

Alma Mater Studiorum Università di Bologna Archivio istituzionale della ricerca

The relationship between species and spectral diversity in grassland communities is mediated by their vertical complexity

This is the final peer-reviewed author's accepted manuscript (postprint) of the following publication:

Published Version:

Conti, L., Malavasi, M., Galland, T., Komárek, J., Lagner, O., Carmona, C.P., et al. (2021). The relationship between species and spectral diversity in grassland communities is mediated by their vertical complexity. APPLIED VEGETATION SCIENCE, 24(3), 1-8 [10.1111/avsc.12600].

Availability:

This version is available at: https://hdl.handle.net/11585/829957 since: 2021-08-19

Published:

DOI: http://doi.org/10.1111/avsc.12600

Terms of use:

Some rights reserved. The terms and conditions for the reuse of this version of the manuscript are specified in the publishing policy. For all terms of use and more information see the publisher's website.

This item was downloaded from IRIS Università di Bologna (https://cris.unibo.it/). When citing, please refer to the published version.

(Article begins on next page)

This is the final peer-reviewed accepted manuscript of:

Conti, Luisa; Malavasi, Marco; Galland, Thomas; Komárek, Jan; Lagner, Ondřej; Carmona, Carlos P.; Bello, Francesco; Rocchini, Duccio; Šímová, Petra: *The relationship between species and spectral diversity in grassland communities is mediated by their vertical complexity*

APPLIED VEGETATION SCIENCE VOL. 24 ISSN 1402-2001

DOI: 10.1111/avsc.12600

The final published version is available online at: <u>https://dx.doi.org/10.1111/avsc.12600</u>

Terms of use:

Some rights reserved. The terms and conditions for the reuse of this version of the manuscript are specified in the publishing policy. For all terms of use and more information see the publisher's website.

This item was downloaded from IRIS Università di Bologna (<u>https://cris.unibo.it/</u>)

When citing, please refer to the published version.

The relationship between species and spectral diversity in grassland communities is mediated by their vertical complexity.

Running title: Spectral-species diversity in grasslands

Luisa Conti^{1,2}, Marco Malavasi¹, Thomas Galland^{2,3}, Jan Komárek¹, Ondřej Lagner¹, Carlos P. Carmona⁴, Francesco de Bello^{3,5}, Duccio Rocchini^{1,6,7,8}, Petra Šímová¹

¹Faculty of Environmental Sciences, Czech University of Life Sciences, Prague, Czech Republic

²Institute of Botany, The Czech Academy of Sciences, Třeboň, Czech Republic

³Department of Botany, Faculty of Sciences, University of South Bohemia, České Budějovice, Czech Republic

⁴ Institute of Ecology and Earth Sciences, University of Tartu, Tartu, Estonia

⁵ Centro de Investigaciones sobre Desertificación (CSIC-UV-GV), Valencia, Spain

⁶ Department of Biological, Geological and Environmental Sciences, Alma Mater Studiorum University of Bologna, Bologna, Italy

⁷Department of Cellular, Computational and Integrative Biology (CIBIO), University of Trento, Povo (TN), Italy

⁸Department of Biodiversity and Molecular Ecology, Research and Innovation Centre, Fondazione Edmund Mach, S. Michele all'Adige (TN), Italy

Correspondence:

Luisa Conti

Faculty of Environmental Sciences, Czech University of Life Sciences, Prague, Czech Republic.

Email: conti@fzp.czu.cz

1

2 Abstract

3 Aims

4 The link between spectral diversity and ground plant biodiversity is one promising approach

5 for using remote sensing for biodiversity assessment. Nevertheless, there is little evidence to

6 whether this link is maintained at fine scales, as well as to how it is influenced by vegetation's

7 vertical complexity. Here we test, at community level in grasslands, the link between diversity

of the spectral signal (SDiv) and taxonomic diversity (TDiv), and the influence of verticalcomplexity.

10 *Methods*

We used 196 1.5×1.5 m experimental communities with different biodiversity levels. To 11 measure vertical complexity, we quantified height diversity (HDiv) of the most abundant 12 species in the community. TDiv was calculated using Shannon index based on species cover. 13 Canopy spectral information was gathered using an UAV mounted with a multi-spectral sensor 14 providing spectral information via six 10 nm bands covering the visible and near-infrared region 15 at 3 cm spatial resolution. We measured SDiv in a core area of 1×1 m within the communities 16 as mean Euclidean distance of all pixels in a feature space spanned between the two first 17 components of a PCA calculated for the complete raster stack. We modelled SDiv through 18 mixed effect linear models, using TDiv in interaction with HDiv as fixed effect predictors. 19

20 *Results*

Contrarily to our expectations, TDiv was negatively linked to SDiv. The diversity in plant height was positively related to SDiv. More importantly, diversity in plant height and TDiv had a significant negative interaction, meaning the more complex the vegetation was in terms of height, the more the SDiv-TDiv relationship became negative.

