and Ho, 2003; Perez,
51 2012). Mine exploitation generates large amounts of tailings, which can still contain high
52 quantities of Potentially Harmful Elements (PHEs). Tailings are poorly colonized by vegetation
53 because of unfavourable chemical-physical conditions for plant growth (Martínez-Sánchez et
54 al., 2012), thus PHEs could easily spread out in the environment contaminating soil, water, air,
55 and wildlife reaching also the food chain (Dore et al., 2020). In many cases PHEs are essential
56 elements at a low concentration (e.g., Cu, Zn, Mn, Fe) that turn to be toxic to one or more
57 species when reaching higher concentrations (Bini and Wahsha, 2014; Vamerali et al., 2010).
58 Other elements (e.g., As, Cd, Cr, Ni, and Pb) show toxic impacts on plants and animals even
59 at low concentrations (Singh et al., 2011).

60 Remediation is the only intervention to reduce or definitively solve the environmental
61 contamination problem. Conventional remediation technologies are often expensive, labour-
62 intensive, destructive, and not eco-friendly (Meuser, 2013; Yao et al., 2012). On the other hand,
63 phytoremediation, defined as the use of plants for degradation of xenobiotics or
64 extraction/immobilization of PHEs from water or soil substrates (USEPA, 2000), is cost-
65 effective, widely acceptable, sustainable, applicable in large areas, and economically
66 exploitable, particularly when native plants are used (Pandey et al., 2015; 2016).

67 Plants for phytoremediation must be resistant to both contaminants and unfavourable climatic
68 conditions like drought and heat, especially in Mediterranean areas (Poschenrieder et al.,
69 2012), and should display a high growth rate to absorb considerable quantities of toxic

70 elements in their tissues (Mendez and Maier, 2008). For these reasons, plants are often native
71 of the environment in which they will be used (Yoon et al., 2006).

72 Phytoremediation includes five types of strategies adopted by plants:
73 phytoextraction/phytoaccumulation, phytostabilization, phytodegradation, rhizofiltration and
74 phytovolatilization; only the first two are considered for PHEs (Mahar et al., 2016; Pandey and
75 Bajpai, 2019). As reported by Mendez and Maier (2008), plants for phytostabilization should
76 accumulate PHEs in roots and not transfer them to shoots, to avoid further transfer into the
77 food chain. Instead, plants eligible for phytoextraction should be tolerant to PHEs, absorb and
78 accumulate them in the aboveground plant parts, grow fast and be easy to harvest (Mendez
79 and Maier, 2008).

80 A plant's phytoremediation capacity is generally assessed by means of a large number of
81 different quantitative indicators (Buscaroli, 2017). These are calculated as ratios between
82 element contents in aerial parts and roots, or as ratios between element contents in plant parts
83 and soil. In literature, the element concentration in soil is assessed by adopting different
84 analytical procedures such as X-ray fluorescence or by several wet extraction methods (e.g.,
85 Aqua regia, EDTA, DTPA, etc.) and thus, resulting ratios between plant parts could significantly
86 differ. Abreu et al. (2008) defined the Bioconcentration Coefficient (BC) the ratio between the
87 element content in leaves and available fraction of the corresponding soil element, extracted
88 with DTPA aqueous solution. Plants are considered tolerant when the BC value is greater than
89 1 ($BC > 1$). Similarly, Sidhu et al. (2017) named the same abovementioned ratio
90 Bioconcentration Factor (BCF) and stated that the BCF values >1 indicate the potential of a
91 plant species for remediation of metal polluted soils. Regardless of the name, when the root/soil
92 ratio is > 1 , the plant is considered a tolerant species (Abreu et al., 2008) useful for remediation
93 of metal polluted soils (Sidhu et al., 2017). Moreover, when the leaf/root ratio is > 1 the element
94 is efficiently transferred from roots to shoots proving that the plant is a phytoextractor while, if
95 the leaf/root ratio is < 1 no element translocation occurs, and the plant is suitable for
96 phytostabilization (Bolan et al., 2011).

97 *Helicrysum italicum* (roth) G. Don is a perennial subshrub of the genus *Helicrysum* of the
98 family Asteraceae, characteristic of the Mediterranean area, and it grows on barren, dry, sandy
99 and poorly developed soils in a wide altitudinal range from the sea level up to 2200 m a.s.l.
100 (Galbany-Casals et al., 2011; Ninčević et al., 2019). The scientific and industrial interest for *H.*
101 *italicum* is increasing due to its rusticity, versatile biological activities, cosmetic and
102 pharmaceutical applications, and ornamental uses (Bianchini et al., 2009; Melito et al., 2015;
103 Ninčević et al., 2019). Moreover, for the utilization and commercialization of derivatives from
104 *H. italicum* the European Union requires certified absence of chemical impurities and heavy
105 elements (Bullitta et al., 2010).

106 *Helicrysum* spp., are indicated as metallophyte, metal tolerant plants growing on soils enriched
107 or contaminated by several elements (Nkoane et al., 2003; 2007; Koosaletse-Mswela, 2015).
108 *Helicrysum* spp. have already been considered for bio-remediation purposes in contaminated
109 mine tailings (Bacchetta et al., 2017; 2018; Bini et al., 2017; Barbafieri et al., 2011; Cao et al.,
110 2004; Leita et al., 1989, studied *H. italicum*) and soils (Brunetti et al., 2018, studied *H. italicum*)
111 in Italy and in many other regions of the world (Conesa et al., 2006; 2011; García et al., 2002,
112 2005, studied *H. decumbens*; Fitamo and Leta, 2010, studied *H. odoratissimum*; Hesami et
113 al., 2018, studied *Helichrysum* Spp.). However, only the recent research paper by Brunetti et
114 al. (2018) considered as many PHEs as this study, even if Brunetti's work was conducted as a
115 pot experiment, while this work is an in situ experiment. The other papers alternatively
116 investigated Pb, Zn, Cd and sometimes Cu. For these reasons there is still a lack of knowledge
117 upon PHEs uptake and translocation in *Helicrysum* spp. in different environments, as well as
118 their interaction mechanisms.

119 A previous study, conducted with the same criteria, was performed by Buscaroli et al. (2017)
120 on *Dittrichia viscosa*, another rustic plant growing on the Montevecchio mine tailings. The ability
121 of phytostabilization and translocation shown by *D. viscosa* in this environment, justifies also
122 the interest for *H. italicum*.

123 The aim of this study is to evaluate *H. italicum* for phytoremediation applications in metal-
124 contaminated sites. Major and trace elements total concentrations in soil samples were
125 measured, while Cu, Cd, Fe, Ni, Pb and Zn were also quantified as bioavailable soil fractions
126 to be compared with total amounts extracted from plants. Detailed objectives of this study are:
127 i) to assess elements accumulation potential and interaction mechanisms in different parts of
128 *H. italicum*; ii) to study elements uptake capability in the roots and the translocation to aerals
129 plant parts; iii) to evaluate differences in plant behaviour in contaminated and uncontaminated
130 sites; iv) to compare elements concentration and phytoremediation properties with the existing
131 literature for *H. italicum* subspp.

132

133 **2. Materials and methods**

134

135 **2.1. Site description and sample collection**

136 Mining activities related to Pb and Zn extraction have been representing the main economic
137 activity for centuries in the South-West Sardinian mining districts (Italy), largely impacting the
138 environment (particularly soil and water) and landscape (Boni et al., 1999; Dore et al., 2020).
139 In this area the mining activities ended in 1991 leaving many abandoned heaps of waste
140 materials now exposed to gravity movements and water and wind weathering. It is estimated

141 that about 297 hectares are occupied by landfills and about 4.9 million m³ is the volume of
142 abandoned heaps. The cost for reclaiming activities is estimated at more than 485 million Euros
143 (Italian Government, 2001). The mining area has been included in the Italian list of polluted
144 sites since 2001 but, until today, no remediation activities have occurred.

145 In this work two broad areas were selected for the sample collection: one including the
146 contaminated sites in the Montevocchio area (CS) and the other including the uncontaminated
147 reference sites either close to the mining area or in the Emilia-Romagna area (US) (Fig. 1).

148 The CS were entirely located in the Montevocchio mining district, in the Southwestern Sardinia,
149 close to Montevocchio and Ingurtosu villages (Fig. 1A). The bedrock consists of low-grade
150 meta-sedimentary and meta-volcanic Cambrian-Ordovician rocks, with intrusions of Arburese
151 igneous complex occurred at the end of the Hercynian orogeny (Cuccurru et al., 2016; Moroni
152 et al., 2019 and references within). This complex is constituted by granodiorite and leucogranite
153 with radial fractures filled with acid and basic magmatic dykes, and with quartz and
154 metalliferous hydrothermal deposits exploited by the Montevocchio-Ingurtosu mines (Moroni
155 et al., 2019). The ore-veins are composed of galena, sphalerite and quartz with local intrusions
156 of carbonates (Moroni et al., 2019). During the mines' activity (1848 - 1991), approximately 3
157 Mt of Pb and Zn were extracted from the Montevocchio district. Nowadays several uncontrolled
158 waste rock piles generate relevant sources of contamination (Caboi et al., 1993; Concas et al.,
159 2006) due to the scarce vegetation cover and intense erosion.

160 The US include 2 sampling sites near but outside the Montevocchio mining district and 4 in the
161 Appennine chain between eastern Emilia-Romagna and Tuscany (Fig. 1B). These sites are
162 developed on different types of sedimentary materials. The two US near Montevocchio area
163 were sampled on aeolian sandy deposits. In the Appennine, the bedrock is made up of
164 alternations of sandstones and marls (Marnoso-Arenacea Formation) followed by a thin band
165 of evaporitic gypsum (Gypsum Vein), formed during the Messinian salinity crisis. Close to the
166 plain there are Pliocene clays and Pleistocene yellow sands (Lancianese and Dinelli, 2015).
167 Among the four sampling sites, two were chosen on evaporitic gypsum (1 and 2 in Fig. 1B) and
168 the other two on the Marnoso-Arenacea Formation (3 and 4 in Fig. 1B).

169 In these two broad areas, 28 sample sites were selected and sampled. At each site the entire
170 *H. italicum* plants and a composite rhizospheric soil, sampled at a depth of 5-30 cm, were
171 collected, stored in plastic bags and brought to the laboratory for analysis. Overall, 22 soil
172 samples and 22 plants were collected in CS corresponding to the major mine tailings deposits
173 (Fig. 1A), while 6 soil samples and 6 plants were collected in US (Fig. 1A and 1B).

