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Hydro-Power Production and Fish Habitat Suitability: Assessing Impact and Effectiveness of Ecological Flows at Regional Scale

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Abstract

Anthropogenic activities along streams and rivers may be of major concern for fluvial ecosystems, e.g. abstraction and impoundment of surface water resources may profoundly alter natural streamflow regimes. An established approach aimed at preserving the behavior and distribution of fluvial species relies on the definition of ecological flows (e-flows) downstream of dams and diversion structures. E-flow prescriptions are usually set by basin authorities at regional scale, often without a proper assessment of their impact and effectiveness. On the contrary, we argue that e-flows should be identified on the basis of (i) regional and (ii) quantitative assessments. We focus on central Italy and evaluate the effects on habitat suitability of two near-threatened fish species (i.e. Barbel and Chub) and an existing hydro-power network when shifting from the current time-invariant e-flow policy to a tighter and seasonally-varying soon-to-be-enforced one. Our example clearly shows that: (a) quantitative regional scale assessments are viable even when streamflow observations are entirely missing at study sites; (b) a priori e-flows policies may impose releases that exceed natural streamflows for significantly long time intervals (weeks, or months); (c) unduly tightening e-flow policies may heavily impact regional hydro-power productivity (15% and 42% losses on annual and seasonal basis, respectively), yet resulting in

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either marginal or negligible improvements of fluvial ecosystem.

Keywords: water resources management, hydro-electric production, e-flow, PHABSIM, Barbel, Chub

1. Introduction

Hydro-power is the worldwide leading renewable source for electricity production, with a capacity increase of more than 30% between 2007 and 2015 (WCE, 2016). Despite its economic relevance, several environmental concerns are associated with hydro-power production. Indeed, hydro-power plants are known to severely affect flows downstream of abstraction points over limited time periods (Person et al., 2014; Vigano et al., 2016). Under these conditions, the river natural flow regime (NFR), defined as the river signature in terms of flow magnitude, frequency, timing, duration and rate of change (Poff et al., 1997), is no longer guaranteed. NFR is a key driver of ecological and geomorphological processes (Bunn and Arthington, 2002; Allan and Castillo, 2007; Young et al., 2011; Ceola et al., 2014; Ceola and Pugliese, 2014), and thus, any flow disturbance may significantly affect and alter fluvial ecosystem dynamics (see e.g. Poff and Allan, 1995; Bradford et al., 2011; Ceola et al., 2013; Vanzo et al., 2016).

As recognized by several water protection policies, e.g. the European Water Framework Directive (EU, 2000), the water laws in Africa (1998) and China (2002), the Australian Environment Act (1999), ecological flows (e-flows) are commonly defined in order to sustain freshwater ecosystems and the river ecological status.

The concept of e-flows has existed for more than 40 years (Tharme, 2003; Acreman and Dunbar, 2004; Snelder et al., 2014) and it is widely applied throughout the globe, though presenting significant differences across site-specific applications. E-flows can be generally grouped under two main categories, based on the methodology they rely upon.

On one side, one may find the classical hydrologically-based methods (e.g. minimum flow, flow percentiles, see Tharme, 2003). This category embeds easily-applicable and simple approaches that can be employed across large areas, but do not focus on any ecological variable, being thus somehow in contrast with the definition of e-flows. Quite frequently, e-flows defined within this category are described by constant flows during the year, thus

32 disregarding the inter-annual flow variability that controls species life stages
33 (Stromberg et al., 2010).

34 On the other side, there are the so-called micro-scale and meso-scale phys-
35 ical habitat modeling methods, based on in-situ and experimental measure-
36 ments to analyze optimal environmental conditions for target species. Several
37 habitat suitability models are described in the scientific literature, see e.g.
38 PHABSIM (Bovee, 1982), RHYHABSIM (Jowett, 2010), RIVER2D (Steffler
39 and Blackburn, 2002), WHYSWESS (Yi et al., 2010) and CASiMIR (Munoz-
40 Mas et al., 2012) at the micro-scale, and MesoHABSIM (Parasiewicz, 2001),
41 MesoCASiMIR (Schneider et al., 2001) and RHM (Maddock et al., 2001) at
42 the mesoscale. Among these PHABSIM and MesoHABSIM are probably the
43 most widely used and representative ones.

44 While MesoHABSIM refers to specific hydromorphologic units (i.e. HMUs
45 Bovee et al., 1998b; Parasiewicz, 2001) and performs a 2D analysis based on
46 detailed input data, PHABSIM analyzes environmental conditions based on
47 1D hydraulic variables through the definition of habitat suitability curves
48 within the Instream Flow Incremental Methodology (IFIM, Bovee et al.,
49 1998a) framework. The ecological variables are key elements of this physical
50 habitat approach, which considers specific target species and requires detailed
51 and site-specific data. In case of limited data availability, expert knowledge is
52 a common practice. The IFIM context allows PHABSIM to identify improve-
53 ments in habitat state from different flow regimes, thus making predictions
54 and supporting the negotiation of suitable water delivery scenarios (Booker
55 and Dunbar, 2004).

56 According to recent e-flow prescriptions, all flow components, from base
57 flow to flood regime, are to be included as operational targets for a sustain-
58 able water resources management (EU, 2000). In this respect, Flow Duration
59 Curves (FDCs), a classical hydrological tool that embeds details on stream-
60 flow regime, which is widely used for flood control, water quality management
61 and hydro-power purposes, represent a meaningful tool for analyzing several
62 ecohydrological issues, such as e.g. the effects of e-flow scenarios on riverine
63 habitat (Vogel and Fennessey, 1995).

64 The scientific literature collects a plethora of studies investigating the
65 potential impacts of different flow releases downstream of hydro-power plants
66 on energy production and riverine ecosystems (see e.g. Snelder et al., 2014;
67 Person et al., 2014; Hirsch et al., 2014; Ayllon et al., 2014; Yin et al., 2014;
68 Vigano et al., 2016; Yi et al., 2017), though only a few of them has employed
69 FDCs, particularly at a regional scale (CAPRA et al., 1995; Ayllon et al.,

70 2012; Pragana et al., 2017). In addition, quantitative assessments of e-flows
71 impacts are considered to be unviable when the availability of hydrological
72 data is limited (i.e. a frequent condition even for high-income countries).
73 To overcome this issue, FDCs and Top-kriging are powerful hydrological
74 tools that can be used to reconstruct streamflow regimes at ungauged sites
75 (Pugliese et al., 2014; Farmer, 2016), thus enabling one to evaluate the hydro-
76 power production and ecological status across large catchments and regions
77 (see e.g. Popescu et al., 2012; Cuya et al., 2013).

78 In this study we quantitatively analyze the effects of alternative e-flow
79 prescriptions on hydro-power production and fish habitat suitability for two
80 Italian river basins by employing FDCs and Top-kriging techniques. In par-
81 ticular, we perform a regional-scale analysis by considering two different e-
82 flow policies (i.e. current policy and a tighter future one, see Section 2.4)
83 identified on the basis of empirical methods and set a-priori by the local
84 Regional Authority without any former insight on possible effects on river
85 biota. We demonstrate how to cope with a limited availability of streamflow
86 data at locations of interest, thus supporting a quantitative assessment of
87 the impacts and the effectiveness of e-flows at regional scale.

88 Our analysis focuses on Barbel (*Barbus barbus*) and Chub (*Leuciscus*
89 *cephalus*) species, which are considered to be near-threatened in Italy (Zeru-
90 nian, 2007) and require high protection level at regional scale. Barbel and
91 Chub spawning occurs from April to June and their habitat requirements are
92 well known (Rambaldi et al., 1997; Bicchi et al., 2006). While it is straight-
93 forward to anticipate a decrease in hydro-power production for higher e-flow
94 releases (i.e. future prescriptions), ecological effects on the considered fish
95 species cannot be easily predicted, nor were ever assessed for the study area
96 by local authorities in charge of defining e-flow policies. To this aim, we em-
97 ploy different habitat suitability criteria to examine whether a loss or a gain
98 in habitat suitability is associated with a modification in the e-flow releases
99 (PTA, 2010). For assessing the ecological effect, we employ the classical
100 PHABSIM approach and we elaborate an analytic approach based on FDCs
101 and on a simpler hydraulic model, hereafter labeled as Habitat Suitability
102 Duration Curve (HSDC), which can be easily applied across hydrologically
103 and ecologically homogeneous large catchments and regions and particularly
104 in any ungauged site.