25 *Conclusions*

Our results suggest that in order to exploit the SDiv-TDiv link for monitoring purposes, it needs to be contextualized. Moreover, they highlight that communities' functional characteristics (i.e. plant height) mediate such link, calling for new insights on the relation between SDiv and functional diversity.

30

Keywords: spectral diversity, grasslands, species diversity, community scale, vertical
stratification, height diversity, taxonomic diversity, mixed effect models, remote sensing, multispectral sensor, UAV.

34 Introduction

Improving our ability to monitor biodiversity across scales is critical to understand its role in 35 shaping ecosystem functioning across the world, as well as to track ecological communities' 36 responses to rapid environmental changes. Despite a growing interest in biodiversity in recent 37 decades, the ability to measure and monitor biodiversity through traditional methods continues 38 39 to lag behind current global changes and its related challenges (Palmer et al. 2002; Wang and Gamon 2019). National biodiversity monitoring programs differ widely, most data sets are 40 inconsistent, and few data are shared openly; further, measuring biodiversity on the ground is 41 laborious, requiring a lot of time, human, and financial resources, as well as being limited by a 42 lack of standardized procedures for reproducible data gathering (Rocchini, Marcantonio, and 43 Ricotta 2017; but see also the GEOBON initiative, Scholes et al. 2018). Recently, it has been 44 suggested that variation in plant diversity could be promptly monitored through remote sensing 45 (RS) techniques (Féret and Asner 2014; Jetz et al. 2016; Lausch et al. 2016; Skidmore et al. 46 2015; Wang and Gamon 2019), which provide a continuous source of information on different 47 facets of plant diversity on a wide range of scales in a consistent, borderless, and repeatable 48 49 manner (Turner 2014).

Among the many approaches used to estimate plant diversity through RS (for a review 50 see Wang and Gamon 2019), "Spectral Diversity" (SDiv), has been gaining momentum 51 (Rocchini et al. 2021). SDiv, defined as the variation in spectral reflectance across a set of 52 neighbouring pixels (Palmer et al. 2002; Ustin and Gamon 2010), was originally proposed 53 54 under the framework of Spectral Variation Hypothesis (SVH; Palmer 1992). This framework suggests that a higher spectral diversity can be interpreted as a higher ecosystem spatial 55 heterogeneity, thus as a potentially higher number of available niches and their related species. 56 However, this relationship also depends on the scale of the RS data. With the development of 57 RS techniques and the advances of instruments such as Unmanned Aerial Vehicles (UAV) it is 58 now possible to flexibly acquire data at very high spatial resolutions (Whitehead and 59 Hugenholtz 2014). Such high-resolution data can capture the direct link between the spectral 60 61 information at leaf and canopy level and the functional characteristics of the different plant species in the analysed area (Asner and Martin 2009; Wang and Gamon 2019). It follows that 62 with a larger availability of fine spatial resolution images, SDiv can be used to directly estimate 63 plant species diversity, since a higher spectral diversity should relate to a higher number of 64 species with corresponding species-specific optical traits at leaf and canopy level (Asner and 65 66 Martin 2009; Wang and Gamon 2019). Several RS indices based on SDiv have been applied so 67 far to assess plant diversity and they have been recently grouped in three main categories (Wang and Gamon 2019): metrics based on variation in vegetation indices (e.g. the Normalized 68 69 Difference Vegetation Index - NDVI) which often scales with species richness (Gillespie 2005); metrics based on information theory (e.g. spectral entropy) providing information on the 70 "dimensionality" of a dataset for a given area, which can then be related to the number of 71 species present in that area (Rocchini et al. 2021); and metrics based on spectral species where 72 73 these are considered proxies or analogues for biological species, and spatial variation in spectral 74 species can be used to infer species richness (see Féret and Asner 2014). .

Nonetheless, while the link between SDiv and plant diversity has been confirmed in 75 several empirical studies (Rocchini et al. 2016; Wang et al. 2018; Wang and Gamon 2019), it 76 has also been criticized for being unstable and not reliable in every context (Schmidtlein and 77 Fassnacht 2017). These contradictory findings might be due to a lack of studies systematically 78 considering the most important factors influencing such relationship such as scale, phenology, 79 used metrics, and vertical structure of the observed ecosystems. In particular, the vertical 80 81 structure of the vegetation may hamper the optical detection of some of the plant species in the 82 area. In a closed vegetation with a complex vertical structure, taller species would tend to obscure short species from RS detection (occlusion effect), which would lead to an 83 84 underestimation of ground plant diversity. However, a complex vertical structure could also give rise to underexposed pixels or shadow patterns, which on the contrary could determine a 85 86 higher diversity of spectral values and lead to an overestimation of ground plant diversity. To our knowledge, no studies have addressed the influence of the vertical complexity on the link 87 88 between SDiv and plant diversity. Moreover, no study has addressed this issue in herbaceous communities, where despite having one single canopy level, species height variability can give 89 90 rise to an important gradient of vertical complexity (Brown and Cahill 2019).