174

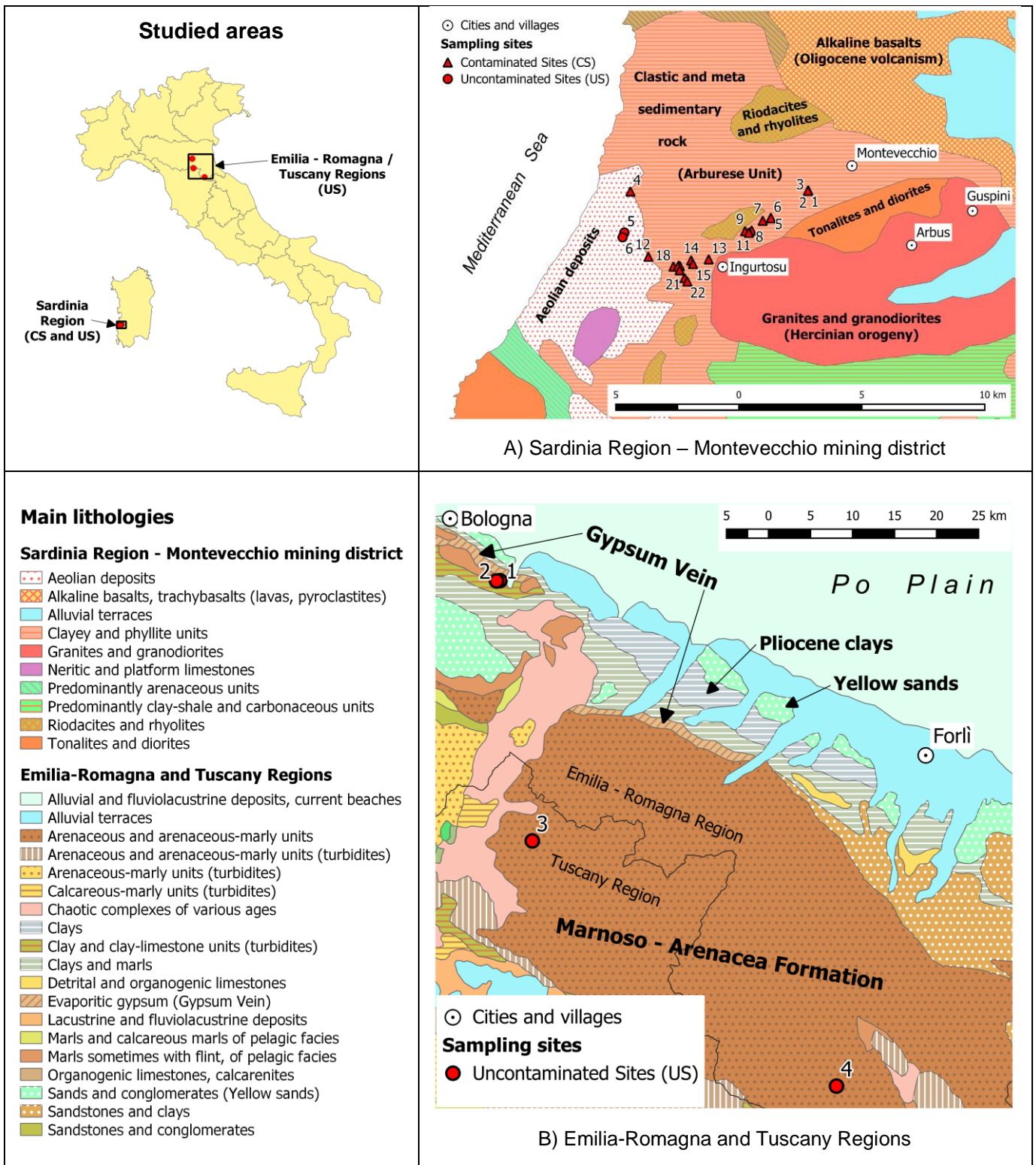


Figure 1. Location of sampling sites in the two study areas (a and b) with their main lithologies and related legend. Modified from ISPRA (<http://sgi2.isprambiente.it/arcgis/rest/services/servizi/cartageologica500k/MapServer>).

175

176 2.2. Chemical analysis of solid material

177 Soil material was air-dried at room temperature for two weeks, crushed and sieved through a
 178 2 mm screen, and this fraction was retained for analysis. Soil sand, silt and clay determination
 179 was performed according to the hydrometer method devised by Day (1965). Soil reaction (pH)

180 and electrical conductivity (EC) were determined in a 1:2.5 (m/V) soil-water suspension. The
181 total limestone was determined by volumetric method according to ISO 10693 method (ISO,
182 1995a). Total Organic Carbon (TOC) and Total Nitrogen (TN) were determined by elemental
183 analyser Thermo Fisher EA Flash 2000 CHNS-O according to ISO 10694 method (ISO,1995b).
184 In order to estimate the organic and carbonate content in soil, the Loss of weight On Ignition
185 (LOI) was determined by placing a soil sample in a muffle furnace at 950 °C for 6 h and
186 measuring the percentage weight loss, as indicated by Heiri et al. (2001).

187 In this paper the soil PHEs concentrations were analysed using three different methodologies:
188 i) extraction with a diethylenetriaminepentaacetic acid-based solution (hereafter DTPA)
189 (element available for root uptake); ii) aqua regia extraction (hereafter AR) (pseudo-total
190 element concentration); iii) X-Ray Fluorescence (hereafter XRF) (total element concentration)
191 (Wang et al., 2021).

192 The bioavailable elements in soil (Cd, Cu, Fe, Ni, Pb and Zn) were extracted with a DTPA-
193 based solution according to Lindsay and Norvell (1978) and ISO 14870 methods (ISO, 2001).
194 Concentrations of Fe, Zn, Pb for all sites and Cd only in CS were measured by Flame Atomic
195 Absorption Spectrometry (FAAS) using a Perkin-Elmer Analyst 100. Instead, concentration of
196 Cu, Ni for all sites and Cd only in US were measured by Graphite Atomic Absorption
197 Spectrometry (GFAAS) using a Perkin-Elmer HGS-800. The analyses were performed at
198 Biological, Geological, and Environmental Sciences Department (BiGeA Dept.) of Bologna
199 University using calibration standards from 0.2 to 5 mg/l and from 2 to 100 µg/l for FAAS and
200 GFAAS, respectively. As reported by Kumpiene et al. (2017), this methodology is widely
201 applied on non-acid soils and fitted perfectly to the US but it could be not appropriated for a
202 few soil samples with low pH in CS. However, in order to obtain comparable results, DTPA
203 extraction was adopted for all soil and tailing samples.

204 Aqua regia extraction was executed following the ISO 11047 method (ISO, 1998) and As, Cd,
205 Co, Cr, Cu, Fe, Ni, Pb, V and Zn were quantified by Inductively Coupled Plasma Mass
206 Spectrometry (ICP-MS) Perkin-Elmer ELAN DRC-eat at BiGeA Dept. of the Bologna University
207 with calibration standards from 0.01 to 5 mg/l.

208 The total concentrations of major and trace elements in soils were determined by X-Ray
209 Fluorescence (XRF) using a Panalytical Axios 4000 spectrometer, following the analytical
210 methodology by Franzini et al. (1972) and Leoni and Saitta (1976) for matrix correction
211 methods.

212 To ensure accuracy and precision in soil PHEs analysis, reagent blanks and certified reference
213 materials were used. Quality control of DTPA-extractable PHEs was performed analysing the
214 NCSDC85102a certified reference material. The obtained recoveries (mean value ± standard

215 deviation in %) are the following: Cd 102 ± 2 , Cu 108 ± 4 , Fe 90 ± 1 , Ni 91 ± 2 , Pb 80 ± 3 , Zn
216 70 ± 1 .

217

218 **2.3. Chemical analysis of plants**

219 After the separation into roots, stems and leaves, the plant samples were placed in an
220 ultrasonic bath to remove soil particles, washed with deionized water, dried in a stove at 40 °C
221 and minced.

222 For the determination of the total element concentrations, 250 mg of each plant part was
223 digested with a mixture of 2 ml H₂O₂ 30% and 6 ml HNO₃ 65% using a microwave Milestone
224 mls 1200 Mega. After the filtration the digested solutions were stored in 50 ml volumetric flasks.
225 In the plant's digested solutions Fe, Zn, Pb for all sites and Cd only in CS were analysed by
226 FAAS (Perkin-Elmer Analyst 100) while Cu, Ni for all sites and Cd only in US were analysed
227 by GFAAS (Perkin-Elmer HGS-800). The analyses were performed at BiGeA Dept. of the
228 Bologna University using calibration standards from 0.2 to 5 mg/l and from 2 to 100 µg/l for
229 FAAS and GFAAS, respectively.

230 As for soil, also for plant reagent blanks and certified reference materials were used.

231 Quality control on total PHEs concentrations in plants was performed analysing the IAEA – 359
232 (Cabbage - Trace elements) certified reference material in three replicas. The obtained
233 recoveries (mean value \pm standard deviation in %) are the following: Cd 127 ± 5 , Cu 126 ± 8 ,
234 Ni 96 ± 14 , Zn 105 ± 7 .

235

236 **2.4 Data quality control and statistical analysis**

237 On the results, several statistical elaborations were performed using R Software version 3.3.2:
238 the Shapiro Wilk test (Shapiro and Wilk, 1965) was adopted to verify the normal distribution of
239 the data; a Mann-Whitney non-parametric test was performed both between the different
240 sampling sites and the different plant tissues for the evaluation of the statistical significance of
241 the difference between the medians. The Spearman Rank Correlation Coefficient (SRCCs)
242 (Spearman, 1904) was applied in the determination of correlations between elements in soils
243 (both total and bioavailable) and in plants from CS using the R software version 3.3.2. The
244 obtained significance correlations were compared with < 0.05 and < 0.01 p-value.

245

246

247 **3 Results and discussion**

248

249 **3.1 Soils**

250 The main compositional and textural features for the analysed soils are summarized in Table
251 1 (the entire dataset is available in Table S1 and S2 of the Supplementary Materials). The CS
252 revealed a sandy loam texture with little silt and clay, whereas in the US, silt and sand were
253 the dominant soil fractions, although a wide variation in texture was present ranging from sand
254 to silty-loam. The CS showed a neutral-sub acid pH (5.4 - 6.7 as Min-Max range), while in the
255 US subalkaline to alkaline pH were observed (7.6 – 9.1). EC was below 0.3 dS/m in all sites,
256 while total lime was significant in the US and negligible in almost all the CS. In both the CS and
257 US, TOC range was large (2 – 60 g/kg) with median values around 25 g/kg. Median values of
258 TN content were identical (2 g/kg, for both CS and US), while the high TOC/TN ratio values (>
259 12) could be affected by *H. italicum* plant residues that hardly decompose in soils (Brady and
260 Weil, 2010).