105 2. Study Area: Data and e-flow Scenarios

106 The quantitative assessment of the effectiveness of two alternative e-flow
107 scenarios and their impact on hydro-power production and fish habitat suit-
108 ability is applied to hydro-power plants located within the Potenza and Chi-
109 enti river basins, in the Marche administrative district in Central Italy (Fig-
110 ure 1 and Table 1).

111 2.1. River basins description

112 The Potenza and Chienti river basins, with a catchment area of 640 and
113 1070 km², respectively, flow northeasterly from the Apennines to the Adriatic
114 Sea. The elevation ranges between 40 and 1400 m asl (above sea level) for
115 Potenza, and 20 and 2000 m asl for Chienti. In these two catchments, agri-
116 cultural areas (59%), forests and semi-natural areas (38%) share the majority
117 of land covers, while human settlements and impervious areas are around 3%
118 (EEA, 2007). Two major geological units dominate the study area from a
119 hydro-geological perspective. The head-water catchments are dominated by
120 fractured carbonate limestones, with frequently emerging subsurface water
121 (see Figure 1), while the downstream area mainly presents sandstones and
122 marble calcarenites. The study area shows a maritime streamflow regime (see
123 Castellarin et al., 2004b,a), whose typical hydrologic year is characterized by
124 a maximum monthly discharge during winter and minimum during summer.
125 The climate of this area is conditioned by the close presence of both the
126 Adriatic Sea and the Apennines, with average annual temperatures ranging
127 from 8°C to 15°C. Mean Annual Precipitation (MAP) values evaluated at
128 catchment scale are reported in Table 2 for the study stream gauges.

129 2.2. Hydro-Power Plants in the Study Area

130 We consider 14 hydro-power plants (see Table 1) operated by the energy
131 multinational power company ENEL Group Ltd. and located within the
132 Potenza and Chienti river basins (see Figure 1). In particular, as illustrated
133 in Table 1, we consider three storage (DAM) and eleven run-of-the-river
134 (ROR) hydro-power plants that share a common feature, i.e. power-houses
135 are located downstream the corresponding dams or barrages and off-line re-
136 lative to the river course. Hence, the water used for hydro-power production
137 needs to be diverted and is returned to the river only downstream the ab-
138 straction point. As a consequence, the operation of the hydro-power plant
139 alters the natural streamflow regime within the river stretch between the

140 abstraction point and the power-plant outlet, which in several cases is a few
141 kilometers long.

142 Several characteristics of the study dams/barrages and hydro-power plants
143 illustrated in Table 1 are accessible from ENEL Group Ltd. technical re-
144 ports and publications (ENEL, 1992; Galeati, 2013a,b). Observed data on
145 the natural streamflow regime (NFR), instead, is sparse or completely miss-
146 ing for barraged and dammed river cross-sections considered in our study.
147 We therefore estimate the natural streamflow regime at abstraction points
148 by referring to the streamflow data described in Section 2.3 and by applying
149 a geostatistical procedure that interpolates empirical flow-duration curves of
150 daily streamflow (FDCs) along the stream-network (see Pugliese et al., 2014;
151 Farmer, 2016).

152 Measurements of stream network hydraulic properties (i.e., river width
153 w , water depth d , flow velocity v , and discharge Q) are available only for a
154 subset (5 out of 14) of the considered hydro-power sites (see red filled symbols
155 in Figure 1). These features are recorded downstream the hydro-power plants
156 in correspondence of four distinct cross-sections within a nearly 100 m long
157 river reach.

158 *2.3. Hydrologic Data*

159 Natural daily streamflow series are available for the study region at 17
160 stream gauges belonging to the former National Hydrographic Service of Italy
161 (SIMN). Observed flow series span over the time period 1920-2000, with an
162 observation period ranging from 5 to 40 years (average record length: 18
163 years). Table 2 reports drainage area and mean annual precipitation (MAP)
164 of catchments upstream each stream gauge as well as some statistics of daily
165 streamflow series (mean annual flow, MAF, minimum and maximum flows,
166 75%, 50% and 25% exceeded flow values). Empirical MAP values, relative to
167 each of the 17 catchments, are estimated using data collected from a rather
168 dense rain gauge network (i.e. 1 rain gauge per $\sim 50\text{km}^2$ on average) during
169 the same time interval of daily streamflow records. Our daily streamflow
170 dataset includes only complete years; missing daily streamflow records have
171 been linearly interpolated for time intervals shorter than one week, while for
172 longer time-intervals of missing observations we have discarded the entire
173 year (see Castellarin et al., 2004a).

174 *2.4. Regional River Regulation: e-flow Prescriptions*

175 Two alternative e-flow scenarios prescribed by the Marche administrative
 176 district are considered in our study: the current time-invariant experimental
 177 e-flow release and the soon-to-be-enforced time-variant e-flow release based
 178 on Water Protection Plan prescriptions (PTA, 2010), hereafter labeled as
 179 PILOT and WPP, respectively. PILOT e-flows will be authorized until Dec.
 180 31, 2019, while from Jan. 1, 2020 the Regional Authority is going to enforce
 181 WPP e-flows.

182 The PILOT e-flow release results from an experimental program agreed
 183 among the Regional administration and ENEL Group Ltd., which allowed a
 184 reduced e-flow release compared to WPP. This scenario refers to the most
 185 common practice in reservoir management, namely to consider a constant e-
 186 flow value across the whole year, regardless of the natural intra-annual flow
 187 variability. Activities aimed at monitoring the environmental effect of e-flows
 188 downstream of dams and barrages are still undergoing and at this stage data
 189 are not currently available.

190 The WPP e-flow release, by recalling the EU Water Framework Directive
 191 (EU, 2000), includes a temporal regulation of e-flows, thus reproducing nat-
 192 ural streamflow regimes of river reaches downstream the abstraction points,
 193 which supposedly enhances ecosystem conservation. WPP e-flow releases
 194 (Q_{WPP} [m³/s]) are computed from the following empirical expression:

$$Q_{WPP} = k \cdot MAF \cdot B \cdot E \cdot \max\{N, I_f\} \cdot G \cdot T, \quad (1)$$

195 where k [-] is an empirical parameter ranging from 0.05 to 0.1; MAF [m³/s]
 196 is the mean annual flow; B [-] is a parameter that takes into account the
 197 hydrogeologic features of the study area ($B = 2$ in upstream river reaches
 198 mainly consisting of fractured carbonate limestones, $B = 1$ in downstream
 199 areas presenting sandstones and marble calcarenites, see Section 2.1 and
 200 Figure 1); E [-] represents the river ecological status (ranging from 1 to 1.4 for
 201 very good or very poor conditions, respectively); N [-] represents the degree of
 202 wilderness of the area around the river reach (ranging from 1.3 for protected
 203 areas, i.e. natural parks, to 1 for urban and rural areas); I_f [-] represents
 204 the river functionality (ranging from 1 to 1.2 for very good to very poor
 205 river functionality, respectively); G [-] is geomorphologic parameter related to
 206 hydraulic and morphological characteristics of the river reach (ranging from
 207 0.9 to 1.1); T [-] is the temporal factor identifying different flow seasons in a
 208 year. In this context, the term season refers to one of the four time intervals

209 identified in the Water Protection Plan of the Marche administrative district,
210 namely: November - January (total duration: 92 days, $T=1.3$), February -
211 March (total duration: 59 days, $T=1.5$), April - June (total duration: 91
212 days, $T=1.3$), July - October (total duration: 123 days, $T=1.0$). Table 3
213 reports PILOT and WPP e-flow releases for the 5 hydro-power sites (see
214 red filled symbols in Figure 1) for which both hydro-power production and
215 habitat suitability are assessed.

216 *2.5. Fish Species for Habitat Suitability Assessment*

217 The study fish species, Barbel (*Barbus barbus*) and Chub (*Leuciscus*
218 *cephalus*), belong to the *Cyprinidae* family and are typical in the study
219 area. Three different life stages are examined, namely juvenile, spawning
220 and adult. Given that the spawning season for both species is between April
221 and June, we consider juvenile fish as those small fish, hatched from eggs
222 spawned in the same year, whereas the adult stage represents individuals
223 older than 1 year. By using the WPP e-flow seasonality described in Section
224 2.4, we associate each life stage with a specific e-flow season as follows: ju-
225 venile is associated with July-October, spawning with April-June, and adult
226 with the whole year, from January to December.