Here, we focus on a grassland experiment situated on a mesic meadow in South 91 Bohemia (Czech Republic) to test the relationship between SDiv and communities' taxonomic 92 plant diversity (TDiv) while accounting for the effect of the grassland vertical complexity. We 93 expect this relationship to be positive, although weaker in more vertically complex 94 95 communities, due to the occlusion effect. The unique settings of the experiment enabled us to test this relationship by using a considerable span of plant diversity within grassland 96 ecosystems. Semi-natural meadows are considered to be one of the most important plant species 97 diversity hotspots worldwide (Dengler et al. 2014) and improving our ability of diversity 98

99 monitoring is essential to mitigate the loss of vascular plant species and associated ecosystem
100 services (Edwards and Kučera 2019).

101

102 Methods

To explore the relationship of SDiv and plant diversity at community level, we used data from 103 a permanent grassland experiment, established in 2015 and situated on a mesic meadow in 104 South Bohemia, Czech Republic (49.331N, 15.003E; for further details see Galland et al., 105 106 2019). The experiment comprises 40 mesic grassland communities which cover independent gradients of plant functional and phylogenetic diversity (Figure 1). Each sown community 107 108 comprised a combination of 6 species obtained from a pool of 19 species naturally present in Czech mesic meadows. Each community was sown in two randomly situated 1.5×1.5 m plots 109 (fertilized and unfertilized), with a buffer zone of 0.5 m between each of them. Monoculture 110 plots were also sown in three replicates each, for a total of 196 plots. These represent the sown 111 communities, which are the basis for the actual communities present in each plot, composed by 112 sown and colonizer species (species present in the surroundings that invaded the experiment), 113 which were the communities actually sampled (see Supporting Information, Appendix S1). 114

In May 2018, at the peak of the growing season, species composition and percentage cover were measured in each of the plots. Data regarding vegetative plant height was measured on site for each species (sown or colonizer) that cumulatively covered 80% in at least one of the plots, on fully developed individuals and following standardized protocols (Pérez-Harguindeguy et al. 2013).

Shortly after the vegetation sampling, multi-spectral imagery at high spatial resolution 120 (~3 cm) was acquired over the grassland experiment by an Unmanned Aerial Vehicle. 121 122 Specifically, a hexacopter Kingfisher (Robodrone Industries) equipped with a Tetracam µMCA6 global shutter (Kelcey and Lucieer 2012), providing optical image data across six 10-123 20 nm width bands ranging from 490 to 900 nm (Table 1). Acquired imagery was processed 124 using Structure from Motion and Multi-View Stereo algorithms in Photoscan image-matching 125 software (version 1.4, Agisoft LLC 2018). Orthomosaics were built and accurately 126 georeferenced using six Ground Control Points placed across the experimental site and 127 surveyed with a Leica GPS1200 GNSS aperture in RTK mode. The surface reflectance values 128 were calculated using a grey calibration target for which the spectral properties were known 129 through spectrometer measurements to produce radiometrically calibrated orthomosaics. 130

132 Statistical analyses

In each plot, first we determined vertical complexity, TDiv, and SDiv. Vertical complexity was quantified through diversity in plant height (HDiv), measured as mean Euclidean distances between the height values (normalized via log-transformation) of the species present in each of the plots, weighted by the abundance of each species in terms of cover (de Bello et al. 2016). TDiv was calculated using Shannon index based on species cover (Shannon and Weaver 1949).