261 The median concentrations of major elements such as Si (282938 mg/kg), Al (89741 mg/kg),
262 Fe (40509 mg/kg) and K (30018 mg/kg), confirmed the abundance of sheet silicates in CS
263 (Table 1). Compared with previous results by Buscaroli et al. (2017), the dominant silicates
264 feature was confirmed with weak increase in Fe and LOI median values and weak decrease in
265 Ti, Al, Mg, Na and K median values. Instead, median concentrations of Ca and LOI (129691
266 and 211908 mg/kg, respectively) had the highest values in US reflecting the calcareous
267 contribution of the substrate in the area as recorded in stream sediment by Lancianese and
268 Dinelli (2015).

269 In terms of total concentration (XRF) the CS were characterized by high Pb and Zn contents
270 (median 604 mg/kg and 11455 mg/kg, respectively). Compared with soils by Buscaroli et al.
271 (2017) from the same area, soils from this study had slightly higher Pb and Cu median values,
272 whereas Zn median concentration was double. The Ni total content (39 mg/kg) agrees with
273 values found by Buscaroli et al. (2017). Compared with soil samples from the adjacent
274 Ingurtosu mine area (Barbafieri et al., 2011), this study denoted around double Zn and half Pb
275 content (7800 and 1800 mg/kg, respectively). Mean total PHEs concentrations found in this
276 study area were comparable with concentrations of the nearby Barraxiutta mine district (De
277 Agostini et al., 2020) and Campo Pisanu (Bacchetta et al., 2018), except for Cu and Pb that
278 were 5 and 4 times lower, respectively. Overall, this high variability in values denoted the strong
279 heterogeneity of mine tailings.

280 The US had low median total concentration of PHEs with only Cr, Ni, V and Zn ranging around
281 100 mg/kg (Table 1). These results were aligned to the soils from the GEMAS project (Reimann
282 et al., 2014), to the streambed sediments of the same area analysed by Lancianese and Dinelli
283 (2015), and also to the natural background maps by Regional Soil Service, although related to
284 the adjacent plain area (RER, 2016).

285 Results of the AR extraction showed Cd and Zn median values (67 and 9771 mg/kg,
286 respectively) much higher than the Italian environmental legislation thresholds (Italian
287 Government, 2006) for an industrial use in CS (15 and 1500 mg/kg, respectively). If the
288 thresholds for public and residential destinations are considered (for As, Cd, Co, Cu, Pb and
289 Zn are 10, 2, 20, 120, 100, and 150 mg/kg, respectively), all median values exceeded the limit
290 and for Cd, Pb and Zn all samples were above the thresholds (bold values in Table 1). The US
291 samples did not reveal concentrations exceeding the national environmental prescriptions,
292 except for mean and maximum As concentration (16 mg/kg and 36 mg/kg, respectively). Being
293 part of a dominant calcareous unit, the high As concentration in US is determined by a direct
294 control of calcite abundance on As bio-accessibility (Raimondi et al., 2021).

295 Regarding bioavailable element concentrations (DTPA-extracted) (Table 1), the CS were
296 characterized by high median concentrations of Zn followed by Cd, Pb, Fe and Cu (630, 9, 6,
297 4, 4 mg/kg, respectively). In particular, the median concentrations of Zn, Cd, Pb and Cu were
298 respectively enriched by 3, 2, 1 and 1 orders of magnitude in CS compared to US; the Ni was
299 double in CS compared to US, while Fe was comparable. Except for Pb, all the considered
300 PHEs showed higher maximum values compared to Buscaroli et al. (2017). In comparison with
301 this study Bacchetta et al. (2018) reported lower mean bioavailable concentrations of Cd, Pb
302 and Zn (3.9, 13 and 117 mg/kg, respectively) in Campo Pisanu mine district. The lower
303 bioavailable concentrations measured in US were aligned with unpolluted data by Buscaroli et
304 al. (2017).

305 The SRCCs for elements in soils from CS revealed significant positive correlations only
306 between bioavailable Cd and Ni (0.56), bioavailable Cd and Zn (0.66) and between total and
307 bioavailable Zn (0.72). Negative significant correlation existed between bioavailable Fe and
308 total Zn and bioavailable Zn (-0.69 and -0.68, respectively). The negative correlation between
309 bioavailable Zn and Pb was weak but significant (-0.52) (Table S3 of the Supplementary
310 Materials).

311

312 Table 1. Minimum, median, maximum and mean values for soils main characteristics in sampling areas. Contaminated sites
 313 (CS) (n = 22) and Uncontaminated sites (US) (n = 6). Bold values in the aqua regia results exceeded the Italian environmental
 314 legislation (Italian Government, 2006) for soils in public and residential areas.
 315

Areas		CS				US			
Statistics		Min	Median	Max	Mean	Min	Median	Max	Mean
Sand 2000-50 µm	g/kg	241	695	874	649	335	413	954	516
Silt 50-2 µm	g/kg	87	228	677	265	13	447	538	390
Clay < 2 µm	g/kg	39	83	138	86	33	52	176	93
Reaction in H ₂ O	pH	5.4	6.2	6.7	6.1	7.6	8.0	9.1	8.2
EC	dS·m ⁻¹	0.2	0.3	0.3	0.3	0.1	0.2	0.3	0.2
Total lime	g/kg	0	0	67	8	3	157	291	146
TOC	g/kg	2	22	34	20	2	26	60	26
TN	g/kg	0	2	3	2	0	2	5	2
TOC/TN		5	13	20	13	9	12	31	15
Major elements and LOI									
Si	mg/kg	245288	282938	353616	293225	54917	164808	408674	275377
Ti	mg/kg	1352	4040	5085	3638	499	1615	3559	1844
Al	mg/kg	33902	89741	115742	80668	22022	41253	69146	43051
Fe	mg/kg	22962	40509	78330	43247	5875	16208	39645	20045
Mn	mg/kg	845	1396	3809	1641	152	552	697	462
Mg	mg/kg	3973	7593	11630	7692	2749	7112	26934	11770
Ca	mg/kg	1102	4483	48660	7089	18775	129691	407810	149083
Na	mg/kg	1077	5142	12834	5481	824	4554	8531	4576
K	mg/kg	12274	30018	43922	28494	5767	13933	23878	14454
P	mg/kg	271	558	1006	554	128	572	700	469
LOI	mg/kg	30032	55600	122774	67108	19900	211908	234504	152533
Total elements (XRF)									
As	mg/kg	19	94	191	103	2	6	50	13
Co	mg/kg	11	16	43	19	1	3	13	5
Cu	mg/kg	1	102	706	170	2	10	26	12
Cr	mg/kg	26	57	99	55	22	65	144	80
Ni	mg/kg	11	39	73	40	5	40	105	49
Pb	mg/kg	138	604	4619	1240	10	18	22	17
V	mg/kg	39	64	114	66	10	60	161	79
Zn	mg/kg	5020	11455	41200	14873	8	29	93	39
Extractable elements (AR)									
As	mg/kg	17	80	333	98	10	11	36	16
Cd	mg/kg	5	67	228	67	0	0	1	0
Co	mg/kg	16	23	33	23	1	3	19	6
Cr	mg/kg	6	28	49	31	3	30	127	49
Cu	mg/kg	12	130	485	147	1	8	40	13
Ni	mg/kg	7	34	62	36	2	20	91	36
Pb	mg/kg	170	819	3442	1235	5	8	16	10
V	mg/kg	4	53	149	63	14	42	88	42
Zn	mg/kg	1227	9771	27286	9377	10	23	63	31
Bioavailable elements (DTPA)									
Cd	mg/kg	1	9	29	9	0	0	0	0
Cu	mg/kg	0	4	26	5	0	1	2	1
Fe	mg/kg	0	4	39	7	1	4	8	4
Ni	mg/kg	0	1	3	1	0	0	1	0
Pb	mg/kg	0	6	143	29	0	1	1	1
Zn	mg/kg	101	630	894	555	0	1	2	1

316
 317
 318 The Table 2 reports the DTPA/total concentration ratios expressed as a percentage and used
 319 to evaluate the elements' behaviour. The percentages were generally low in CS, especially
 320 regarding Fe, although the maximum values reached 45% and 27% for Cu and Pb,

321 respectively. The differences between CS and US were limited and significant only for Fe in
 322 US and Ni in CS. Regardless of the soil conditions in CS and US, Cu and Pb showed highest
 323 bioavailability followed by Zn, Ni and Fe although no systematic order was observed. Previous
 324 work by Buscaroli et al. (2017) in Montevecchio district (in brackets in Table 2) reported lower
 325 median values of Cu, Pb and Zn, even if maximum values were aligned with the ones from this
 326 study. The US ratios from this study weakly differ from Buscaroli et al. (2017) confirming that
 327 North Appennine district was well characterized by the collected samples.

328

329 Table 2. Minimum, median and maximum bioavailable/total element ratio in soils, expressed as percentage (%). Contaminated
 330 sites (CS) (n = 22); Uncontaminated sites (US) (n = 6). The letters (C for Contaminated, U for Uncontaminated) indicate the
 331 presence of statistically significant differences between percentages of each element in each area according to the Mann–
 332 Whitney test. No letter means absence of statistically significant differences between areas. The considered significant levels
 333 are p-value < 0.1 (*) and < 0.05 (**). Values between brackets are by Buscaroli et al. (2017).

334

Areas	Statistics	Cu	Fe	Ni	Pb	Zn
CS	Min	0.67 (1.28)	0.0003 (0.003)	0.67 (0.48)	0.09 (1.2)	1.34 (3.87)
	Median	3.96 (6.66)	0.008 (0.01)	1.41 U** (1.03)	1.53 (5.51)	3.25 (7.07)
	Max	45 (9.2)	0.116 (0.05)	6.2 (1.79)	26.6 (23.3)	10.4 (10.82)
US	Min	0.41 (1.74)	0.02 (0.01)	0.13 (0.25)	1.44 (3.8)	1.53 (0.18)
	Median	5.38 (7.75)	0.03 C* (0.03)	0.64 (0.71)	5.47 (6.03)	3.14 (1.96)
	Max	19.5 (12.6)	0.04 (0.07)	1.32 (1.52)	9.45 (12.6)	8.19 (2.63)

335

336

337 3.2 Plants

338 The total PHEs concentrations in the different parts of *H. italicum* plant are shown in Table 3
 339 grouped into CS and US, while the entire database is presented in Table S4 of the
 340 Supplementary Materials. The PHEs concentrations in plants were higher in CS than in US. In
 341 particular, the differences were statistically significant for Cd, Fe, Ni, Pb and Zn (“U” in Table
 342 3), while there were no significant differences for Cu, although median concentrations in plants
 343 were slightly higher in CS than US.