227 **3. Methods**

228 *3.1. Estimation of Natural Flow Regime at Hydro-Power Sites*

229 In order to reconstruct the unknown natural inflows to the considered
230 hydro-power sites, we employ observed streamflow data and apply a geosta-
231 tistical technique. The procedure, which was originally proposed and applied
232 by Pugliese et al. (2014) to the same study region, adapts the Topological-
233 kriging (or Top-kriging, see Skøien et al., 2006), a block-kriging with vari-
234 able support area coinciding with the catchment watershed (see Skoien et al.,
235 2014), and enables the user to predict FDCs in ungauged basins by linearly
236 weighting empirical curves constructed at available stream gauges. Kriging
237 weights used in the linear weighting scheme take catchment size and nest-
238 ing structure of the stream network into account. Using the Top-kriging
239 adaptation by Pugliese et al. (2014), we predict long-term FDCs of daily
240 streamflows at all 14 abstraction points of interest by interpolating empirical
241 period-of-record (i.e. constructed on the basis of all available daily streamflow
242 observations) FDCs. According to the regional river regulation illustrated in
243 Section 2.4, minimum streamflow requirements have to be identified on a

244 seasonal basis (the term season in this context coincides with one of the four
245 time intervals specified in the regional regulation). Therefore, together with
246 the long-term annual FDCs that describe the natural streamflow regime we
247 also predict the long-term seasonal FDCs for the four periods of interest, as
248 defined in the Water Protection Plan of our study area (PTA, 2010). The
249 prediction of seasonal FDCs represents a novel application of the procedure
250 developed by Pugliese et al. (2014). The validation is based on the same
251 leave-one-out cross-validation scheme used in Pugliese et al. (2014) for assess-
252 ing the accuracy of predicted long-term yearly FDCs. The results prove the
253 suitability of the selected approach since the accuracy of predicted seasonal
254 FDCs results to be comparable with the accuracy of predicted yearly curves
255 and certainly acceptable for the scopes of the present analysis (i.e. overall
256 Nash-Sutcliffe Efficiency computed for predicted log-flows in cross-validation
257 varies between 0.91 and 0.94, and is equal to 0.96 for yearly curves, see also
258 Pugliese et al., 2014).

259 *3.2. Computation of Hydro-Power Production*

260 The present section summarizes the different steps required by the com-
261 putation of annual and seasonal hydro-power productions, distinguishing be-
262 tween run-of-the-river and storage power-plants (see Table 1). It is worth
263 emphasizing here that our study neglects the interaction between power-
264 plants located along the same stream (i.e. we do not consider the possible
265 effects of streamflow regulation upstream the considered power-plant, that
266 is we always adopt the NFR as inflow condition). This simplifying working
267 hypothesis is correct when only run-of-the-river power plants exist upstream
268 any given river dam/barrage, but is certainly associated with an approxima-
269 tion when artificial reservoirs with significant storage capacity exist upstream
270 the location of interest. Nevertheless, the hypothesis seems viable in our
271 study given the limited number of hydro-power plants located downstream
272 the study dams (see Figure 1).

273 Concerning the eleven run-of-the-river hydro-power plants listed in Table
274 1, the assessment of hydro-power production under various constraints on
275 e-flow release is straightforward when annual and seasonal FDCs relative to
276 the NFR are available for the barraged river cross-sections (see e.g. Vogel
277 and Fennessey, 1995). Therefore, seasonal FDCs of daily streamflow are
278 predicted at all hydro-power plants via Top-kriging (see Section 3.1). Figure
279 2a provides a graphical example for Montefranco hydro-power plant (see

280 Table 1), which clearly shows that for roughly 10% of the season duration
281 the e-flow value is higher than natural streamflows.

282 Concerning the three storage power plants (see Table 1), since they can
283 store and manage inflow water volumes, the assessment of their hydro-electric
284 productivity cannot be based solely on FDCs representative of the NFR, but
285 it requires continuous, and possibly multi-annual, daily streamflow series
286 and a conceptualization of reservoir management and functioning. Figure
287 2b,c illustrates reconstructed inflows, together with outflows relative to an
288 arbitrarily selected year at Polverina dam (see Table 1). In particular, the
289 figure reports the reconstructed daily inflows (blue line) and the seasonally
290 variable e-flow releases (red line, WPP scenario), which are used as inputs,
291 and the daily series of simulated outflows downstream the reservoir (black
292 line).

293 The computation of yearly and seasonal hydro-power production for e-
294 flow scenarios PILOT and WPP for run-of-the-river and storage power plants
295 relies also on (i) hydro-power plant characteristics (e.g. minimum and max-
296 imum exploitable discharge, see Table 1) and (ii) seasonal e-flow values for
297 the considered scenario (i.e. PILOT and WPP).

298 A detailed description of the computational steps for the evaluation of
299 hydro-power production for any given site and season is reported in the
300 Appendix.

301 3.3. *Habitat Suitability Assessment*

302 The potential impact of PILOT and WPP e-flow scenarios on Barbel
303 (*Barbus barbus*) and Chub (*Leuciscus cephalus*) suitability to the physical
304 habitat within the considered river basins is assessed by coupling the out-
305 flows from the hydro-power sites with habitat suitability criteria (HSC). HSC
306 describe species habitat preferences, ranging from 0 (unsuitable) to 1 (most
307 suitable), by accounting for the effects of hydro-morphological variables (i.e.
308 water depth, HSC_d , flow velocity, HSC_v , and river substrate, HSC_s) on
309 species distribution. Given that habitat suitability of target species changes
310 during a lifetime, HSC are generally defined and associated with different life
311 stages (see Section 2.5). The formulation of HSC should be generally based
312 on field investigations, providing detailed ecological information of the target
313 species from the study area. However, due to the difficulty of collecting suf-
314 ficient data on species habitat, these data are not always available, as in the
315 present study. When local information are missing, expert knowledge is a

316 significant basis and multiple HSC, showing similar hydrological, morpholog-
317 ical and ecological properties to those characterizing the study area, should
318 be considered in order to test for consistency and account for the effects of
319 different formulations. Here we consider two alternative HSC provided by
320 Bicchi et al. (2006) and Rambaldi et al. (1997), both referring to the Cen-
321 tral Apennines in Italy, and therefore suitable for our study area (see Figure
322 S1). Habitat suitability values are then combined together by computing a
323 composite habitat suitability (HS), as the product of $HSC_d \cdot HSC_v \cdot HSC_s$.
324 It is worth highlighting here that, since the formulation proposed by Bicchi
325 et al. (2006) neglects the effect of river substrate (i.e. by assuming a constant
326 maximum preference regardless the substrate characteristics, $HSC_s=1$), for
327 consistency we implement the same condition in the HSC from Rambaldi
328 et al. (1997).

329 We employ two different methodologies to quantify a synthetic indicator
330 of the habitat quality (i.e. Suitable Area Index, SAI [m^2/m]) associated
331 with the different release scenarios illustrated in Section 2.4: (i) the classical
332 PHABSIM procedure and (ii) an analytic method based on FDCs, hereafter
333 labeled Habitat Suitability Duration Curve (HSDC). The habitat quality
334 indicator SAI is then estimated for each hydro-power site (i.e. a total of 5
335 sites, see red filled symbols in Figure 1), each fish species (i.e. 2 species,
336 Barbel and Chub), each life stage (i.e. 3 life-stages: juvenile; spawning;
337 adult), each HSC (2, i.e. Rambaldi et al., 1997; Bicchi et al., 2006) and each
338 flow regime (3, i.e. NFR, PILOT, WPP).

339 Due to the limited hydro-ecologic data availability, we adopt PHAB-
340 SIM although more recent alternatives (see e.g. PHABSIM, RHYHABSIM,
341 RIVER2D, WHYSWESS, CASiMIR, MesoHABSIM, MesoCASiMIR, RHM)
342 are consolidated across the scientific literature. Furthermore, when analyz-
343 ing e-flows at the micro-habitat level within an IFIM context (Bovee et al.,
344 1998a), different flow scenarios and habitat suitability models are to be con-
345 sidered in order to assess the ecological effects and then negotiate e-flows
346 to be prescribed. In our case study, given that e-flow scenarios prescribed
347 by the Regional Authority were determined a-priori without performing any
348 assessment of the effects on river biota, a sort of backward application of
349 IFIM is performed (i.e. from prescribed e-flows, the current and the soon-to-
350 be-enforced, to ecological and hydro-power production effects).