138 To measure SDiv we used an index based on spectral entropy (Wang and Gamon 2019), which is now the main approach used to quantify spectral diversity as a proxy for plant diversity 139 140 (see Schweiger et al. 2018, Wang et al 2018, Rocchini et al. 2021). We first performed a standardized PCA (centered and scaled input image for equal weighting of all layers) to 141 summarize the values of the six bands (using 'rasterPCA' function in R package 'RStoolbox'), 142 a common practice for spectral entropy metrics (Rocchini et al 2021). Then, for each plot, we 143 extracted the values of the first and second principal component axes derived from the PCA. 144 These values were extracted considering the pixels within a core area of 1×1 m inside the plot, 145 in order to avoid edge areas and thus avoid sampling bias due to potential overlap between 146 adjacent communities or between buffer zone and the actual sampled plot (Figure 1). We then 147 calculated the mean Euclidean distance of these two sets of values in each of the core areas. We 148 also quantified two separate SDiv measures, corresponding to the spectral diversity when 149 considering only the first axis (SDiv₁) or only the second axis (SDiv₂), results regarding these 150 151 indexes are in line with the results presented here and can be found in Supporting Information, Appendix S2. We modelled the variation in SDiv through mixed effect linear models (R 152 package 'nlme4'), using TDiv and HDiv (both scaled and centred), separately and in interaction, 153 as fixed effect predictors, and the combination of diversity level (type of community sowed) 154 and fertilization treatments as random effect. In addition, for each model we calculated marginal 155 and conditional R²s to assess the variance explained by fixed effect and both fixed and random 156 effects, respectively (function 'r.squaredGLMM' in R package 'MuMIn'). A visual 157 representation of all the values used in the models on the study area is found in Appendix S3, 158 159 Supporting Information. Finally, we controlled for potential effect of visible flowerheads in our images. We compared values of SDiv and HDiv in plots with or without flowers though 160 Kruskal-Wallis rank sum test. We did not find any effect of the flowers in either SDiv or HDiv 161 (see Supporting Information, Appendix S4). All analyses were performed in R statistical 162 software (version 3.5.2, Team R Core 2013) 163

164

165 **Results**

The two first principal components derived from the standardized PCA accounted for nearly 166 90% of the variance in the spectral signal of the communities (Table 2). When correlating these 167 components to the reflectance values in each of the bands, Pearson's correlation coefficients 168 showed that the first component is highly and positively correlated with bands 2, 4, and 6 169 (Green, Red Edge, and Near Infrared 2, respectively), while the second component is positively 170 correlated mainly with bands 1 and 3 (Blue and Red respectively), while negatively correlated 171 with bands 5 and 6 (Near Infrared 1 and 2; Table 3). Therefore, the higher variance in the 172 spectral signal of these communities is enclosed in the red edge, near infrared, and green 173 regions, which is represented in the first component axis. The rest of the spectral variance, 174 175 specially the one in the blue and red regions, is described mainly by the second component axis.

The mixed model showed that TDiv was significantly linked to SDiv, although with a 176 negative coefficient (estimate = -0.26, standard error = 0.05, p-value < 0.001; Figure 2). On the 177 contrary, HDiv, i.e. the vertical structure of the communities, was positively related to SDiv 178 (estimate = 0.36, standard error = 0.05; p-value < 0.001 Fig. 2). More importantly, HDiv and 179 TDiv had a significant negative interaction (estimate = -0.27, standard error = 0.06, p-value < 180 0.001; Fig. 2), meaning that the more complex the vertical structure of the vegetation, the more 181 negative the relationship between TDiv and SDiv became (Fig. 2). The marginal R² calculated 182 was of 0.36, thus the fixed effects explained more than 30% of the variability. Together with 183 the random effects, the model explained 61% of the total variability (conditional $R^2 = 0.61$). 184

185

186 Discussion

Generally, our results show how the relationship between SDiv and communities' taxonomic 187 diversity (TDiv) is mediated by grasslands' vertical complexity. The negative relationship 188 found between SDiv and TDiv is in contrast with previous empirical studies (Rocchini et al. 189 2016; Wang et al. 2018; Wang and Gamon 2019). This found relationship highlights that SDiv 190 can't be used routinely to predict plant species diversity (e.g. Schmidtlein and Fassnacht 2017). 191 This is also suggested by the fact that in our models, the random effect (i.e. the fertility and 192 diversity treatment of each plot), explained nearly 30 percent of the variability in SDiv. Our 193 results do show both a relationship between SDiv and HDiv, and an interaction effect of HDiv 194

195 and TDiv. This means that herbaceous communities with a higher diversity in terms of height, i.e. a more complex vertical structure, have a more diverse spectral signal. More importantly, 196 197 herbaceous communities with the most complex vertical structure manifest a more negative relationship between SDiv and TDiv; meaning that these communities will express high SDiv 198 199 even when having low TDiv (probably because of taller species contributing to spectral 200 diversity by casting shadows and hence generating underexposed pixels) and will express low 201 SDiv even with high TDiv (due to potential occlusion effect). These relationships were also confirmed when analysing the spectral variability using single PCA axes, meaning that they are 202 valid across different optical features of the communities (Supporting Information, Appendix 203 S2). 204

Results regarding the PCA highlighted that the highest variability in terms of spectral 205 signal is expressed in the wavelength range corresponding to the RE, G and NIR2 spectral 206 bands, while the rest of the spectral variability is expressed in the bands B and R but also 207 negatively through NRI1 and NRI2 (Table 2). The variability expressed in these ranges is linked 208 209 to species-specific differences in pigment content, leaf structure and canopy structural 210 components such as leaf area index (Jones and Vaughan 2010). The second principal 211 component linked mainly to R and B bands, which relate to pigments, including carotene and 212 xanthophylls. However, this axis is also expressing a negative correlation with NIR2, in contrast with the first axis being positively correlated to NIR2. This variability could reflect the different 213 214 composition of the communities also in terms of phenological stage, e.g. flowers presence and leaf senescence variability. These results show that despite the dimensionality reduction applied 215 216 to the spectral variance, SDiv metrics based on PC components, which belong to the 217 information theory type of metrics, are able to convey information across all bands as well as 218 important vegetation properties.