344 The median concentrations in leaf were generally higher than the other plant parts, although
 345 statistical significance (p-value < 0.01) occurred only for Cu, Fe and Zn in CS (leaves marked
 346 with “r” and “s” in Table 3). In US no significant differences in element concentrations among
 347 plant parts were found. Only Cu in leaves was statistically different from Cu stem concentration.
 348 In both CS and US, the concentrations of PHEs in the stems were lower than in the other parts
 349 of the plant, although they were significantly different only for Cd (compared to root content),
 350 Cu, Fe, Pb and Zn in CS and for Cu in US, mainly compared to the leaves (Table 3). The ratio
 351 among median element concentrations in plant parts from CS and US revealed the following
 352 enrichment ranking: Pb > Zn > Cd > Fe > Ni > Cu (Table 3).

353 The median concentrations of Pb in plants were 40 mg/kg in roots, 27 mg/kg in stems and 64
354 mg/kg in leaves with the CS/US ratios ranging from 33 to 64 for stems and leaves, respectively
355 (Table 3). The median values of Zn in plants were 576 mg/kg in roots, 391 mg/kg in stems and
356 1206 mg/kg in leaves, this last proven statistically different from the others (Table 3).
357 *Helicrysum Italicum* plants showed from 13 to 26 times more Zn in CS than in US. The median
358 concentrations of Cd in plant were quite homogeneous: 3.1 mg/kg in stems, 6.0 mg/kg in leaves
359 and 5.7 mg/kg in roots, these last two significantly higher than in stems. These concentrations
360 of Cd were from 10 to 15 times more enriched in CS than in US. Median concentration of Fe
361 in leaves (1523 mg/kg) was significantly different from roots (701 mg/kg) and stems (62 mg/kg)
362 and *H. Italicum* plants were from 3 to 5 times more Fe-enriched in CS than in US. The median
363 concentrations of Ni were similar in plants: 4.4 mg/kg in roots, 3.5 mg/kg in stems, 4.6 mg/kg
364 in leaves. Plants from CS had only 2 – 3 times more Ni than the ones from US. The median
365 concentrations of Cu in CS plants were 13 mg/kg in roots, 10 mg/kg in stems and 20 mg/kg in
366 leaves, this last significantly different from roots and stems. No differences of Cu content
367 existed between CS and US (median ratio was 1 in Table 3).

368 Few authors studied element distribution in *H. italicum* plant parts including stems. In an
369 adjacent mine area, Barbafieri et al. (2011) showed similar Cd concentrations and distribution
370 in the same plant parts with stems as the lowest accumulation site. Instead, the same authors
371 reported increasing concentrations in Pb and Zn from roots to leaves, but compared to this
372 study mean concentrations were one order of magnitude higher for Pb and slightly lower for
373 Zn. Also Brunetti et al. (2018) evaluated Cd, Cu, Ni, Pb and Zn abundance in roots, stems and
374 leaves of *H. italicum* grown in a polluted soil in Apulia Region (Italy). Their abundances of Cu
375 and Ni in plants were aligned with the concentrations of this study, while Cd, Pb and Zn were
376 one order of magnitude lower. Brunetti et al. (2018) concluded that *H. italicum* stores PHEs in
377 the roots with stems as the least concentrated part. Although not clearly evident in this study,
378 the trend to accumulate Pb in roots rather than in other aerial parts is widely demonstrated in
379 *H. italicum* (Barbafieri et al., 2011; Brunetti et al., 2018) and also other plants such as *Oryza*
380 *sativa* (Ashraf et al., 2020) and *Crambe abyssinica* (Gonçalves et al., 2020), confirming the
381 poor Pb translocation.

382 In a pyrite-mine site in Tuscany (Italy), Bini et al. (2017) showed preferential accumulation of
383 PHEs in *H. italicum* roots. Compared to this study, Fe and Pb concentrations in plants were
384 aligned, Cd and Ni were 10 times lower, Zn was three orders of magnitude lower and only Cu
385 resulted 10 times higher. For *H. italicum* subsp. *tyrrhenicum* (*H. tyrrhenicum*), in the adjacent
386 mine area of Campo Pisanu, Bacchetta et al. (2018) detected higher concentrations in roots
387 and leaves for all considered elements (Cd, Pb and Zn) and, especially for Pb, the differences
388 with this study were quite important.

389 Pot trials at different soil contamination of Cu and Pb executed on *Helichrysum splendidum*
 390 Less revealed a reduction of chlorophyll content (phytotoxicity sign) only when the Cu leaves
 391 concentration was about 290 mg/kg, while for the Pb a constant chlorophyll reduction was
 392 evident starting from 90 mg/kg in leaves (Banda et al., 2021). *Helichrysum italicum* never
 393 reached high concentrations of Cu in this work, while maximum concentrations of Pb exceeded
 394 90 mg/kg in CS plants. Since it was not within the aims of the work, no surveys regarding the
 395 health status of the plants were conducted, therefore the presence of *H. italicum* plants with
 396 phytotoxic symptoms in CS could not be excluded.

397

398 Table 3. Minimum, median, maximum and mean element concentration in plant parts for Contaminated Sites (CS, n = 22) and
 399 Uncontaminated Sites (US, n = 6). Concentrations are in mg/kg. The letters R, S and L represent roots, stems and leaves,
 400 respectively. According to the Mann–Whitney test, the Statistical Significant Difference (SSD) between the analysed plant
 401 tissues within each area was indicated by the letters r and s. According to the Mann–Whitney test, the letter U indicates
 402 statistically significant difference of the plant tissue from CS with US. No letter means absence of statistically significant
 403 differences. The considered significant levels are p-value < 0.1 (*), < 0.05 (**) and < 0.01 (***).
 404

Elements	Plant parts	CS					SSD	US				SSD	Median CS/US ratio
		Min	Median	Max	Mean	Min		Median	Max	Mean			
Cd	L	0.5	6.0	90.0	12.4	U**	0.1	0.4	1.3	0.5		15	
	S	0.6	3.1	22.3	5.8	U**	0.1	0.3	1.1	0.4		10	
	R	1.4	5.7	50.9	11.1	s**U**	0.1	0.5	1.7	0.6		11	
Cu	L	7.8	20.2	98.4	26.0	r*s***	9.4	13.9	29.8	16.8	s*	1	
	S	3.2	10.2	32.3	12.6		6.3	9.3	11.9	9.1		1	
	R	6.5	12.9	103.2	22.4		6.3	9.8	19.0	10.9		1	
Fe	L	359	1523	5991	1994	r**s***U**	174	290	677	343		5	
	S	267	662	1905	748	U**	112	206	427	228		3	
	R	142	701	7140	1245	U**	150	252	581	319		3	
Ni	L	1.3	4.6	11.3	4.8	U**	0.6	1.7	4.1	1.9		3	
	S	2.0	3.5	5.6	3.8	U**	0.4	1.3	2.1	1.2		3	
	R	0.6	4.4	11.0	4.7	U*	0.8	1.8	2.4	1.7		2	
Pb	L	5.9	63.7	288.4	74.5	s**U**	0.6	1.0	13.0	3.6		64	
	S	7.8	26.5	181.4	37.4	U**	0.5	0.8	5.9	1.8		33	
	R	5.2	39.5	384.4	76.5	U**	0.4	0.9	4.4	1.6		44	
Zn	L	177	1206	9837	1764	r*s**U**	28	47	112	61		26	
	S	132	391	1959	630	U**	19	30	51	32		13	
	R	220	576	3337	936	U**	26	28	46	32		21	

405

406

407 3.2.1 Interaction mechanisms of PHEs in *H. italicum* plants

408 The interaction mechanisms between PHEs could reveal synergistic or antagonistic effects
 409 able to improve or reduce element uptake and translocation in plant species. Only for plants
 410 collected in CS (n = 22), the SRCs and their significance levels were calculated for each
 411 element (Cd, Cu, Fe, Ni, Pb and Zn) and for different plant parts and results are presented in
 412 Table 4.

413 Results of this study indicated that Zn in *H. italicum* was positively and significantly correlated
414 with Cd in all considered plant parts (around 0.90 for R, S and L). Moreover, Zn and Cd were
415 also themselves correlated with values of 0.92 and 0.9 between leaves and stems. The Zn and
416 Cd, together with Pb were the most enriched elements in *H. italicum* compared to US (Table
417 3). Nevertheless, Zn and Cd were also positively and significantly correlated with all the other
418 elements in leaves (from 0.55 to 0.7), except for Pb that showed correlations only with Cd in
419 leaves (0.67). The relation between Zn and Cd is known in literature and depends by the
420 element similarity (Fernández et al., 2017). Indeed, Cd plant uptake is hindered by high soil Zn
421 concentrations (Choudhary et al., 1995; Oliver et al., 1994) because they share the same
422 transportation protein in plants and Zn is selectively preferred (Hart et al., 2002). Kutrowska et
423 al. (2017) documented the synergistic effect between Cd and Zn in *Brassica juncea* where Zn
424 increased the accumulation of Cd in leaves. The same authors reported that Pb increased the
425 Cd in stems, as identified also in the present study (Table 4).

426 In *H. italicum* grown at CS Cu was positively correlated with itself, especially between stems
427 and leaves (0.74), and clearly positively correlated also with Zn for almost all plant parts with
428 a peak of 0.86 in leaves. As Fe, also Cu and Zn are micronutrients for plants and serve in
429 physiological processes; so, their synergism was expected for US, but it was found also at
430 elevated Zn concentration in CS (Table 3). In other species like *Brassica juncea* antagonism
431 between Zn and Cu was demonstrated and indicated as happening not in the roots, but later
432 during xylem loading/unloading (Kutrowska et al., 2017). A previous study (An et al., 2004)
433 revealed that Cu and Cd act antagonistically resulting in decreased accumulation of both
434 metals in *Cucumis sativus*. In the present study Cu and Cd were weakly correlated with
435 coefficients around 0.6 between Cu in leaves and Cd in roots, stems and leaves (Table 4).

