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Mercury and selenium status of bottlenose dolphins (*Tursiops truncatus*): A study in stranded animals on the Canary Islands

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6 **Mercury and selenium in bottlenose dolphins (*Tursiops truncatus*): a**  
7 **study in stranded animals from the Canary Islands**

8

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28 **ABSTRACT:**

29

30 The mercury (Hg) level in the marine environment has tripled in recent decades,  
31 becoming a great concern because of its high toxic potential. This study [is the first report of inorganic pollutants Hg and](#)  
32 [selenium \(Se\) status, and the first Se/Hg molar ratio assessment](#) in bottlenose dolphins  
33 (*Tursiops truncatus*) inhabiting the waters of [the](#) Canary Islands. Total Hg and selenium  
34 (Se) concentrations were determined in the blubber and liver collected from 30  
35 specimens stranded along the coasts of the archipelago from 1997 to 2013. The  
36 median values for total Hg in the blubber and liver were 80.83 and 223.77  $\mu\text{g g}^{-1}$  dry  
37 weight (dw), and the median levels for Se in both tissues were 7.29 and 68.63  $\mu\text{g g}^{-1}$   
38 dw, respectively. Hg concentrations in the liver were similar to those obtained in  
39 bottlenose dolphins from the North Sea, the Western Atlantic Ocean and several  
40 locations in the Pacific Ocean. The Mediterranean Sea and South of Australia are the  
41 most contaminated areas for both elements in this cetacean species. However, it must  
42 be stressed that the [hepatic contents levels](#) of Hg and Se [in the liver](#) showed an increasing trend with  
43 the age of the animals. Furthermore, as expected, a strong positive correlation  
44 between Hg and Se was observed ( $r_s=0.960$ ). Surprisingly, both younger and older  
45 specimens had a Se/Hg molar ratio different from 1, suggesting that these individuals  
46 may be at greater toxicological risk for high concentrations of both elements or a  
47 deficiency of Se without a protective action against Hg toxicity.

48

49 **Keywords:** [bottlenose dolphins, Tursiops truncatus, Canary Islands, mercury, selenium, blubber, liver, Mediterranean Sea, North Atlantic Ocean, Pacific Ocean](#)

50

51



75 It is well known that the toxic potential of Hg is suppressed in the presence of  
76 sufficient amounts of selenium (Se) (Parizek and Ostadalova, 1967). This effect has  
77 been shown in studies in a variety of species, including marine mammals (Cuvin-Aralar  
78 and Furness, 1991; Frodello et al., 2000; Gui et al., 2014; Sakamoto et al., 2013)  
79 exposed to these elements for a long time, even before the industrial period (Holsbeek  
80 et al., 1998). Thus, several mechanisms for resistance to the adverse effects of Hg have  
81 been proposed. On the one hand, Se can easily combined with various forms of Hg to  
82 yield complexes with lower toxicity, such as methylmercury selenide (MeHg-Se),  
83 methylmercury selenocysteinate (MeHg-Sec), or mercury selenide, tiemannite (HgSe),  
84 which is considered the last step in Hg detoxification (Palmisano et al., 1995). These  
85 compounds also contribute to the mobilization of mercury from the most vulnerable  
86 targets (such as kidney or nervous system tissues) to other less sensitive bodily  
87 regions, such as muscle. Furthermore, Se competes with Hg for its various biological  
88 targets, which also contributes to lowering the potential toxicity of Hg (Khan and  
89 Wang, 2009). Therefore, the Se/Hg molar ratio has been widely used (McHuron et al.,  
90 2014; Mendez-Fernandez et al., 2014; Squadrone et al., 2015; Vos et al., 2003), and  
91 many authors have established that Se<sub>i</sub> in a molar ratio of 1:1 or above with Hg<sub>i</sub>  
92 protects against the toxic effects of this latter metal (Ganther et al., 1972; Ralston et  
93 al., 2007; Ralston and Raymond, 2010; Sormo et al., 2011; Squadrone et al., 2015).  
94 However, paradoxically, this protective action can be harmful to the body because  
95 complex formation also results in the sequestration of both elements, causing them to  
96 become biologically unavailable (Martoja and Berry, 1980). Se is a well-known  
97 essential element with multiple biological functions, such as its critical participation in  
98 reproduction, the metabolism of thyroid hormones or DNA synthesis, in addition to its