We observed an overall negative relationship between SDiv and TDiv, which suggests 219 that the inconsiderate use of spectral variation to quantify species diversity, at this scale, might 220 be misleading (Fig. 3). The use of spectral diversity as a surrogate of species diversity has been 221 indeed long debated, as this link is highly dependent on various factors that might be hard to 222 223 control (Schmidtlein and Fassnacht, 2017). Indeed, our results seem to confirm the contextdependency of this method, with the relationship between SDiv and TDiv in this case being 224 opposite to expectations, and generally driven by the diversity of heights, i.e. vertical 225 complexity, in the communities analysed (as seen in Fig. 3). Despite our results being 226 227 supposedly able to capture the direct spectral footprint of each plant individual given their high spatial resolution, i.e. exploiting the known direct link between the spectral information at leaf and canopy level and the characteristics of plant species (Asner and Martin 2009), the relationship between SDiv and TDiv was actually more negative when analysing more vertically complex communities. Reasons leading to the variation of the SDiv-TDiv relationship may be connected partially to spatial grain effects, as well as to the effect of the spatial configuration of the vegetation on the spectral diversity, as discussed below.

234 The effect of spatial grain on the link between SDiv and TDiv has been highlighted by several authors (Rocchini et al. 2016; Wang et al. 2018). In particular, Wang et al. (2018), using 235 236 central-European experimental grasslands, highlighted that in order to observe a positive relationship, the used pixel size should reflect the individual plant size. Despite the use of a fine 237 238 spatial grain (~3 cm, corresponding to the average size of the individuals found in our communities), the SDiv measure used presented a negative link to TDiv. In our system and at 239 this scale, more taxonomically diverse communities were less diverse in terms of spectra, and 240 less taxonomically diverse species had higher spectral diversity. One explanation might be that 241 242 within this particular setting species' taxonomic identity doesn't correspond to the species' 243 spectral identity. On the one hand individuals of the same species might have slightly different spectral signals, because of intra-specific variability or different phenological stages; on the 244 245 other hand, different species might have similar spectral signals, as they might possess similar plant functional traits. In order to better discriminate between species specific signals, a 246 247 hyperspectral sensor could be preferred over the here adopted multi-spectral sensor (e.g. Lopatin et al. 2017; Möckel et al. 2014). In any case, plant functional types (defined by plant 248 249 traits), are more closely and more mechanistically linked to the spectral identity of species. 250 Therefore, functional diversity would be more strongly related to spectral variability (as seen 251 in Schweiger et al. 2018; and Ustin and Gamon 2010), compared to species or taxonomic diversity. Which is also highlighted by the direct link we found between diversity in spectral 252 variability and plant height diversity. 253

Our findings show that SDiv was linked more strongly to the vertical structure of the sampled communities, compared to TDiv (Figs. 1 and 3). While the link between vegetation structure and different remote sensing indicators has been highlighted by previous studies, especially in woody vegetation (Campos et al. 2018; Wood et al. 2012), the present work is the first work assessing this link using optical data within grasslands. Spatial configuration of vegetation is an important functional characteristic, tightly linked with different ecosystem functions such as biomass production and ecological niche availability (Zuppinger-Dingley et

261 al. 2014). Indeed, remote sensing techniques have been employed to quantify spatial configuration of vegetation and its spatial variability. In particular, regarding optical data, a 262 263 series of textural indices have been linked to vegetation vertical structure (Campos et al. 2018; Wood et al. 2012). These studies have been performed at wide extent and coarse spatial 264 265 resolution, as they focus mainly on differences in the vertical structure of different vegetation types. Our results suggest that, when using the proper resolution, communities' spatial structure 266 267 can be quantified even within a seemingly homogeneous vegetation type, with diversity of plant height being directly related to the diversity of reflectance within a certain area. 268