351 *3.3.1. Estimation of Suitability Area Index from PHABSIM and HSDC*

352 PHABSIM divides river cross-sections into several vertical elements, or
 353 grid cells, each one characterized by given flow velocity and water depth for
 354 a given streamflow value. More specifically, flow velocity and water depth
 355 are derived from 4 distinct cross-sections within a nearly 100 m long river
 356 reach in correspondence of each hydro-power site. Given that PHABSIM
 357 hydraulic modeling can be sometimes controversial (Shirvell, 1986; Ghanem
 358 et al., 1996; Dunbar et al., 1998), we calibrated the hydraulic model on the
 359 basis of concurrent observations of discharge and water level through HEC-
 360 RAS simulations (Brunner, 2016), which we then use as inputs to PHABSIM
 361 (see e.g. Nikghalb et al., 2016). The composite habitat suitability is com-
 362 puted for every grid cell and the Weighted Usable Area (WUA, i.e. the
 363 available habitat area for the target species within a river reach [m²/m]) is
 364 then evaluated as a weighted sum of composite suitability and cell area for
 365 each flow scenario (NFR, PILOT and WPP). We estimate WUA values for
 366 a given set of discharges (i.e. streamflow values sampled from FDCs and
 367 associated with 16 durations within the range 0.005-0.995) and then com-
 368 bine these WUA values to quantify SAI as the integral of the WUA-duration
 369 curve:

$$SAI = \int_{D=0}^{D=1} WUA(D)dD \quad (2)$$

370 where $WUA(D)$ is the Weighted Usable Area associated with a duration D ,
 371 ranging from 0 to 1.

372 Concerning the HSDC approach, the composite habitat suitability HS is
 373 based on FDCs and on a simpler hydraulic procedure and it is evaluated
 374 for the entire river cross-section (i.e. without dividing the cross-section into
 375 computational grid cells). By following the procedure proposed by Vogel
 376 and Fennessey (1995), we combine the relation between HS and discharge
 377 (Figure 3b) with the predicted FDC (Figure 3a) and define the HSDC as the
 378 relationship between the composite habitat suitability and the duration or
 379 exceedence probability of the discharge value associated with that HS (Figure
 380 3c). We then compute the Habitat Suitability Index (HSI) as the integral of
 381 the habitat suitability duration curve (shaded areas in Figure 3c):

$$HSI = \int_{D=0}^{D=1} HSDC(D)dD \quad (3)$$

382 where D represents the duration associated with each composite habitat
 383 suitability value.

384 Finally, the Suitable Area Index (SAI) is defined by simply multiplying
385 HSI and the wetted river width, w , from the equivalent rectangular cross-
386 section. A detailed description of the 3-step procedure for the computation
387 of HS as a function of discharge values is reported in the Appendix.

388 4. Results

389 4.1. Computation of Hydro-Power Production

390 Figure 4 shows through a box-plot representation the distribution of rel-
391 ative differences of hydro-power production for the set of 14 plants belonging
392 to Potenza and Chienti river basins (i.e. run-of-the-river and storage power
393 plants) for each reference period (i.e. the entire year and four sub-periods,
394 namely Nov.-Jan.; Feb.-Mar., Apr.-Jun. and Jul.-Oct.). Each value is com-
395 puted as the difference between the hydro-power production associated with
396 PILOT (i.e. current) e-flow releases and with releases that are compliant
397 with the regional water protection plan (WPP e-flow releases, soon-to-be-
398 enforced), divided by the former.

399 All computations of hydro-power production refer to daily inflow series
400 (storage hydro-power plants) or period-of-record yearly or seasonal FDCs
401 (run-of-the-river hydro-power plants) reconstructed for a multiannual time
402 span, roughly extending between 1920 and 2000. Hence, the resulting hydro-
403 power production should be regarded as a long-term prediction of the hydro-
404 power potential for any given study plant. Each prediction is necessarily
405 associated with some degree of uncertainty, resulting from all simplifying
406 assumptions adopted in our study. It is worth noting, though, that we are
407 mainly interested in comparing different estimates of long-term hydro-power
408 productivity rather than assessing their absolute values, which mitigates the
409 impact of simplifying assumptions. One striking feature of Figure 4 is that
410 values are all positive, meaning that the enforcement of WPP releases will
411 result in losses of hydro-power production. This result was expected as the
412 current prescriptions on e-flows are less stringent than WPP ones (see Table
413 3). Another feature of Figure 4 is the dependence of the reduction of hydro-
414 power production on the considered time-period. Average relative differences
415 (in %) are equal to 14.7, 13.7, 5.8, 10.2 and 42.8 for Year, Nov.-Jan., Feb.-
416 Mar., Apr.-Jun. and Jul.-Oct. reference time-intervals, in this order.

417 Production losses are significant, or extremely significant, over the study
418 area. More than 50% of the plants show a production loss larger than 10%

419 on an annual scale, or during the time period between November and Jan-
420 uary, which is one of the two wet seasons for the study catchments, the other
421 spanning between March and early June. Losses become extremely impor-
422 tant between July and October, with 50% of the plants showing losses larger
423 than 44% and in excess of 58% in 25% of the cases. This result is associ-
424 ated with WPP e-flow prescriptions during the summer season (i.e. between
425 July and October), which are particularly severe. During several weeks of
426 the simulation time interval 1920-2000, WPP e-flows resulted to be larger
427 than natural streamflows for the majority of study catchments. In particu-
428 lar, relative to PILOT e-flow releases and on the basis of the computations
429 performed in our study, the enforcement of WPP releases is likely to increase
430 the duration of plant-shutdown periods by 61 and 132 days per year on av-
431 erage for the Potenza and Chienti hydro-power plants, respectively (i.e. c.a.
432 two and four months, respectively), which is an extremely significant amount
433 of time.

434 The marked variability of production losses between different hydro-power
435 plants (e.g. between 7 and 29%, or 20 and 82% for Year and Jul.-Oct., re-
436 spectively) results from the extremely high variability of a few empirical
437 parameters used by the expression adopted in the regional WPP for comput-
438 ing the e-flow. A noteworthy example is the parameter B , which is normally
439 equal to 1 and is set to 2 for all basins that are entirely within with a specific
440 geological unit that mainly consists of fractured carbonate limestones (see
441 Section 2.1).

442 4.2. *Habitat Suitability Assessment*

443 Suitable Area Index (SAI) values computed for the 5 hydro-power sites
444 are shown in Figures 5 and 6 for *Barbus barbus* and *Leuciscus cephalus*,
445 respectively. In order to examine the regionalized (i.e. average) behavior of
446 the study area, characterized by comparable hydro-geomorphic properties, we
447 opted for grouping together the considered river cross-sections. The outcomes
448 are presented in terms of mean \pm standard deviation for the 2 alternative
449 methodologies, namely PHABSIM and HSDC, and the 2 habitat suitability
450 criteria (Bicchi et al., 2006; Rambaldi et al., 1997).

451 As expected, regionalized SAI values computed from the HSDC method
452 are generally characterized by a higher variability compared to the PHAB-
453 SIM ones. Indeed, while the application of PHABSIM generally results in
454 at least some (a few) elementary cells of each river cross-section presenting

455 suitable conditions for fish, the HSDC approach, which assumes an equiva-
456 lent rectangular cross-section under uniform flow conditions, may bring forth
457 either totally unsuitable or suitable states, thus showing a larger SAI range.

458 The application of the habitat suitability criteria proposed by Bicchi et al.
459 (2006) shows a fairly good agreement between SAI values from PHABSIM
460 and HSDC for both fish species (Figure 5a,b,c and 6a,b,c). For the adult life
461 stage (Figure 5c and 6c), PHABSIM and HSDC methodologies consistently
462 reveal a preference for the WPP e-flow scenario, whereas both PILOT and
463 WPP e-flow scenarios present similar outcomes for juvenile (Figure 5a and 6a)
464 and spawning (Figure 5b and 6b) life stages. The NFR always presents the
465 lowest SAI values compared to PILOT and WPP e-flow scenarios. This result
466 is likely to be associated with the large range of discharge values (i.e. from
467 $10^{-1}\text{m}^3/\text{s}$ to $10^3\text{m}^3/\text{s}$), whose extreme conditions (i.e. floods and low-flows)
468 are mostly unsuitable for fish. Indeed, habitat suitability is usually assessed
469 for low-flows, thus entirely disregarding high or very high flow conditions
470 (Booker and Dunbar, 2004). Furthermore, the HSDC approach reveals that
471 juvenile (Figure 5a, 6a) and spawning (Figure 5b, 6b) life stages, which
472 prefer low water depths and flow velocities, and thus lower discharges, may
473 experience particularly small (and even negligible) SAI values under NFR.
474 This condition intimately depends on the discharge values associated with
475 the composite habitat suitability, which may lie outside the discharge range
476 gathered from the FDC (see Figure 3). More specifically, when applying
477 HSDC under NFR, negligible or even null composite suitability values can
478 be associated with the majority of discharge values sampled from the FDC.
479 As a consequence, SAI values may result in extremely low or even null figures.