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99 important antioxidant role (anticarcinogenic activity), among other functions (Schwarz  
100 and Foltz, 1957; Taylor et al., 2009; Zhang et al., 2014). Therefore, the presence of high  
101 levels of Hg could lead to Se deficiency, which could even cause the death in extreme  
102 cases (Chen, 2012; Sunde, 2006). Thus, the toxicological effects might be due to both  
103 MeHg toxicity and the induced Se deficiency (Zhang et al., 2014). However, Se levels  
104 have increased dramatically in many marine areas, presenting an environmental  
105 toxicity problem (Lavery et al., 2008). Se pollution probably occurs as a result of  
106 anthropogenic activities such as coal burning, smelting, ceramic and glass  
107 manufacturing, or copper refining (van de Merwe et al., 2010). Therefore, to evaluate  
108 the health status of the ecosystems, the simultaneous study of Hg and Se and the  
109 relationship between them is of great interest, particularly in those species usually  
110 considered as sentinels for environmental pollution.

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111  
112 Because of its large size, longevity, and high position within the food chain, many  
113 authors have proposed the cetaceans as good sentinels for ocean health. Species with  
114 a worldwide distribution such as the bottlenose dolphin (*Tursiops truncatus*), are  
115 usually employed to assess global pollution and regional variations (Wilson et al.,  
116 2012). Therefore, this species has been selected for the present study because  
117 previous reports indicate that bottlenose dolphins clearly reflect the contamination of  
118 the waters of the Canary Islands (Eastern Atlantic Ocean) due to their proximity to  
119 likely anthropogenic sources (Garcia-Alvarez et al., 2014a; Garcia-Alvarez et al.,  
120 2014b). Moreover, these cetaceans have been extensively studied, allowing  
121 comparison of the results ~~of the results~~ of this research with other marine areas  
122 around the world, to obtain more comprehensive approach to pollution observations.

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123  
124 [Bottlenose dolphins inhabit the Canary Islands as local resident populations that shows](#)  
125 [inter-island movements within the archipelago \(Tobeña et al., 2014\). This species faces](#)  
126 [a high exposure to organic pollutants \(Garcia-Alvarez et al., 2014b\) and is considered a](#)  
127 [valuable biomarker of the health status of the marine ecosystems. A high](#)  
128 [concentration of contaminants has also been reported in humans from the](#)  
129 [Archipelago \(Luzardo et al., 2012; Luzardo et al., 2009\) and in other marine animals](#)  
130 [from the Canary Islands waters \(Camacho et al., 2014\) and other nearby areas](#)  
131 [\(Camacho et al., 2013\). Although there is a previous research concerning a few](#)  
132 [inorganic pollutants in 12 bottlenose dolphins stranded on the canary coasts \(Carballo](#)  
133 [et al., 2004\), there is a need of more recent and comprehensive data from this marine](#)  
134 [region.](#)

135  
136 The major goals of this study were as follows: 1) adding to recently published  
137 information on chemical pollution in bottlenose dolphins from the Canary Islands  
138 [\(Garcia-Alvarez et al., 2014a; Garcia-Alvarez et al., 2014b\), focusing on the Hg and Se](#)  
139 concentrations in the blubber and liver of stranded animals, and studying their  
140 relationships and toxicity, and 2) reviewing published studies to date on both elements  
141 in bottlenose dolphins worldwide to assess the global impact of these elements on this  
142 species.

143  
144 **2. MATERIALS AND METHODS**

145  
146 *2.1. Study area*

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148 The Canary Archipelago is located in the Eastern North Atlantic Ocean near Europe and  
149 North Africa. These islands are a protected territory with 12 marine Special Areas of  
150 Conservation (SACs) because of the presence of bottlenose dolphins, species listed in  
151 Annex II and IV in the European Habitats Directive (EC, 1992).

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152

153 Bottlenose dolphins inhabit the Canary Islands as a local resident populations that  
154 shows inter-island movements within the archipelago (Tobeña et al., 2014). This  
155 species faces a high exposure to organic pollutants (Garcia-Alvarez et al., 2014b) and is  
156 considered a valuable biomarker of the health status of the marine ecosystems. A high  
157 concentration of contaminants has also been reported in humans from the  
158 Archipelago (Luzardo et al., 2012; Luzardo et al., 2009) and in other marine  
159 animals (Camacho et al., 2014) in nearby areas (Camacho et al., 2013). However, a lack  
160 of data exists concerning heavy metals and other inorganic pollutants from this marine  
161 region.