Finally, not only HDiv was directly linked to SDiv, it mediated the relationship between 269 SDiv and TDiv, and thus provided an explanation for the found negative link between the two 270 variables. For highly structured communities in terms of their vertical configuration, we found 271 a more negative relationship between the diversity of species measured at ground level and the 272 diversity of the spectral signal (negative interaction, Fig. 2 and 3). Moreover, communities with 273 a low species diversity but high HDiv showed a high spectral diversity. This may depend on 274 275 the horizontal configuration of the communities, which is constrained by the pre-established size of the experimental plot. At high HDiv, the species-poor communities might have more 276 space compared to species-rich communities which express a low SDiv. That is to say that 277 278 wherever in our experimental plots we observe a species-rich community with high HDiv, its species, in order to populate the plot, may be squeezed in a relatively small area, overlapping 279 280 one another in the vertical and horizontal space. In this case, fewer species would be able to lean out in the higher strata and mask the lower layers, which results in a more homogeneous 281 282 canopy sensed and therefore lower SDiv (i.e., occlusion effect).

283

284 Conclusions

Ultimately, our results show that the link between SDiv and species diversity in grasslands is mediated by vegetation's vertical complexity. In particular, they suggest that in order to exploit the link for monitoring purposes, it needs to be contextualized. Despite recent evidence pointing at the use of remote sensing data for estimating species richness and diversity (Rocchini et al. 2016; Wang et al. 2018; Wang and Gamon 2019), this approach is not straightforward, especially when focusing on grassland systems on community scale. However, the relationship between SDiv and structural characteristics of the communities found here calls for more insights regarding the link between SDiv and functional diversity (as seen in Schweiger et al.2018).

Finally, although our findings partly reject the theorized link between SDiv and plant species diversity, this does not mean that spectral diversity cannot be useful for biodiversity monitoring. Besides the promising usefulness of its link with functional diversity (e.g. Schweiger et al. 2018), our results indicate that spectral diversity can be used when appropriately calibrated to the context analysed. In particular, our work highlights the importance of considering vegetation structural characteristics affecting the relationship between spectral diversity and plant diversity.

301

Acknowledgements: We want to acknowledge the people involved in maintenance and data
sampling of the experimental communities, particularly Miroslav Šrůtek who gave access to his
land.

Author contributions: LC and MM conceived the idea, CC designed the experimental communities, TG, FdB, and CC sampled and/or analysed ground data, JK and OL sampled and cleaned the drone images, DR and PS helped analyse spectral data, LC analysed the complete dataset and wrote the first draft of the manuscript. All authors contributed to the writing.

309 Data availability statement: Aggregated data (metrics) used in this manuscript is available in
310 Supporting Information Appendix S1, Table S1.1.

311

313 **References**

- 314 Agisoft LLC (2018) Agisoft PhotoScan Professional (v. 1.4). [Software].
- Asner, G.P. and Martin, R.E. (2009) Airborne spectranomics: mapping canopy chemical and
- taxonomic diversity in tropical forests. *Frontiers in Ecology and the Environment* 7, 269–276.
 doi: 10.1890/070152.
- de Bello, F., Carmona, C.P., Lepš, J., Szava-Kovats, R. and Pärtel, M. (2016) Functional
- diversity through the mean trait dissimilarity: resolving shortcomings with existing paradigms
 and algorithms. *Oecologia* 180, 933-940. doi: 10.1007/s00442-016-3546-0
- Brown, C. and Cahill, J.F. (2019) Vertical size structure is associated with productivity and
- 322 species diversity in a short-stature grassland: Evidence for the importance of height variability
- 323 within herbaceous communities. *Journal of Vegetation Science* 30, 789–798. doi:
- **324** 10.1111/jvs.12785
- 325 Campos, V.E., Gatica, G.M., Cappa, F.M., Giannoni, S.M. and Campos, C.M. (2018) Remote
- sensing data to assess compositional and structural indicators in dry woodland. *Ecological*
- 327 Indicators 88, 63-70. doi: 10.1016/J.ECOLIND.2018.01.032.
- 328 Dengler, J., Janišová, M., Török, P. and Wellstein, C. (2014) Biodiversity of Palaearctic
- 329 grasslands: A synthesis. *Agriculture, Ecosystems and Environment* 182, 1–14. doi:
- 330 10.1016/j.agee.2013.12.015.
- Edwards, K.R. and Kučera, T. (2019) Management effects on plant species composition and
 ecosystem processes and services in a nutrient-poor wet grassland. *Plant Ecology* 220, 1009–
 1020. doi: 10.1007/s11258-019-00970-9.
- Féret, J.B. and Asner, G.P. (2014) Mapping tropical forest canopy diversity using highfidelity imaging spectroscopy. *Ecological Applications* 24, 1289–1296. doi: 10.1890/13-1824.1.
- 337 Galland, T., Adeux, G., Dvořáková, H., E-Vojtkó, A., Orbán, I., Lussu, M. et al. (2019)
- Colonization resistance and establishment success along gradients of functional and
- phylogenetic diversity in experimental plant communities. *Journal of Ecology* 107, 2090–
 2104. doi: 10.1111/1365-2745.13246.
- Gillespie, T.W. (2005) Predicting woody-plant species richness in tropical dry forests: a case
 study from south Florida, USA. *Ecological Applications* 15, 27–37. doi: 10.1890/03-5304.
- Jetz, W., Cavender-Bares, J., Pavlick, R., Schimel, D., Davis, F.W., Asner, G.P. et al. (2016)
- 344 Monitoring plant functional diversity from space. *Nature Plants* 2,16024. doi:
- 345 10.1038/NPLANTS.2016.24.
- Jones, H.G. and Vaughan, R.A. (2010) Remote sensing of vegetation: principles, techniques,and applications. Oxford University Press.
- Kelcey, J. and Lucieer, A. (2012) Sensor correction of a 6-band multispectral imaging sensor
 for UAV remote sensing. *Remote Sensing* 4, 1462–1493. doi: 10.3390/rs4051462.
- 350 Lausch, A., Bannehr, L., Beckmann, M., Boehm, C., Feilhauer, H., Hacker, J.M. et al. (2016)
- 351 Linking Earth Observation and taxonomic, structural and functional biodiversity: local to
- ecosystem perspectives. *Ecological Indicators* 70, 317–339. doi:
- 353 10.1016/j.ecolind.2016.06.022.