436 The Fe was significantly correlated between stems and leaves (0.8), in leaves with Pb (0.71)
437 and in leaves with Cd (0.73 and 0.72 for stems and leaves, respectively). Moreover, Fe and Ni
438 resulted significantly correlated in all the *H. italicum* plant parts, with coefficients ranging
439 around 0.70 and peaks for roots (0.74) and leaves (0.71) (Table 4). Khalid and Tinsley (1980),
440 in *Lolium perenne*, reported a common increase of Ni and Fe concentrations in shoots with
441 increasing rates of Ni. Same synergistic effect of Ni on Fe was detected in maize with highest
442 evidence in roots and leaves by Torres et al. (2016).

443 Absence of correlation existed among Pb and Cu as well as Pb and Zn in the plant parts of *H.*
444 *italicum* grown at CS. An antagonistic effect of Pb on the Cu accumulation was documented
445 also in *Brassica juncea* and related to a competition between metals at the plant uptake site
446 (Kutrowska et al., 2017). Also for Israr et al. (2011), Pb showed antagonistic effect on the
447 accumulation of Cu, Ni and Zn in *S. drummondii* species, probably due to the competition
448 between metals at the plant uptake sites. Yet, Wong et al. (1986) reported a reduced uptake

449 of Cu in the presence of Pb for *Brassica chinensis*. The inhibition of essential nutrient transfer
450 (such as Cu and Zn) in plant biomass due to Pb elevated concentration has been also proposed
451 by Yoon et al. (2006) for numerous plants grown in Florida contaminated site. In addition to all
452 the side effects of Pb in plants, Pourrut et al. (2011) reported impaired uptake of essential
453 elements, such as Mg and Fe. On the contrary, An et al. (2004) showed positive correlation of
454 Zn and Pb in *Cucumis sativus* suggesting a synergistic effect.

455

456

457 Table 4. Main Spearman Rank Correlation Coefficients (SRCCs) and their significance levels, calculated for each element
 458 (Cd, Cu, Fe, Ni Pb and Zn) between the different plant parts (R=roots, S=stems and L=leaves) and, for each of them, between
 459 the different elements. The SRCCs were calculated considering only the Contaminated Sites (CS) (n=22). The considered
 460 significant levels are p-value < 0.05 (**) and < 0.01 (***).
 461

		R	S	L	R	S	L	R	S	L	R	S	L	R	S	L	R	S	L			
		Cd			Cu			Fe			Ni			Pb			Zn					
R	Cd	1																				
S		0.76 ***	1																			
L		0.81 ***	0.92 ***	1																		
R	Cu	0.55 **	0.42	0.44 **	1																	
S		0.22	0.30	0.31	0.60 ***	1																
L		0.66 ***	0.56 ***	0.66 ***	0.69 ***	0.74 ***	1															
R	Fe	0.37	0.19	0.20	0.57 ***	-	0.01	0.24	1													
S		0.36	0.64 ***	0.53 **	0.47 **	0.30	0.31	0.49 **	1													
L		0.55 **	0.73 ***	0.72 ***	0.47 **	0.30	0.49 **	0.33	0.80 ***	1												
R	Ni	0.27	0.08	0.17	0.55 **	-	0.01	0.24	0.74 ***	0.32	0.37	1										
S		0.39	0.46 **	0.51 **	0.65 ***	0.20	0.37	0.55 **	0.69 ***	0.68 ***	0.62 ***	1										
L		0.64 ***	0.66 ***	0.75 ***	0.48	0.01	0.41	0.31	0.52 **	0.71 ***	0.54 **	0.73 ***	1									
R	Pb	0.12	-	-	0.39	-	-	0.03	0.67 ***	0.17	0.08	0.56 ***	0.36	0.09	1							
S		0.40	0.40	0.48 **	0.15	0.09	0.23	0.42	0.47 **	0.46 **	0.42	0.35	0.41	0.53	1							
L		0.46 **	0.51 **	0.67 ***	0.20	0.08	0.29	0.33	0.53 **	0.71 ***	0.41	0.49 **	0.60 ***	0.36	0.86 ***	1						
R	Zn	0.90 ***	0.68 ***	0.70 ***	0.77 ***	0.33	0.70 ***	0.50 **	0.39	0.56 ***	0.46 **	0.52 **	0.64 ***	0.26	0.34	0.37	1					
S		0.78 ***	0.93 ***	0.87 ***	0.54 **	0.45 **	0.73 ***	0.24	0.61 ***	0.66 ***	0.12	0.43	0.60 ***	-	0.15	0.35	0.40	0.74 ***	1			
L		0.87 ***	0.82 ***	0.89 ***	0.60 ***	0.45 **	0.86 ***	0.27	0.38	0.59 ***	0.19	0.42	0.60 ***	-	0.09	0.28	0.42	0.82 ***	0.90 ***	1		

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463

464

3.3 Soil-plant relationship

465 The relationship between the soil bioavailable pool and the concentrations in plant parts of the
 466 investigated elements are represented by box plots in Figure 2. In terms of soil bioavailable
 467 pool, Fe and Ni showed comparable concentrations in CS and US even if Ni in US had a wider
 468 range of variation. At these low concentrations the *H. italicum* was able to absorb and
 469 concentrate the elements in the plant (2 and 1 order of magnitude for Fe and Ni, respectively).
 470 Although the soil bioavailable fractions were similar the plant concentrations between CS and
 471 US were statistically different (Table 3) with lower levels in the latter.

472 The Cu and Pb were enriched in CS compared to US, but due to the wider range of variations
 473 (Cu in US and Pb in CS) a clear separation was not evident. Regarding these elements the

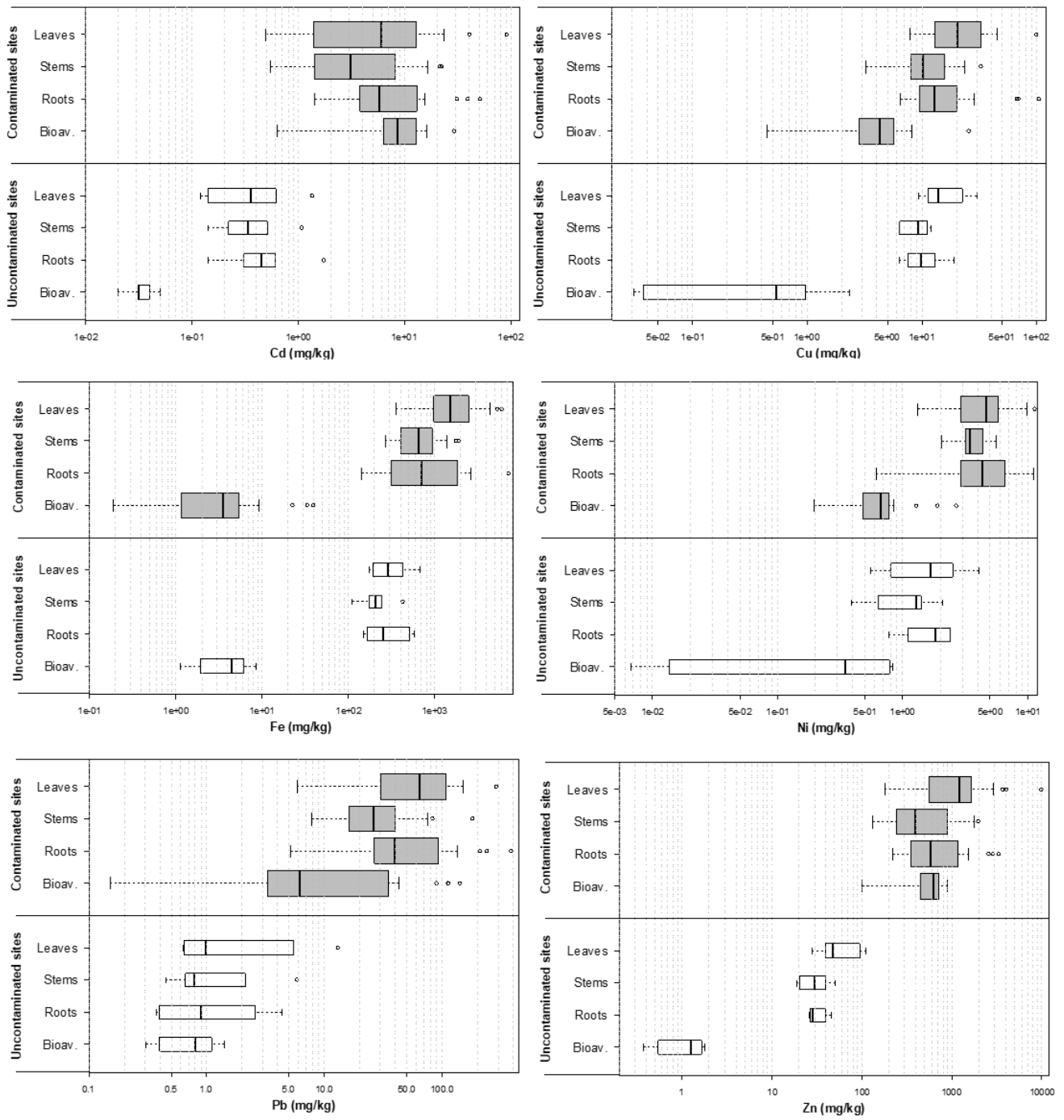
474 plant was able to absorb and concentrate them compared to the soil bioavailable pool (1 order
475 of magnitude for Pb). The Cu concentrations in the plant were not statistically different between
476 CS and US (Table 3) indicating that CS could not be considered polluted by Cu. In fact,
477 Buscaroli et al. (2017) in Cu-contaminated Libiola mine and Brunetti et al. (2018) in Apulia soil
478 reported bioavailable Cu concentration greater than this study by 1 and 2 orders of magnitude,
479 respectively. The Pb in CS was absorbed and concentrated by *H. italicum*, while in US it was
480 absorbed but not concentrated compared to soil bioavailable fractions (Fig. 2).

481 Soil bioavailable concentration of Cd and Zn were more abundant in CS than US of 2 and 3
482 times, respectively. In CS bioavailable Cd and Zn were elevated (more than Bacchetta et al.
483 (2018) and Brunetti et al. (2018)) and *H. italicum* plant absorbed them, but weakly concentrated
484 Cd and Zn in the plant tissues. Instead, in US *H. italicum* was able to absorb and concentrate
485 Cd and Zn (Fig. 2), even if the plant concentrations remained significantly lower than CS (Table
486 3). Despite this behaviour within plants, *H. italicum* growing in CS accumulated two orders of
487 magnitude more Pb and one order of magnitude more Zn and Cd compared to US (Fig. 2).