480 Concerning the application of the habitat suitability criteria proposed by
481 Rambaldi et al. (1997), a satisfactory match between PHABSIM and HSDC
482 approaches on SAI values associated with the different flow regimes is evident
483 only for the adult life stage of both fish species (Figure 5f, 6f). In this case,
484 adult fish reveal an overall preference for NFR conditions, with the lowest
485 values for the PILOT e-flow scenario. Conversely, juvenile (Figure 5d, 6d)
486 and spawning (Figure 5e, 6e) life stages show rather contradictory outcomes
487 with divergent SAI values. Evidently, the linkage among FDCs and the
488 relationship between HS and discharge plays a key role.

489 Our analysis reveals that the two habitat suitability criteria employed in
490 this study, both referring to Apennines rivers with comparable hydrologic
491 regime and water resource availability, despite some differences in terms of
492 absolute quantities, show a rather consistent trend when comparing the three

493 flow regimes (i.e. NFR, PILOT and WPP) either with the PHABSIM or the
494 HSDC methodology. In particular, given that (i) for juvenile and spawning
495 life stages, the two e-flow releases present analogous impacts on habitat qual-
496 ity on the considered fish species (see Figure 5, 6), and (ii) the hydro-power
497 production losses are significant within the associated sub-periods (see Fig-
498 ure 4), from a practical and operational perspective our analysis may suggest
499 within the IFIM framework a review of the WPP e-flow releases, as prescribed
500 by PTA (2010), possibly allowing smaller outflows during the aforementioned
501 seasons.

502 Interestingly, the variability associated with different habitat suitability
503 criteria is reasonably comparable with the overall variability between
504 the proposed methodologies. The analytic HSDC approach, which is based
505 on relatively few and simple hydraulic properties (see also Appendix), can
506 thus constitute a valid alternative to the more complex and data-demanding
507 PHABSIM approach for a fast and rapid identification of potential ecological
508 impacts of different e-flow scenarios at the regional scale. This alternative
509 approach can be successfully applied across hydrologically and ecologically
510 homogeneous river networks, as our case study. Furthermore, to get more
511 reliable results, multiple and possibly site-specific habitat suitability criteria
512 should be considered in future studies.

513 5. Conclusions

514 In this paper we perform a quantitative analysis of the effects of alterna-
515 tive e-flow prescriptions at regional scale. In particular, we focus on hydro-
516 power production and fish habitat suitability (Barbel and Chub fish species)
517 across Chienti and Potenza river basins (Italy) referring to the current time-
518 invariant regional e-flow prescription (PILOT) versus a new time-variant re-
519 gional prescription (WPP), which will be enforced from Jan. 1, 2020. We
520 employ natural and altered flow-duration curves (FDCs) to estimate both
521 hydro-power production and an index of habitat quality (i.e. Suitable Area
522 Index, SAI). The ecological effects are also assessed, for the sake of compar-
523 ison, through the classical PHABSIM approach. The following conclusions
524 are worth summarizing:

- 525 • An evident reduction of hydro-power production shifting from PILOT
526 to WPP e-flow releases emerges without any significant uncertainty.
527 At the annual time-scale, average relative differences are equal to 15%.

528 Higher losses (43% on average) characterize the July-October time in-
529 terval.

- 530 • In addition, we find that WPP e-flows are frequently greater than the
531 actual surface water availability at various cross-sections within the
532 study river networks, thus causing a significant enhancement of shut-
533 down periods for hydro-power plants located in the upstream part of
534 the study area (i.e. fractured carbonate limestones).
- 535 • Given the prescription of a stricter e-flow scenario (i.e. WPP) by the
536 Regional Authority, even though a significant hydro-power production
537 loss is found, a clear outcome does not emerge from the habitat suit-
538 ability assessment.
- 539 • From the ecological perspective, regardless of habitat suitability cri-
540 teria and the methodology employed for assessing habitat conditions,
541 increasing e-flow releases does not show a clear and consistent improve-
542 ment of habitat status for Barbel and Chub. In order to get more accu-
543 rate indications, future studies should preferably consider site-specific
544 habitat suitability criteria or, alternatively, may benefit from adopting
545 multiple (i.e more than one or two) criteria associated with rather ho-
546 mogeneous hydro-geomorphic environments and then refer the average
547 behavior.
- 548 • When comparing NFR and e-flow scenarios for juvenile and spawn-
549 ing life stages, our results show a general preference for e-flows rather
550 than natural streamflow conditions. This result was expected since
551 these life stages tend to prefer low-flow conditions, whereas high- or
552 very high-flows are indeed scarcely suitable for them. Outcomes are
553 not as consistent when it comes to adult life stages. Different SAI val-
554 ues emerge due to the interrelation between FDCs and the composite
555 habitat suitability criteria.
- 556 • The variability of our results associated with different habitat suitabil-
557 ity criteria is comparable with the variability between PHABSIM and
558 HSDC approaches. The HSDC approach proves indeed to be a valu-
559 able alternative method for rapidly assessing habitat suitability at the
560 regional scale, when data availability (both hydrological and ecologi-
561 cal) is limited and hydrologically and ecologically homogeneous river
562 networks are considered.

563 Concluding, the proposed research is not intended to substitute site-
564 specific and detailed studies, but rather to provide regional-scale guidance
565 towards the identification of effective and sustainable e-flow policies for the
566 conservation of fluvial ecosystems, when eco-hydrological data availability is
567 limited and streamflow observations are entirely missing at the locations of
568 interest.

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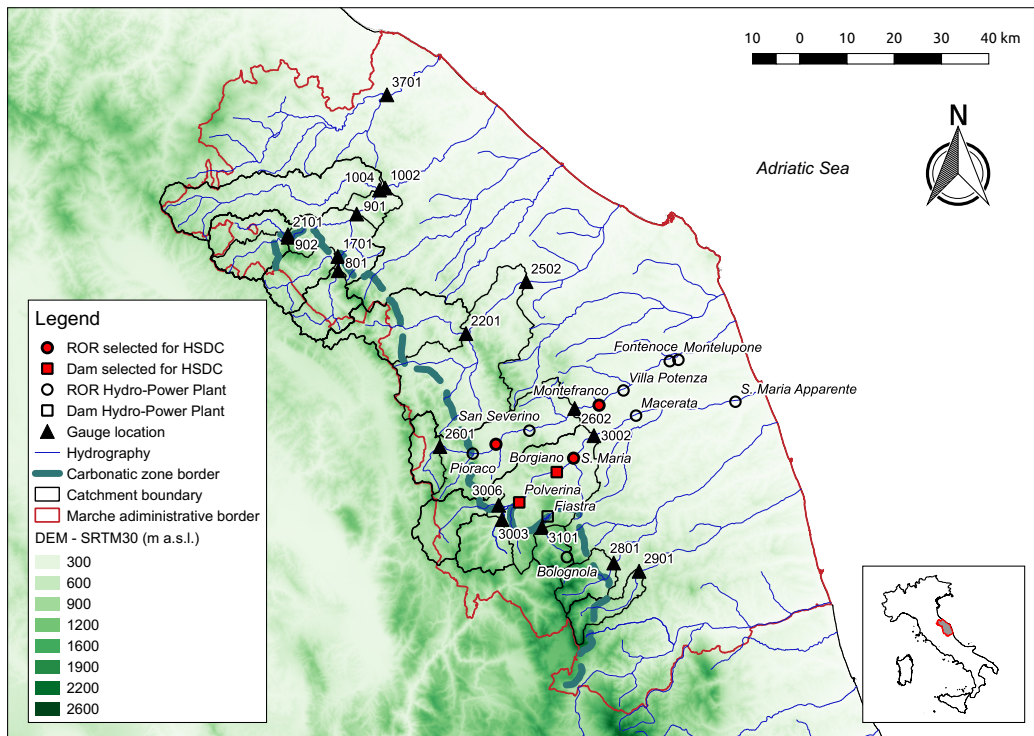


Figure 1: Study area: hydro-power plants considered for assessing hydro-power production (black circles for ROR and black squares for DAM) and fish habitat suitability (red filled symbols); 17 available stream gauges (black triangles) and corresponding upstream catchments (black solid lines) used for the discharge computation at ungauged sites (i.e. hydro-power plants); boundary of the carbonatic zone described in Section 2.1 (dashed thick grey line); Marche Region administrative border (red line).