- 354 Lopatin, J., Fassnacht, F.E., Kattenborn, T. and Schmidtlein, S. (2017) Mapping plant species
- in mixed grassland communities using close range imaging spectroscopy. *Remote Sensing of*
- 356 *Environment* 201, 12-23. doi: 10.1016/j.rse.2017.08.031.
- 357 Möckel, T., Dalmayne, J., Prentice, H.C., Eklundh, L., Purschke, O., Schmidtlein, S. and Hall,
- 358 K. Classification of Grassland Successional Stages Using Airborne Hyperspectral Imagery.
- 359 *Remote Sensing* 6, 7732-7761. doi: 10.3390/rs6087732
- 360 Palmer, M.W. Earls, P.G., Hoagland, B.W., White, P.S. and Wohlgemuth, T. (2002)
- 361 Quantitative tools for perfecting species lists. *Environmetrics* 13, 21–137. doi:
- 362 10.1002/env.516.
- 363 Pérez-Harguindeguy, N., Díaz, S., Garnier, E., Lavorel, S., Poorter, H., Jaureguiberry, P. et al.
- 364 (2013) New handbook for standardised measurement of plant functional traits worldwide.
- 365 Australian Journal of Botany 6, 167-234. https://doi.org/10.1071/BT12225
- Rocchini, D., Boyd, D.S., Féret, J-B., Foody, G.M, He, K.S., Lausch, A., et al. (2016)
- Satellite remote sensing to monitor species diversity: potential and pitfalls. *Remote Sensing in Ecology and Conservation* 2, 25–36. doi: 10.1002/rse2.9.
- Rocchini, D., Marcantonio, M. and Ricotta, C. (2017) Measuring Rao's Q diversity index
- from remote sensing: an open source solution. *Ecological Indicators* 72, 234–238. doi:
 10.1016/j.ecolind.2016.07.039.
- 372 Rocchini, D., Thouverai, E., Marcantonio, M., Iannacito, M., Da Re, D., Torresani, M., et al.
- 373 (2021) rasterdiv an Information Theory tailored R package for measuring ecosystem
- heterogeneity from space: to the origin and back. *Methods in Ecology and Evolution*.
- Accepted Author Manuscript. doi: 10.1111/2041-210X.13583
- 376 Schmidtlein, S. and Fassnacht, F.E. (2017) The spectral variability hypothesis does not hold
- across landscapes. *Remote Sensing of Environment* 192, 114–125. doi:
- 378 10.1016/j.rse.2017.01.036.
- 379 Scholes, R. J., Mace, G. M., Turner, W., Geller, G. N., Jürgens, N., Larigauderie, A., et al.
- 380 (2008). Ecology: Toward a global biodiversity observing system. *Science*. 321, 1044-1045.
 381 doi: 10.1126/science.1162055
- 382 Schweiger, A.K., Cavender-Bares, J., Townsend, P.A., Hobbie, S.E., Madritch, M.D., Wang,
- R., Tilman, D. and Gamon, J.A. (2018) Plant spectral diversity integrates functional and
- 384 phylogenetic components of biodiversity and predicts ecosystem function. *Nature Ecology &*
- 385 *Evolution* 2, 976-982. doi: 10.1038/s41559-018-0551-1
- Shannon, C.E. and Weaver, W. (1949) A mathematical model of communication. Urbana, IL:University of Illinois Press.
- 388 Skidmore, A.K., Pettorelli, N., Coops, N.C., Geller, G.N., Hansen M., Lucas, R. et al. (2015)
- Environmental science: Agree on biodiversity metrics to track from space. *Nature* 523, 403–
 405. doi: 10.1038/523403a.
- Team R Core. (2013) R. A language and environment for statistical computing (v. 3.5.2)
 [Software]
- 393 Turner, W. (2014) Sensing biodiversity. *Science* 346, 301–302. doi:
- 394 10.1126/science.1256014.
- 395 Ustin, S.L. and Gamon, J.A. (2010) Remote sensing of plant functional types. *New*