488 Bacchetta et al. (2018) for *H. tyrrhenicum* in the adjacent mine area of Campo Pisanu, detected
489 higher concentrations in plant parts for all considered elements (Cd, Pb, and Zn). Yet,
490 bioavailable pools for Cd and Zn were a quarter the levels of the present study, while the
491 bioavailable pool of Pb was three times higher. Brunetti et al. (2018), starting from bioavailable
492 pools like Bacchetta et al. (2018), presented notably lower plant concentrations. This behaviour
493 could be related to the carbonate-soils studied by Brunetti et al. (2018) that contain high
494 exchangeable Ca. This latter could compete with heavy metals limiting their uptake. The
495 antagonistic effect of Ca on Cd, Cu, Fe, Ni, Pb and Zn uptake was found by Kabata-Pendias
496 (2010) and observed also for *D. viscosa* by Buscaroli et al. (2017).

497 Possible interaction mechanisms between elements in soil and plant from CS were
498 investigated through SRCCs and results are presented in Table S3 of the Supplementary
499 Materials. In general, the correlations between soil and plant concentrations were scarce.
500 The only positive significant correlation existed between total Pb in soil and Pb in roots (0.75).
501 There were weak significant negative correlations between total Ni and Cu in stems (-0.56)
502 and leaves (-0.59), between bioavailable Pb and Cd in stems (-0.53) and leaves (-0.52), and
503 Zn in stems (-0.54) (Table S3).

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Figure 2. Bioavailable Cd, Cu, Fe, Ni, Pb and Zn in soils and their distribution in *H. italicum* plants of CS and US.

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510 **3.4 Root/bioavailable soil element concentration ratio**

511 With the aim of studying the tolerance to PHEs of *H. italicum* and provide evidence of
512 phytostabilization or phytoextraction capacity, the root/bioavailable soil concentration ratio
513 (root/soil) was calculated for each element and the results are shown in Table 5. The mean
514 values were presented only for literature comparison.

515 For all the elements in CS and US the medians for root/soil ratios were > 1 suggesting that *H.*
516 *italicum* was able to absorb the bioavailable elements in roots. According to root/soil median
517 values, in the CS the elements were absorbed following the ranking Fe >> Ni > Pb > Cu,
518 whereas the ratios for Zn and Cd were close to 1. In the US the ranking resulted Fe > Zn > Cu
519 > Cd > Ni and Pb only slightly above 1 (Table 5).

520 Out of the three most abundant bioavailable elements in CS (Zn, Pb and Cd, Table 1), Cd
521 (1.21) and Zn (1.39) had the lowest ratios indicating a limited root uptake by *H. italicum* when
522 soil available pool is abundant. Pb was more absorbed in CS (4.17) than in US (1.21). Same
523 behaviour for Zn, Pb and Cd was detected in *D. viscosa* for the same mine area (Buscaroli et
524 al., 2017).

525 The different ratio of Cu between CS and US depended from its wide variability in US, while Ni
526 showed similar soil/root ratios in both conditions. The Fe is a fundamental micronutrient for
527 plants strongly related to chlorophyll content and plant growth (Terry, 1980). Despite its low
528 availability in soil (Tables 1 and 2), *H. italicum* showed an elevated root uptake capacity for Fe,
529 especially in CS (Fig. 2).

530 Many other authors calculated the root/soil ratios of *H. italicum* (Barbafieri et al., 2011; Bini et
531 al., 2017; Cao et al., 2004; Leita et al., 1989) but only Bacchetta et al. (2018) and Brunetti et
532 al. (2018) applied the DTPA extraction, allowing a direct comparison with ratios from this study.
533 In particular, Bacchetta et al. (2018) found ratios of 21, 29 and 9.6 for Zn, Pb and Cd,
534 respectively, for the *H. tyrrhenicum* in a Sardinian mine site. These ratios were higher than this
535 study, but with lower bioavailable concentrations and significantly higher root concentrations.
536 Recalculated ratios by Brunetti et al. (2018) showed for Cu, Ni, Pb and Zn slightly lower values,
537 under reduced bioavailable soil pool compared to CS of this study. Instead, Cd showed root/soil
538 ratio one order of magnitude higher and bioavailable concentrations one order of magnitude
539 lower compared to this study in CS, but comparable with values in US (Tables 1 and 5). This
540 confirms that Cd is strongly incorporated in roots at low soil bioavailable concentrations
541 (Brunetti et al., 2018), while its absorption is limited when soil concentrations increase (Fig. 2
542 and Bacchetta et al., 2018). This represents an excluding mechanism for Cd already
543 documented in other plant species such as *Thlaspi arvense* (Martin et al., 2012) and
544 *Arabidopsis thaliana* (Zhu et al., 2012), but not yet in *Helicrysum* spp.

545 Table 5. Minimum, median, maximum and mean root/soil values of selected elements in *H. italicum*. CS n=22 and US n=6.

Area	Statistics	Cd	Cu	Fe	Ni	Pb	Zn
CS	Min.	0.14	1.45	3.60	0.81	0.36	0.35
	Median	1.21	3.03	299	6.56	4.17	1.39
	Max.	7.95	28.9	6541	29.0	52.0	5.92
US	Mean	1.76	5.47	988	8.84	11.57	1.92
	Min.	4.43	3.22	23.0	1.33	0.40	15.7
	Median	12.1	19.1	91.2	4.67	1.21	26.3
	Max.	43.1	417	135	272	8.56	74.4
	Mean	18.47	107.73	89.06	70.56	2.54	38.99
	US/CS Median ratio	10	6.3	0.3	0.71	0.29	18.9

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3.5 Leaf/root element concentration ratio

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The capacity of *H. italicum* to transfer PHEs from roots to leaves has been quantified for each element through the leaf/root concentration ratio (leaf/root ratio) and the results are shown in Table 6. The leaf/root ratio was extensively applied for the evaluation of the phytoextraction capacity of plants growing in mine soils (Buscaroli et al., 2017; Martínez-Sánchez et al., 2012; Wang et al., 2019; Yoon et al., 2006) or in contaminated agricultural soils (Dinu et al., 2020; Nadimi-Goki et al., 2014). As stated for root/soil ratio, also leaf/root ratio is widely used by researchers, but under different names such as Translocation factor, Transfer factor or Transportation index (Buscaroli et al., 2017).

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The *H. italicum* in CS always showed leaf/root median ratios > 1 , except for Cd and Ni that were 0.75 and 1, respectively (Table 6). Iron reached a value of 2.36 followed by Zn (1.54) then Pb (1.37) and Cu (1.34). The leaf/root median values in US were comparable with CS except for Fe that was half. The ratios of Zn, Ni, and Cd were almost identical, while slightly higher values were measured for Pb and Cu in US compared to CS (Table 6). However, the maximum ratios were from 2 to 10 times higher in CS than US (Table 6) for all considered elements, indicating an increased inclination of the plants to transfer elements in leaves when growing on contaminated soils.

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As all the *H. spp.*, also *H. italicum* is a terpene-rich species, rarely appreciated as food by wild or domestic herbivorous, reducing the possibilities of PHEs entering the food chain (Rogosic et al., 2006). Moreover, metal accumulation in aerial parts is an evolutionary adaptation that confers to plants also protection against herbivores or pathogens (Galeas et al., 2008).

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In presence of a tolerant plant (root/soil ratio > 1), the higher the leaf/root ratio (e.g., > 2), the greater the capacity to transfer elements to aerial plant parts and the more adapt is the species for phytoextraction strategy (Buscaroli et al., 2017; Yoon et al., 2006). The opposite indicates the suitability of the species for phytostabilization (Rizzi et al., 2004; Yoon et al., 2006). Results from this study indicated *H. italicum* as a tolerant species weakly able to concentrate Fe, Zn,

576 Pb and Cu in the aerial parts and unable to transfer Cd and Ni to leaves. The leaf/root ratio
 577 recalculated by Barbaferi et al. (2011) in *H. italicum* plants revealed almost identical values for
 578 Zn and Cd and double values for Pb compared to the present study. Identical values for the
 579 ratio were reported also by Bacchetta et al. (2018) in Campo Pisano mine site, but for *H.*
 580 *tyrrhenicum* and for the entire epigeal organs. Half values of the ratios for Cd, Cu, Pb and Zn
 581 and about one tenth for Fe were obtained by Bini et al. (2017) in a Tuscany mine district, while
 582 Brunetti et al. (2018) reported lower ratios for all elements (Cu, Ni, Pb and Zn). In the recent
 583 pot trials by Banda et al. (2021) leaves/root ratios of Cu and Pb were weakly above 1 in
 584 *Helichrysum splendidum* Less.

585 Only for comparison purposes, Boechat et al. (2016) reported leaf/root ratios for several
 586 Brazilian species well above 2. In particular, *Baccharis trimera* (Less) DC (5.48), *Cyperus*
 587 *eragrostis* Lam (3.54), *Eryngium horridum* Malme (2.91) and *Dicranopteris nervosa* (Kaulf.)
 588 (2.61) for Pb and *Senecio brasiliensis* (Spreng.) Less (2.93) for Cd.

589

590 Table 6. Minimum, median, maximum, and mean leaf/root total element concentration values of selected elements in *H.*
 591 *italicum*. CS n = 22 and US n = 6.

Area	Statistics	Cd	Cu	Fe	Ni	Pb	Zn
CS	Min.	0.22	0.41	0.17	0.39	0.12	0.80
	Median	0.75	1.34	2.36	1.02	1.37	1.54
	Max.	6.91	4.96	19.3	6.59	14.3	13.1
	Mean	1.20	1.61	3.39	1.46	2.46	2.27
US	Min.	0.40	0.73	0.38	0.30	0.71	1.04
	Median	0.74	1.78	1.08	1.04	1.87	1.51
	Max.	1.35	2.98	2.88	2.11	3.00	3.67
	Mean	0.82	1.70	1.33	1.12	1.83	1.90

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594 3.6 Literature comparison and final remarks

595 The interest for the species *H. italicum* (subsp. *italicum* or *tyrrhenicum*) as a possible
 596 phytoremediation plant, especially in abandoned mine areas, goes long back in time. From
 597 Leita et al. (1989) up to Bacchetta et al. (2018) and Brunetti et al. (2018) many authors have
 598 studied the capacity of *H. italicum*, to uptake PHEs from contaminated soils. With the aim
 599 of summarizing the available literature on these species and outline their phytoremediation
 600 capacities, a literature research browsing the keyword “*Helichrysum italicum* and remediation”
 601 in the Web of Science (last time checked 08/02/2022) was performed. Only 3 publications
 602 appeared Bini et al. (2017), Brunetti et al. (2018) and Boi et al. (2020). The work of Boi et al.
 603 (2020) was dedicated to the seed germination and for this reason excluded. This literature
 604 research demonstrated that the publications already cited in this study represent the most
 605 updated articles dealing with the application of *H. italicum* for phytoremediation purpose. All

606 the available articles considering *H. italicum* for phytoremediation (N=6) (180 total citations and
607 maximum number 75 for Barbafieri et al., 2011) have been reviewed and their main
608 characteristics summarized in Table 7. Particular attention has been paid to the analytical
609 methods applied on soil samples. The DTPA extraction method for the determination of the
610 bioavailable soil fraction has been executed only by Bacchetta et al. (2018) and Brunetti et al.
611 (2018). This last was a greenhouse study conducted on contaminated agricultural soil in Apulia
612 Region and not on mine tailings as all the other considered studies (Table 7). The remaining
613 studies adopted more aggressive soil extraction techniques and, although presented, they
614 cannot be considered in the discussion.