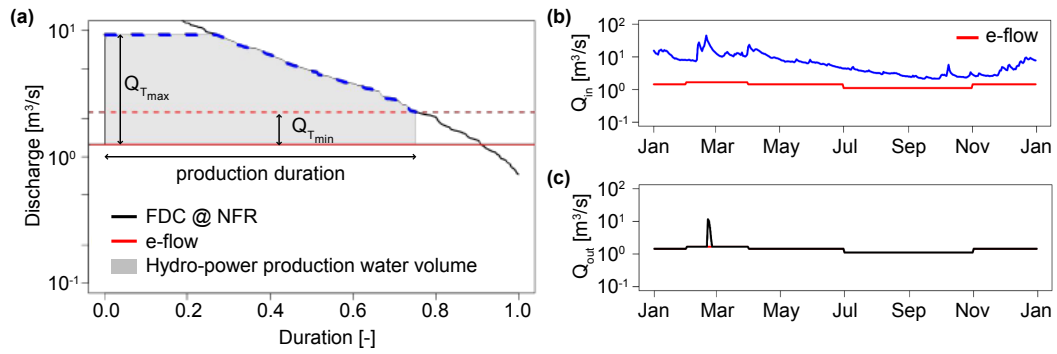


Figure 2: Schematic representation of seasonal hydro-power computation: (a) run-of-the-river power plants (ROR), example of seasonal FDC of daily streamflow estimated for Montefranco via Top-kriging (black solid line), seasonal e-flow (solid red line, current scenario, PILOT), summation of seasonal e-flow and turbine minimum discharge (dotted red line); indication of exploitable discharge (dashed blue line) and water volume (gray shaded area); (b) and (c) storage power plants (DAM), the example refers to Polverina and illustrates for an arbitrarily selected simulation year the reconstructed daily inflows (blue line), seasonal e-flow values (red line, future scenario, WPP), and simulated daily outflows (black line).

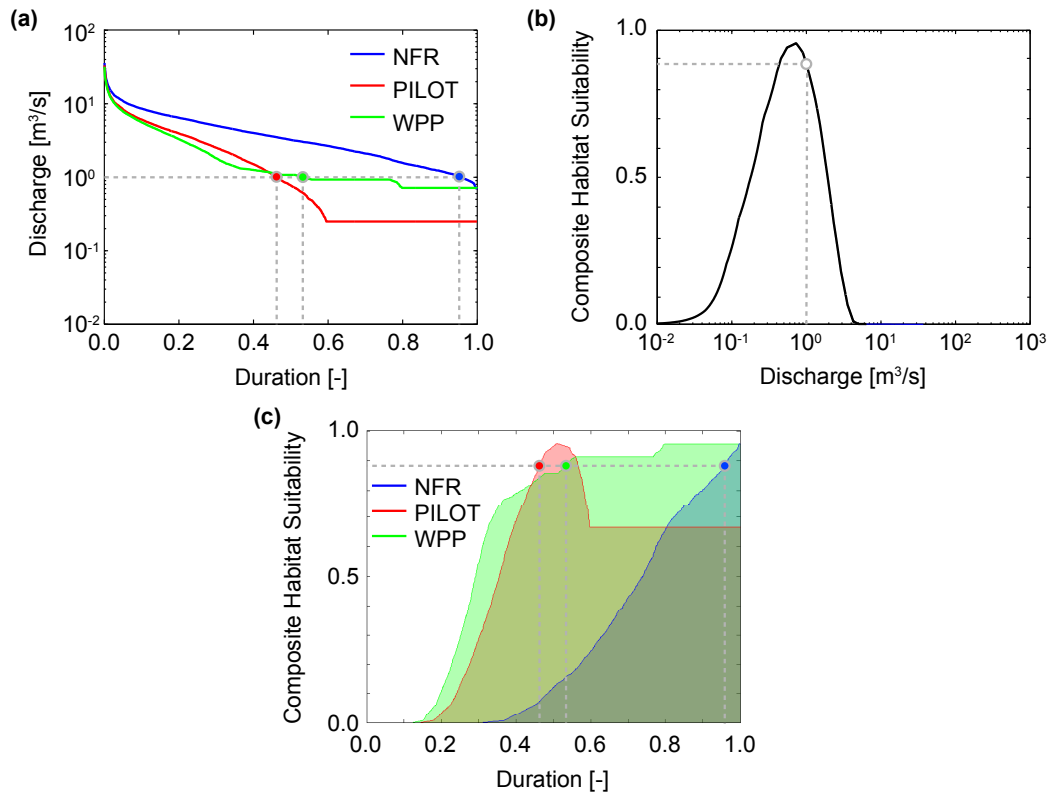


Figure 3: Schematic representation of Suitable Area Index (SAI) computation (example for Castelraimondo power plant): (a) annual flow-duration curves (FDCs) for natural flow regime (NFR, blue line), and current (PILOT, red line) and future (WPP, green line) e-flow policies; (b) composite habitat suitability for Barbel (*Barbus barbus*), adult life stage, derived from Bicchi et al. (2006); (c) habitat suitability duration curves (lines) and Suitable Area Index (shaded areas) associated with NFR (blue), and PILOT (red) and WPP (green) e-flow policies. Blue, red and green dots reported in the plots show the linkage between FDCs, composite habitat suitability and HSDC associated with $Q = 1 \text{ m}^3/\text{s}$.

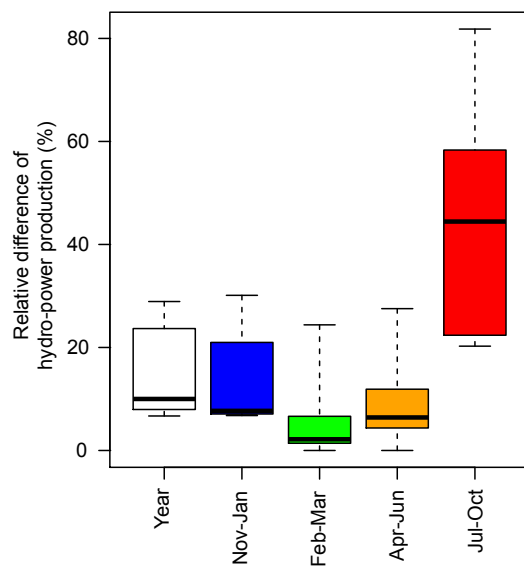


Figure 4: Relative differences (%) of hydro-power production for the set of 14 study plants: each value is computed as the difference between the hydro-power production associated with current (PILOT) and future (WPP) e-flow releases, divided by the former; the box-plots consider the entire year and four sub-periods (namely Nov.-Jan.; Feb.-Mar., Apr.-Jun. and Jul.-Oct.); each box-plot illustrates 25th, 50th (i.e. median, black thick line) and 75th percentiles, together with minimum and maximum values (whiskers).

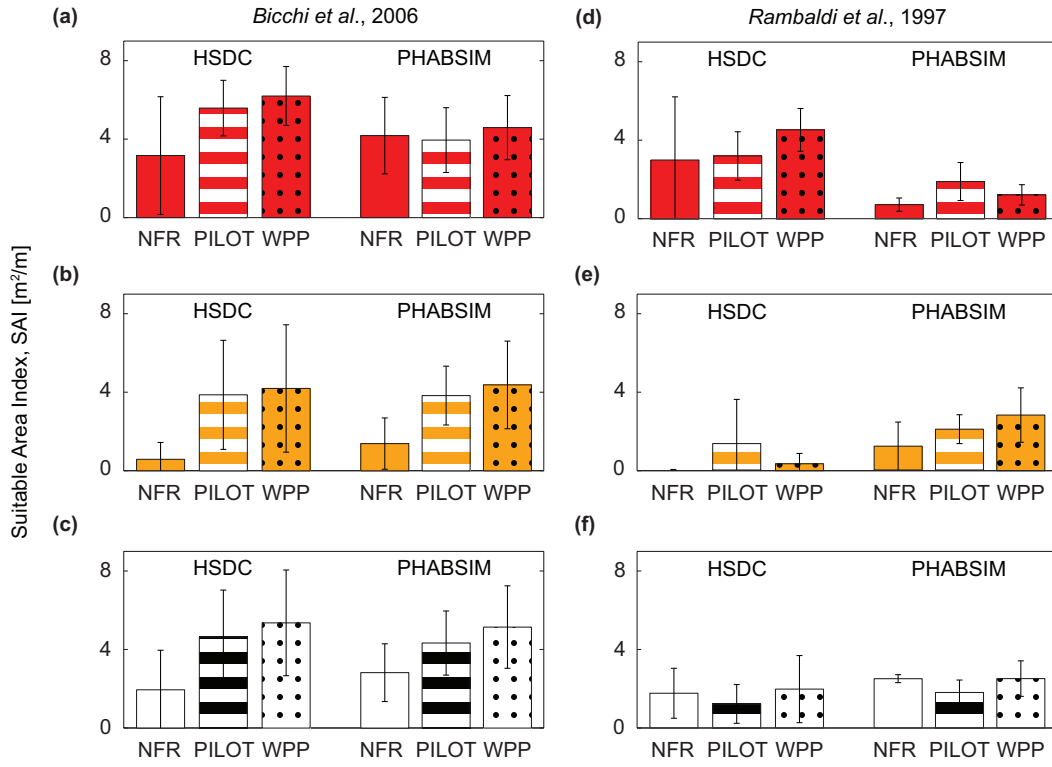


Figure 5: Suitable Area Index values (mean \pm standard deviation) for *Barbus barbuis* for the set of 5 study plants, whose river cross-section properties are available (see Figure 1): (a) and (d) juvenile (period: Jul.-Oct.); (b) and (e) spawning (period: Apr.-Jun.); (c) and (f) adult (period: entire year). Left and right columns refer to the habitat suitability criteria proposed by Bicchi et al. (2006) and Rambaldi et al. (1997), respectively. Filled, striped and dotted patterns represent NFR, PILOT and WPP scenarios, respectively. Bar colors refer to the seasonal representation in Figure 4.