- 396 *Phytologist* 186, 795–816. doi: 10.1111/j.1469-8137.2010.03284.x.
- Wang, R., Gamon, J.A., Cavender-Bares J., Townsend, P.A. and Zygielbaum, A.I. (2018) The
- 398 spatial sensitivity of the spectral diversity-biodiversity relationship: an experimental test in a
- prairie grassland. *Ecological Applications* 28, 541–556. doi: 10.1002/eap.1669.
- Wang, R. and Gamon, J.A. (2019) Remote sensing of terrestrial plant biodiversity, *Remote Sensing of Environment* 231, 111218. doi: 10.1016/j.rse.2019.111218.
- Whitehead, K. and Hugenholtz, C.H. (2014) Remote sensing of the environment with small
 unmanned aircraft systems (UASs), part 1: a review of progress and challenges. *Journal of Unmanned Vehicle Systems* 2, 69–85. doi: 10.1139/juvs-2014-0006.
- Wood, E.M., Pidgeon, A.M., Radeloff, V.C. and Keuler, N.S. (2012) Image texture as a
 remotely sensed measure of vegetation structure. *Remote Sensing of Environment* 121, 516526. doi: 10.1016/J.RSE.2012.01.003.
- 408 Zuppinger-Dingley, D., Schmid, B., Petermann, J.S., Yadav, V., De Deyn, G.B. and Flynn,
- 409 D.F. (2014) Selection for niche differentiation in plant communities increases biodiversity
- 410 effects. *Nature* 515, 08-111.doi: 10.1038/nature13869.
- 411 Appendix S1: Ground diversity values for the communities sampled and spectral diversity values
- 412 measured at each plot are presented in a separate file "TableS1.1.txt"
- 413 Appendix S2: Linear mixed models using SDiv₁ and SDiv₂.
- 414 Appendix S3: Visual representation of SDiv, TDiv, and HDiv values in the study area.
- 415 Appendix S4: Flowers effect on SDiv and HDiv
- 416

Band	Band peak	Band full width at	Corresponding	Abbreviation
	(nm)	half maximum	range	
		(nm)		
Band 1	490	485 - 495	Blue	В
Band 2	550	545 - 555	Green	G
Band 3	680	675 - 685	Red	R
Band 4	720	715 - 725	Red edge	RE
Band 5	800	795 - 805	Near infra-red	NIR1
Band 6	900	890 - 910	Near infra-red	NIR2

Table 1. Spectral bands measured and their corresponding range.

Table 2. Variability explained by the principal components derived from the standardized PCA

420 (centered a	and scaled	input la	ayers)	performed	over the 6	spectral	bands.
-------	------------	------------	----------	--------	-----------	------------	----------	--------

	Comp.1	Comp.2	Comp.3	Comp.4	Comp.5	Comp.6
Standard	2.0532060	1.0874347	0.54525290	0.38164219	0.3254505	0.230133368
deviation						
Proportion	0.7026091	0.1970857	0.04955012	0.02427513	0.0176530	0.008826895
of Variance						
Cumulative	0.7026091	0.8996949	0.94924497	0.97352010	0.9911731	1.00000000
Proportion						

Table 3. Pearsons' correlation estimates between principal components 1 and 2, and spectral

423 bands

	Band.1 (B)	Band.2 (G)	Band.3 (R)	Band.4	Band.5	Band.6
				(RE)	(NIR1)	(NIR2)
Comp.1	0.7725979	0.9039081	0.8194706	0.91142	0.7912322	0.8206282
Comp.2	0.5090558	0.2589284	0.5005981	-0.1966337	-0.5297212	-0.535224



Figure 1. RGB image of the multi-spectral image sampled. Whole area in the left panel, zoomin on the single plots in the right panel. Community plots are shown with black lines while the
core areas considered in the analyses are delineated in cyan.



Figure 2. Coefficient plot of the mixed model showing estimate values for each explanatory
variable and their interaction, as well as their 95% confidence interval. R² marginal and
conditional are shown in-graph.



438 Figure 3. Predicted relationships between SDiv and TDiv derived from the mixed effect model,

439 when setting different values of HDiv (in legend).