615

616

617 Table 7. Locations of the studies, quantity of analysed samples, adopted methodologies for soil and plant analysis, considered
 618 elements and *Helicrysum Spp.* from the literature review for the *H. italicum Spp.*. "n.a." means not available information.
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Literature	N° of CS samples	Methodology for PHEs determination in contaminated soils	N° of plant	Methodology for PHEs determination in plants	Considered elements	Studied <i>Helicrysum</i> subspp.	Location in Italy
Current study	22	DTPA extraction, Lindsay and Norvell (1978)	22	0.25 g plant +2 ml H ₂ O ₂ + 6 ml HNO ₃	Cd, Cu, Fe, Ni, Pb, Zn	<i>H. italicum</i>	Montevecchio mine district (Sardinia)
Barbafieri et al., 2011	3	SEP ⁽¹⁾ with H ₂ O, KNO ₃ , EDTA	9	HNO ₃ /HClO ₄ in 2.5/1 ratio	Cd, Pb, Zn	<i>H. italicum</i>	Ingurtosu mine district (Sardinia)
Bacchetta et al., 2018	5	DTPA extraction, Lindsay and Norvell (1978)	5	0.5 g plant + 9 ml HNO ₃ + 0.5 ml HF	Cd, Pb, Zn	<i>H. tyrrhenicum</i>	Campo Pisano mine district (Sardinia)
Brunetti et al., 2018	10	DTPA extraction, Lindsay and Norvell (1978)	10	HNO ₃ :H ₂ O ₂ :HCl mixture (5:1:1 v/v)	Cd, Co, Cr, Cu, Ni, Pb, Zn	<i>H. italicum</i>	Agricultural area Alta Murgia (Apulia)
Bini et al., 2017	n.a.	0.2 g of soil + 5 ml of aqua regia (37% HCl+65% HNO ₃ , 1:3) +1 ml of 48% HF + 1 ml of cold supersaturated H ₃ BO ₃	n.a.	0.5 g of plants + 5 ml 65% HNO ₃ + 3 ml 30% H ₂ O ₂ in open vessels on the hot plat	Cd, Co, Cr, Cu, Fe, Mn, Ni, Pb, Zn	<i>H. italicum</i>	Nocciola mine district (Tuscany)
Cao et al., 2004	n.a.	SEP ⁽¹⁾ with H ₂ O, KNO ₃ , EDTA	n.a.	Aqua Regia	Pb, Zn	<i>H. italicum</i>	Montevecchio mine district (Sardinia)
Leita et al., 1989	3	10 g soil + 50 ml of 0.05 M EDTA.	3	1 g plants digested in concentrated HNO ₃ - HCl 3:1 at 150 °C	Cd, Cu, Pb, Zn	<i>H. italicum</i>	Wide Iglesias mine district (Sardinia)

(1): Sequential Extraction Procedure

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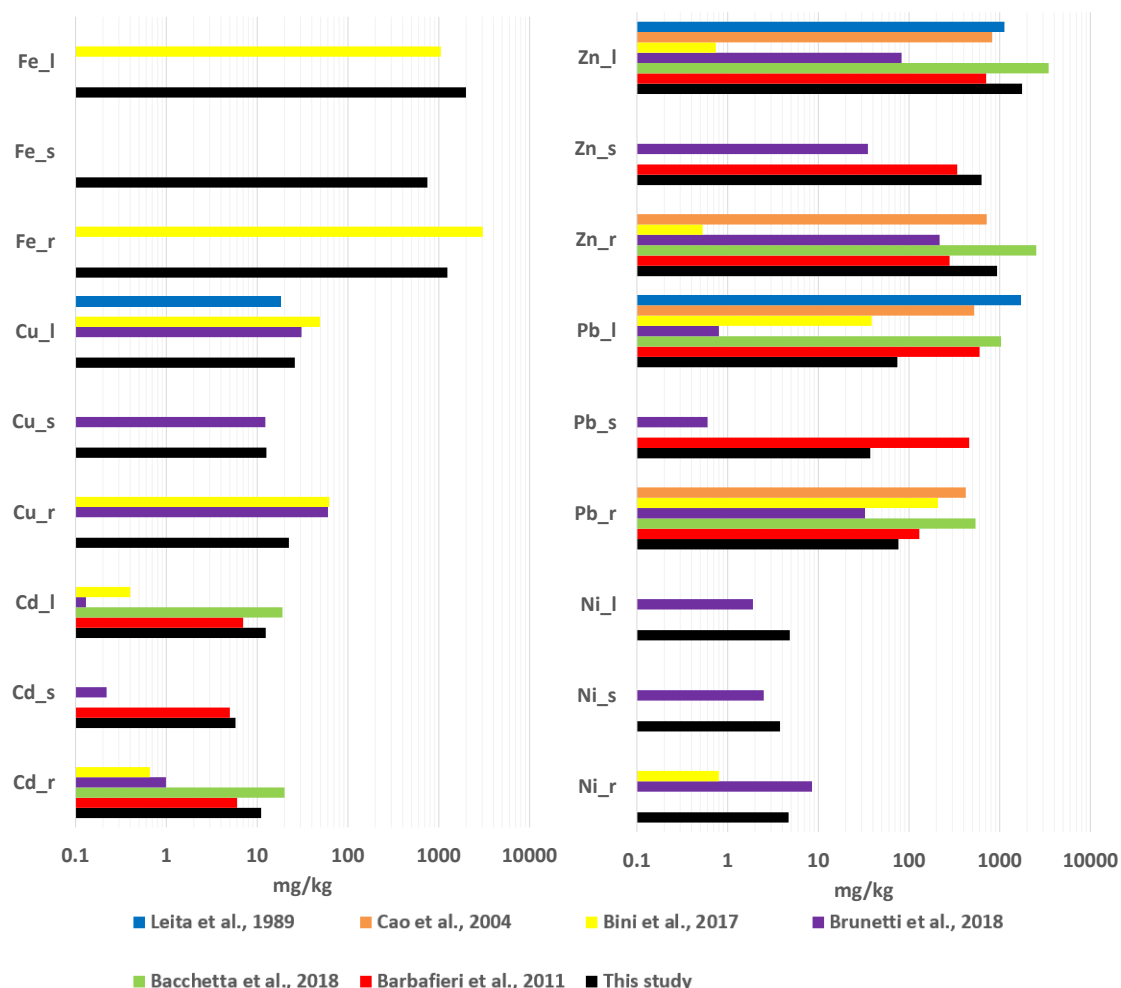
621 Concerning concentrations in plants, the literature review showed that only Zn and Pb were
 622 considered in all 7 studies (including this), Cd was measured in 5, while Cu, Ni and Fe in 4, 3
 623 and 2 studies, respectively (Fig. 3). Only this study, Barbafieri et al. (2011) and Brunetti et al.
 624 (2018) measured PHEs in stems and leaves separately. All the other studies were limited to
 625 roots and leaves and Leita et al. (1989) considered only leaves.

626 Mean element concentrations identified Zn, Pb and Fe as most abundant in *H. italicum* plant
 627 with concentrations around 10³ mg/kg, 5*10² mg/kg and 10³ mg/kg, respectively (Fig. 3).
 628 Instead, Cu, Cd and Ni, were poorly absorbed by the plant and their concentrations ranged
 629 around 10 mg/kg for Cu and below 10 mg/kg for Cd and Ni. The stems were the most
 630 impoverished parts for all considered elements. Based on the thresholds provided by Van der
 631 Ent et al. (2013), *H. Italicum* cannot be considered an hyperaccumulator species.

632 Considering only studies from Sardinia region that shared the same soil element abundance
 633 ranking Zn > Pb > Cd, both *H. italicum* or *H. tyrrhenicum* (our study, Bacchetta et al., 2018;
 634 Barbafieri et al., 2011; Cao et al., 2004; Leita et al., 1989) showed similar abundances for Cd,
 635 Pb and Zn, with *H. tyrrhenicum* reporting higher concentrations. The experiment of Brunetti et
 636 al. (2018) was conducted on compost-contaminated clay-loam soils at basic pH with the

637 following contaminant soil ranking: Zn > Cu > Pb > Ni > Cd, while Bini et al. (2017), in Tuscany,
 638 had different geological settings and the contaminant rank in soil was Fe > Mn > Pb > Zn > Cu
 639 > Ni > Cd. These different experimental conditions justified the evident differences in terms of
 640 element abundances in plant parts of Cd, Zn and Pb, as well as the alignment of Fe and Cu
 641 (Fig. 3).

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644 Figure 3. Mean concentration of the analysed PHEs in *H. Spp.* plant parts (r= roots; s = steams; l = leaves) from this study
 645 and from the works presented in Table 7.