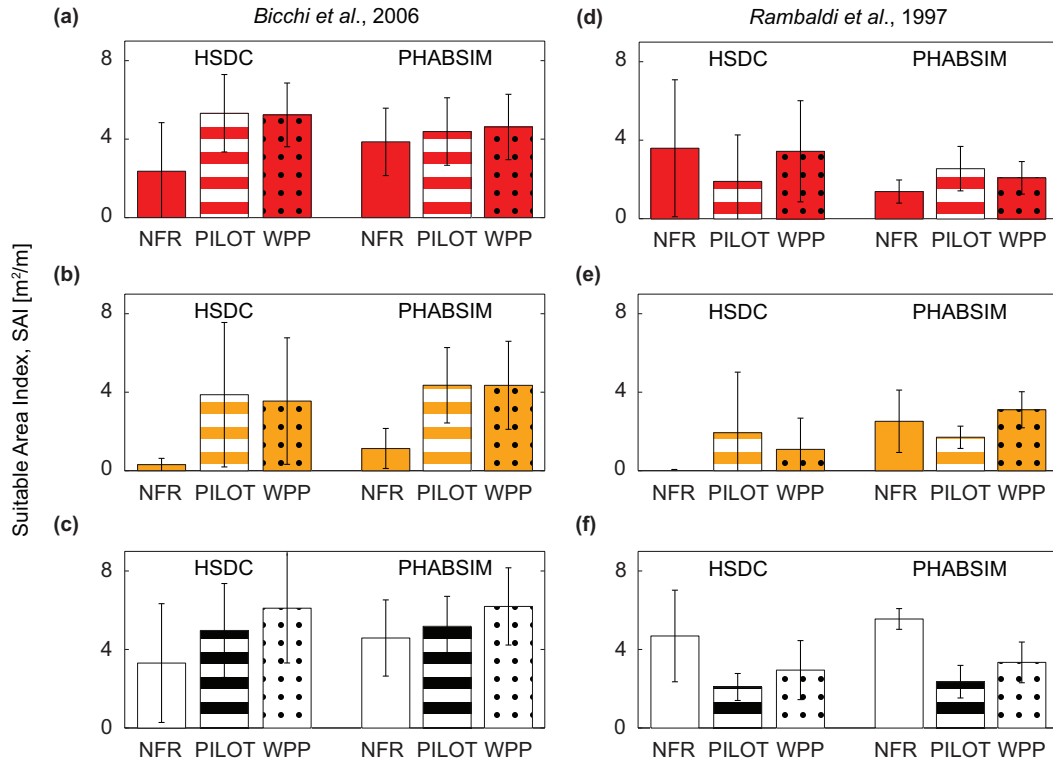


Figure 6: Suitable Area Index values (mean \pm standard deviation) for *Leuciscus cephalus* for the set of 5 study plants, whose river cross-section properties are available (see Figure 1): (a) and (d) juvenile (period: Jul.-Oct.); (b) and (e) spawning (period: Apr.-Jun.); (c) and (f) adult (period: entire year). Left and right columns refer to the habitat suitability criteria proposed by Bicchi et al. (2006) and Rambaldi et al. (1997), respectively. Filled, striped and dotted patterns represent NFR, PILOT and WPP scenarios, respectively. Bar colors refer to the seasonal representation in Figure 4.

Table 1: Main characteristics for the 14 hydro-power plants considered in this study: river name (RN), hydro-power site name (HP), abstraction point location coordinates (Lat, Lon), drainage area (A), Mean Annual Flow (MAF), morpho-environmental parameter for WPP e-flow releases ($C = E \cdot max\{N, I_f\} \cdot G$), minimum and maximum turbine discharges ($Q_{T_{min}}$ and $Q_{T_{max}}$), plant typology (Run-of-the-River, ROR or storage, DAM), storage volume (V).

RN	HP	Lat, Lon (°N, °E)	A (km ²)	MAF (m ³ /s)	C (-)	$Q_{T_{min}}$ (m ³ /s)	$Q_{T_{max}}$ (m ³ /s)	Typology	V (Mm ³)
Potenza	Pioraco	43.18, 12.99	170	3.71	1.30	1.0	6.0	ROR	-
Potenza	Castelraimondo	43.20, 13.05	204	4.32	1.00	0.6	2.5	ROR	-
Potenza	San Severino	43.22, 13.14	316	5.76	1.30	1.5	5.0	ROR	-
Potenza	Montefranco	43.27, 13.32	488	8.04	1.00	1.0	8.0	ROR	-
Potenza	Villa Potenza	43.30, 13.38	532	8.64	1.00	1.0	5.5	ROR	-
Potenza	Fontenoce	43.35, 13.50	643	10.15	1.00	1.5	6.6	ROR	-
Potenza	Montelupone	43.36, 13.52	648	10.24	1.00	1.0	6.5	ROR	-
Chienti	Bolognola	42.98, 13.23	11.3	0.31	1.30	0.2	0.8	ROR	-
Chienti	Fiastra	43.06, 13.18	80.0	1.99	1.30	1.0	10.8	DAM	18.025
Chienti	Polverina	43.09, 13.11	296.0	4.33	1.21	4.0	18.0	DAM	4.047
Chienti	Borgiano	43.14, 13.21	407.0	5.55	1.10	2.0	16.0	DAM	4.263
Chienti	S. Maria	43.17, 13.25	592.0	9.10	1.21	2.0	18.0	ROR	-
Chienti	Città di Macerata	43.25, 13.41	706.0	11.90	1.00	3.0	11.5	ROR	-
Chienti	S. Maria Apparente	43.28, 13.67	1074.0	16.78	1.00	1.5	4.3	ROR	-

Table 2: Key features for the 17 stream gauges considered in this study: streamgauge identifier, name and location (ID, Name, Lat, Lon, in this order), river name (RN), drainage area (A), Mean Annual Precipitation (MAP), years of record (Y), Mean Annual Flow (MAF), Minimum streamflow (Q_{min}), 75th, 50th and 25th streamflow percentiles (Q_{75} , Q_{50} , Q_{25}), Maximum streamflow (Q_{max}).

ID	Name	Lat, Lon (°N, °E)	RN	A (km ²)	MAP (mm)	Y (-)	MAF (m ³ /s)	Q_{min} (m ³ /s)	Q_{75} (m ³ /s)	Q_{50} (m ³ /s)	Q_{25} (m ³ /s)	Q_{max} (m ³ /s)
801	Foci	43.52, 12.64	Burano	123.7	1217.9	7	2.66	0.22	0.65	1.54	3.27	36.4
901	Acqualagna	43.63, 12.69	Candigliano	611.9	1160.1	22	13.4	0.06	2.63	6.86	16	229.3
902	Piobbico	43.59, 12.51	Candigliano	186.1	1163.1	8	4.05	0.03	0.48	1.72	4.64	44.6
1002	Barco di Bellaguardia	43.68, 12.76	Metauro	1040.3	1120	26	21.00	0.20	3.3	10.5	26	361
1004	Calmazzo	43.67, 12.75	Metauro	374.7	1131.1	10	6.36	0.04	0.89	2.86	8.64	93.9
1701	Cagli	43.55, 12.64	Bosso	125.7	1272.5	10	3.05	0.19	0.73	1.58	3.39	111.8
2101	Piobbico	43.59, 12.51	Biscuvio	64.9	1116.5	5	2.14	0.03	0.238	0.86	2.14	40.3
2201	San Vittore	43.40, 12.97	Sentino	262.7	918.1	12	6.3	0.23	1.29	3.42	8.2	127.7
2502	Moie	43.50, 13.13	Esino	795.3	1070.8	11	14.8	1.02	5.29	9.23	17.9	190
2601	Spindoli	43.19, 12.91	Potenza	87.9	1285.2	10	2.52	0.42	1.13	1.93	3.42	16.8
2602	Cannucciaro	43.26, 13.25	Potenza	423.1	1103.7	40	7.94	1.54	3.54	5.69	10	111
2801	Amandola	42.97, 13.35	Tenna	99.3	1145	39	2.82	0.56	1.46	2.32	3.43	42.7
2901	Comunanza	42.96, 13.42	Aso	83.2	946.9	21	2.94	0.77	1.77	2.51	3.42	39.1
3002	Tolentino	43.21, 13.30	Chienti	688.4	1297.5	5	13.7	2.00	4.93	9.81	19	109
3003	Pietrorina	43.06, 13.07	Chienti	115.3	981.0	29	2.16	0.34	0.78	1.66	2.8	20.1
3006	Pontegrove	43.08, 13.06	Chienti	140.9	1112.8	26	1.52	0.06	0.58	1.17	2.05	17.0
3101	Fiume	43.04, 13.17	Fiastrone	60.6	955.7	6	1.66	0.49	0.82	1.42	1.98	7.74

Table 3: Current (PILOT) and future (WPP) e-flow releases for the 5 hydro-power plants considered for both hydro-power production computation and habitat suitability assessment. WPP e-flow releases are reported on a seasonal basis, as defined by the Water Protection Plan of the Marche administrative district (see Section 2.4).

RN	HP	PILOT (m ³ /s)	WPP Nov-Jan (m ³ /s)	WPP Feb-Mar (m ³ /s)	WPP Apr-Jun (m ³ /s)	WPP Jul-Oct (m ³ /s)
Potenza	Castelraimondo	0.250	0.932	1.076	0.932	0.717
Potenza	Montefranco	0.500	0.965	1.113	0.965	0.742
Chienti	Polverina	0.450	1.446	1.668	1.446	1.112
Chienti	Borgiano	0.550	1.314	1.516	1.314	1.011
Chienti	S. Maria	0.700	1.446	1.668	1.446	1.112

777 **Appendix A. Detailed Computational Steps of Hydro-Power Pro-**
778 **duction**

779 *Appendix A.1. Run-of-the-River Hydro-Power Plants*

780 Concerning the eleven run-of-the-river hydro-power plants listed in Table
781 1, the computation of hydro-electric production for any given season consists
782 of the following steps:

- 783 I the production duration (i.e. percentage of time of the season in which
784 the turbine is working) is first identified by comparing the seasonal FDC
785 (black line in Figure 2a) with a constant streamflow value equal to the
786 sum of the seasonal e-flow for the considered scenario and the minimum
787 turbine discharge (red dotted line in Figure 2a);
- 788 II the overall water volume that can be diverted and used for hydro-power
789 production (identified by the gray shaded area in Figure 2a) is then
790 computed by integrating the usable discharge (blue dashed line in Figure
791 2a) over the hydro-power duration identified at step I.;
- 792 III finally, the summation of four seasonal hydro-power productions (i.e.
793 usable water volume) returns the yearly hydro-power production for the
794 considered plant and e-flow scenario.

795 This procedure is repeated for all run-of-the-river power plants and both
796 e-flow scenarios examined in the study.

797 *Appendix A.2. Storage Hydro-Power Plants*

798 As mentioned above, for the three storage power plants natural inflows
799 are not available, therefore synthetic daily streamflow series are generated

800 for these sites by adapting the methodology that was originally presented in
 801 Hughes and Smakhtin (1996) and briefly outlined here. A stream gauge is
 802 selected that is nearby to the target ungauged (i.e. dammed in our case)
 803 cross-section; for this stream gauge the observed daily streamflow series is
 804 continuous (no missing data) and sufficiently long (i.e. at least five years in
 805 this study) and the corresponding watershed is hydrologically similar to the
 806 target site. The observed daily streamflow series is converted into a duration
 807 series by referring to the empirical period-of-record FDC constructed from the
 808 observed streamflow series itself. The duration series is back-transformed into
 809 a daily streamflow series for the target ungauged site by using the long-term
 810 FDC predicted for this site through the geostatistical procedure proposed by
 811 (see Pugliese et al., 2014) and described in Section 3.1. The synthetic daily
 812 natural streamflow series is then used as input to a simplified algorithm that
 813 simulates the reservoir management through the following steps:

- 814 I at any given day, the daily inflow volume is added to the volume stored
 815 during the previous time-step (which is initially set to zero);
- 816 II the code checks the compliance between the stored volume and the e-flow
 817 prescriptions of the scenario (i.e. PILOT or WPP):
 - 818 (a) if the stored volume is larger than or equal to the daily e-flow volume,
 819 the latter is subtracted to the stored volume and the computation
 820 continues to step III;
 - 821 (b) the entire stored volume is released otherwise, and the calculation
 822 moves to the next day (step I) with an empty storage;
- 823 III the stored volume is compared with the maximum, W_{max} , and minimum,
 824 W_{min} , daily volumes that can be exploited for hydro-power production
 825 (i.e. W_{max} is equal to the maximum turbine discharge over a duration
 826 of 24 hours, while W_{min} is equal to the minimum turbine discharge over
 827 a 1-hour duration):
 - 828 (a) if the stored volume is larger than W_{max} , W_{max} is subtracted from
 829 the stored volume and the calculation goes to step IV;
 - 830 (b) if the stored volume is between W_{min} and W_{max} , all stored volume
 831 is used and the computation moves to the next time day (i.e. step
 832 I) with an empty reservoir;
 - 833 (c) if the stored volume is less than W_{min} , the stored volume is held in
 834 the reservoir and the calculation moves to step I;
- 835 IV the stored volume which is left from step III-a is compared with the
 836 reservoir capacity:

- 837 (a) if the stored volume is larger than the reservoir capacity, the ex-
838 cess volume is released downstream, the stored volume is set to the
839 reservoir capacity and the calculation moves to the next day (step
840 I);
841 (b) otherwise, the stored volume becomes the initial volume and the
842 computation starts from step I.

843 The algorithm described above does not aim at faithfully reproducing the
844 real reservoir management and hydraulic behavior, but rather at performing
845 a plausible simulation of reservoir operation at daily timescale, which maxi-
846 mizes hydro-power production while meeting the e-flow prescriptions for the
847 considered scenario.

848 Our simplified numerical code is run for the multi-annual daily inflow
849 time series relative to each one of the three considered storage plants and for
850 all e-flow scenarios. The code returns as outputs the average seasonal and
851 yearly usable water volumes.

852 **Appendix B. Computational steps for the estimation of the com- 853 posite habitat suitability following the HSDC ap- 854 proach**

855 Concerning the HSDC approach, the composite habitat suitability is eval-
856 uated for the entire river cross-section (i.e. without dividing the cross-section
857 into computational grid cells) through the following steps:

- 858 I An equivalent rectangular cross-section with average water depth and
859 flow velocity, derived from river geometry data, is first defined. More
860 specifically, for each of the 4 distinct cross-sections describing the river
861 reach downstream any barrage or dam, we consider the water depth com-
862 puted from HEC-RAS simulations and we then evaluate (i) the wetted
863 area, (ii) the average flow velocity, as the ratio between discharge and
864 wetted area, (iii) the wetted river width and (iv) the water depth asso-
865 ciated with an equivalent rectangular cross-section, as the ratio between
866 wetted area and river width.
- 867 II The relationships between geomorphic features and discharge ($v(Q)$,
868 $d(Q)$) are then computed by applying at-a-station scaling laws devel-
869 oped by Leopold et al. (1964) for each of the 4 distinct cross-sections
870 downstream each barrage or dam. The same 16 discharge values men-
871 tioned earlier (i.e. sampled from FDCs and associated with a duration

872 ranging from 0.005 to 0.995) are then regressed against the correspond-
873 ing average flow velocities and water depths (log-log regression). The
874 regression coefficients computed for the 4 cross-sections are then aver-
875 aged to identify at-a-station coefficients for the river branch downstream
876 each barrage and dam.

877 III For each discharge value gathered from predicted FDCs, the composite
878 habitat suitability HS is finally computed as $HSC_{d(Q)} \cdot HSC_{v(Q)}$.