646

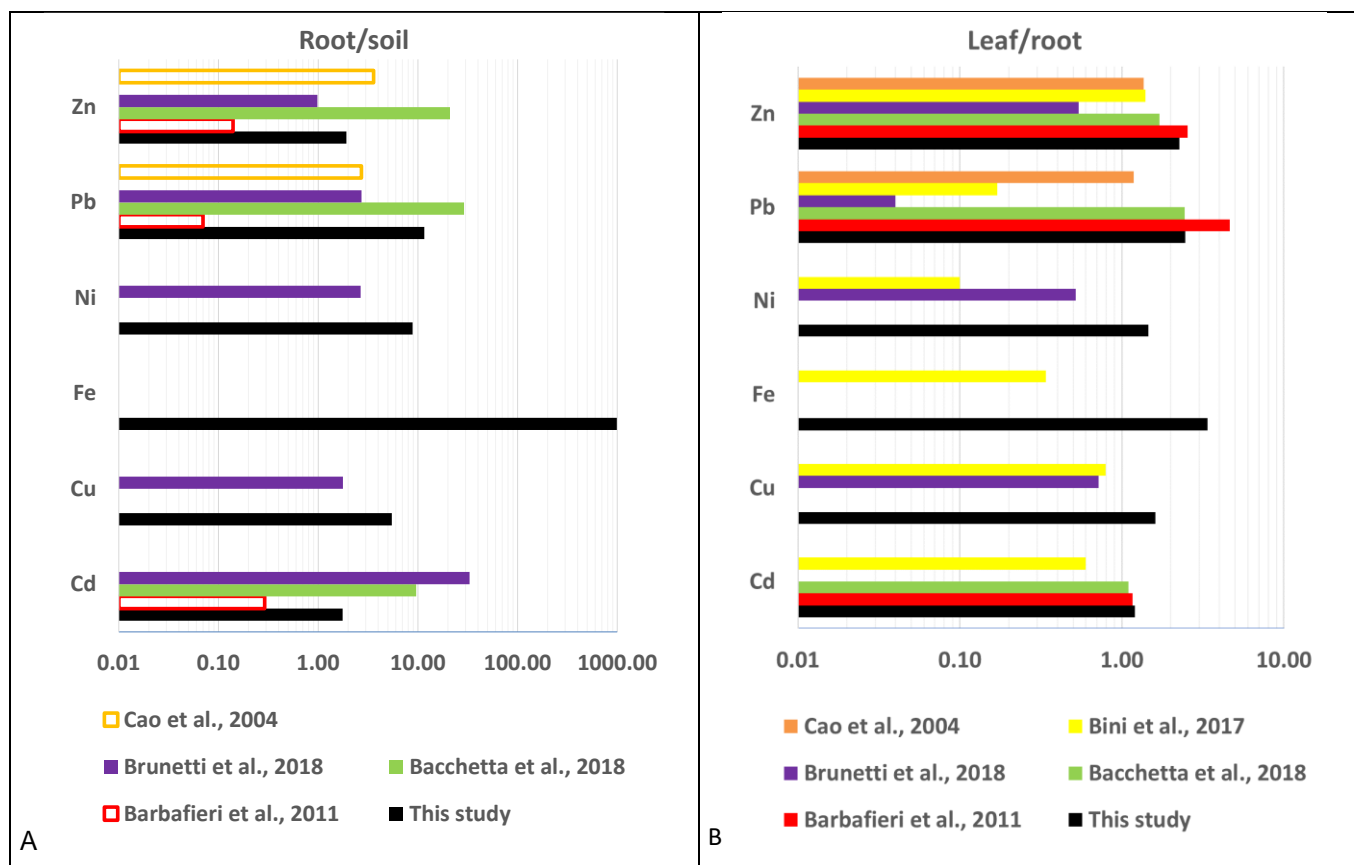
647 As stated above, the root/soil ratio can be compared only against works that adopted the same
 648 DTPA extraction on soil samples, even if ratios from Cao et al. (2004) and Barbaferi et al.
 649 (2011) are also reported (empty bars) in Figure 4A. The uptake capacity was evident for *H.*
 650 *italicum* because all values were > 1 as reported here and also by Brunetti et al. (2018) and
 651 Bacchetta et al. (2018). The ratio values rarely exceed 10 (except for Fe, but it has a biological
 652 function) and when it happens, it is for *H. tyrrhenicum* subsp. for Pb and Zn (Bacchetta et al.,
 653 2018) or for Cd at low bioavailability in soils by Brunetti et al. (2018).

654 *Helicrysum italicum* can be defined as a tolerant species for Cd, Cu, Fe, Ni, Pb and Zn, but it
655 is not suitable for phytoextraction since it shows little capacity to accumulate elements in
656 leaves. Indeed, Cd was equally distributed in the plant and leaf/root ratio was 1 both in the
657 present study and Barbafieri et al. (2011) and Bacchetta et al. (2018), while it was 0.6 in
658 Tuscany mine of Bini et al. (2017) (Fig. 4B). Although the studies conducted in the Sardinia
659 region showed an accumulation of Zn and Pb in leaves with mean leaf/root ratios around 2,
660 and even higher in Barbafieri et al. (2011) (Fig. 4B), the high variability of the leaf/root ratios in
661 this study does not allow a clear and univocal indication on translocation capacity of *H. italicum*
662 (Table 6).

663 In this study, Fe resulted the most absorbed element (Fig. 4A) and contemporarily also the
664 most translocated from roots to leaves (2.36 leaf/root median ratio in Table 6) because of its
665 biological function in photosynthetic process. Despite this, the Fe leaf/root ratio by Bini et al.
666 (2017) was widely < 1 indicating a root accumulation (Fig. 4B). Bioavailable Fe in CS and US
667 were aligned (Table 1) and the higher absorption and translocation in CS was probably related
668 to the lower pH and lime content in respect to US, as reported by Buscaroli et al. (2017). In the
669 mine site investigated by Bini et al. (2017), total Fe was the most abundant element and plants
670 preferred to store it in roots. Similar behaviour has been reported in rice species that are able
671 to oxidise Fe at the root surface, leading to the formation of iron plaques (Green and
672 Etherington, 1977) or accumulating Fe as ferric hydroxides (goethite and lepidocrocite) in roots
673 (Bacha and Hossner, 1977). In both cases Fe precipitation in roots could later influence the
674 uptake of other elements (Armstrong and Armstrong, 1988).

675

676



678 Figure 4. Mean root/bioavailable soil element concentration ratio (A) and mean leaf/root element concentration ratio (B) in *H.*
 679 *Spp.*, calculated from this study and from the other papers presented in Table 7.

680

681 Plants suitable for phytoextraction should possess multiple abilities: first of all absorb (root/soil)
 682 and translocate (leaf/root) to aerial parts heavy metals then rusticity, fast growth, high biomass
 683 yield and easy harvesting (Jabeen et al., 2009).

684 Based on the literature review, the content of all investigated elements, in plant tissues makes
 685 *H. italicum* a tolerant species (especially in respect of Pb and Zn), but it does not reach the
 686 concentrations to be defined an hyperaccumulator plant (Baker et al., 2000): 100 mg/kg of Cd
 687 (of the leaf dry weight), 1000 mg/kg for Ni, Cu and Pb and 10000 mg/kg for Zn. The root/soil
 688 bioavailable ratio > 1 for all elements suggests the use of *H. italicum* for phytostabilization in
 689 mine areas as a pioneering strategy. Although the median leaf/root ratio was > 1 for Zn and Pb
 690 in plants grown in Sardinia mine districts, contrasting mean ratios were achieved for *H. italicum*
 691 in different contaminated sites not allowing a clear evidence of its phytoextraction ability.

692 Moreover, Boi et al. (2020) argued that few kilograms per hectare (6 - 11 kg/ha) can be
 693 recovered by *H. tyrrhenicum* (the most performant subspp. as shown in Fig. 3 and 4) and given
 694 the actual price of Zn, it does not allow economic sustainability.

695 Since *H. italicum* i) is a spontaneous and perennial species, tolerant to PHEs; ii) guarantees
 696 the canopy cover all throughout the year, preventing wind dispersion and water erosion; iii)

697 influences the soil retention capacity and can itself rehabilitate the vegetation cover,
698 reactivating pedological processes; iv) can be propagated sowing directly seeds on
699 contaminated soils allowing cheaper propagation; v) permits the stabilization of mine tailing
700 also from land management point of view, it can be indicated for phytostabilization in
701 abandoned mine districts, reducing the impact of PHEs on the mine sites and surrounding
702 environments (Barbafieri et al., 2011; Bacchetta et al., 2018; Boi et al., 2019; 2020).

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704

705 **4. Conclusions**

706 This study aimed at evaluating phytoremediation properties of *H. italicum* for PHEs (Cd, Cu,
707 Fe, Ni, Pb and Zn) by the determination of elements concentration on the roots, stems and
708 leaves, and by the related root/soil and leaf/root ratios on plants collected from mine tailing
709 deposits in contaminated sites (CS, Montevecchio mine, Sardinia) and in uncontaminated sites
710 (US). Moreover, a literature review on the phytoremediation properties of *Helicrysum* Spp. has
711 been executed in order to elucidate its phytoremediation potentiality.

712 The Cd, Pb and Zn resulted to be the most bioavailable PHEs in soils collected from
713 Montevecchio mine district compared to US, also exceeding the thresholds of the Italian
714 environmental legislation.

715 Element concentrations in plants from CS were higher than those from US. Leaf element
716 concentrations were higher compared to stems and roots, although statistically significant only
717 for Cu, Fe and Zn in CS. Interaction mechanisms (synergistic effects) between Cd and Zn, Fe
718 and Ni, and Cu and Zn in *H. italicum* plants grown on CS were detected.

719 The medians for root/soil ratio were > 1 for Cu, Fe, Ni and Pb in both CS and US, meaning that
720 *H. italicum* was able to accumulate bioavailable elements in roots. Cadmium and Zn in CS had
721 root/soil ratio close to one (1.39 and 1.21, respectively), suggesting their limited uptake when
722 soil bioavailable concentrations are elevated. Based on the root/soil ratios *H. italicum* is
723 considered a metal tolerant species.

724 The medians leaf/root ratio in CS were 2.36, 1.54, 1.37 and 1.34 for Fe, Zn, Pb and Cu,
725 respectively, while ratios were < 1 for Cd (0.75) and Ni (1). Similar ratios were also calculated
726 for US proving that *H. italicum* has a weak phytoextraction capacity.

727 The literature review on the phytoremediation potentiality of *H. italicum* confirmed the analytical
728 findings of this study. Indeed, Zn, Pb and Fe were the most abundant elements in *H. italicum*
729 plants grown on contaminated soils with concentrations around 10^3 mg/kg, $5 \cdot 10^2$ mg/kg and
730 10^3 mg/kg, respectively. The Cu, Cd and Ni, were poorly absorbed by plants and their
731 concentrations ranged around 10 mg/kg of Cu and below 10 mg/kg of Cd and Ni. In light of

732 this, *H. Italicum* cannot be considered a hyperaccumulator species. Overall, the root/soil
733 bioavailable ratio > 1 for all elements suggested the use of *H. italicum* for phytostabilization in
734 mine areas as a pioneering strategy of remediation.

735 Given that, *H. italicum* is a spontaneous and perennial species, which guarantees the canopy
736 cover all throughout the year, rehabilitates the vegetation cover and it can be propagated by
737 directly sowing seeds on contaminated soil, it can be recommended for phytostabilization of
738 abandoned mine districts and for stabilization of mine tailing from land management point of
739 view.

740

741

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749

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