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## 163 2.2. Sample collection

164

165 Over a period from 1997 to 2013, 29 each of blubber and liver samples were collected  
166 from 30 bottlenose dolphins stranded on the Canary Islands coasts. According to the  
167 literature, Hg and Se were found to accumulate in both tissues, reaching the highest  
168 levels in the liver (Beck et al., 1997). Besides, these tissues have been selected to be in  
169 accordance with previous studies of contaminants in stranded dolphins from this  
170 Archipelago (Garcia-Alvarez et al., 2014a). The blubber is considered as a main target

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171 [for pollutant assessment, in order to possible future comparisons with biopsy samples](#)  
172 [from free ranging cetacean. On the other hand, the liver tissue was also selected](#)  
173 [because pattern distribution of metals is tissue specific, being the mercury mostly](#)  
174 [concentrated in the liver](#) (Krishna et al., 2003).

175  
176 Tissue sampling and the state of decomposition of the stranded specimens were  
177 determined by adapting the Geraci and Lounsbury (2005) protocol. Thirteen males and  
178 17 females (including 2 pregnant females) were ~~grouped-divided~~ into age categories  
179 [i.e., newborn \(1\), calf \(1\), juveniles \(5\), subadults \(11\), adult \(11\) and old \(1\), based on](#)  
180 [body length and gonadal appearance](#). The bodily condition of the specimens was  
181 classified from a good to a very poor state according to morphological  
182 ~~characteristics~~[features](#). All of the characteristics of the animals studied are  
183 summarized in [Table 1](#). Samples were stored in plastic bags at -80C in the Cetacean  
184 Tissue Bank of the University of Las Palmas de Gran Canaria (ULPGC) until analysis.

### 186 *2.3. Sample preparation and analysis of trace elements*

187  
188 All samples were first lyophilized (freeze-dried) for a subsequent microwave digestion  
189 method using a Milestone ETHOS ONE oven. The fresh weight of each sample was  
190 recorded such that the results could be expressed both on a dry (dw) and a wet weight  
191 (ww) basis. In different vessels, 0.5 g aliquots of freeze-dried samples were mineralized  
192 with 6 ml of nitric acid plus 50 µl of Itrio (Y) as an internal standard. Each vessel was  
193 placed into the microwave oven to obtain solutions, which were then diluted to a final  
194 volume of 50 ml with distilled water. After digestion, the analysis of the elements was

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195 performed with an Inductively Coupled Plasma-Optic Emission Spectrometry method  
196 (ICP-OES) using a Perkin Elmer Optima 2100 DV instrument. Two blanks were run  
197 during each analysis to check chemical purity, and the accuracy of the method was  
198 verified with reference materials (lyophilized mussel; CRM 278, Community Bureau of  
199 Reference, BCR, Brussels). All the values of the reference materials were within  
200 certified limits. ~~The instrumental detection limits were 0.061 ng ml<sup>-1</sup> ww for Hg and 0.1~~  
201 ~~ng ml<sup>-1</sup> ww for Se.~~ The recovery values for Hg and Se were 120 ± 8% and 115 ± 11 %,   
202 respectively. ~~The instrumental detection limits were 0.061 ng ml<sup>-1</sup> ww for Hg and 0.1~~  
203 ~~ng ml<sup>-1</sup> ww for Se.~~

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#### 205 [2.4. Data analysis](#)

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#### 207 ~~2.4~~ Calculation of the Se/Hg molar ratio

208

209 The molar ratio of Se to Hg was calculated as:

$$210 \text{Se/Hg} = (\text{Se}/78.96) / (\text{Hg}/200.59)$$

211 where 200.59 and 78.96 g mol<sup>-1</sup> are the molar masses of Hg and Se, respectively.

212

#### 213 ~~2.5~~ Statistical analysis

214

215 Statistical analysis was performed with IBM SPSS Statistics v ~~22~~19.0. ~~Because trace~~  
216 ~~elements did not follow~~ ~~When~~ conditions of normality (Kolmogorov-Smirnov ~~and~~  
217 ~~Shapiro-Wilk tests)~~ ~~and-or~~ homogeneity of variances ~~in all variable groups~~ ~~were not~~  
218 ~~satisfied~~, non-parametric tests were used. ~~Thus,~~ ~~The~~ significant differences between

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219 ~~group medians and distributions statistical significance between different categories~~  
220 ~~were was~~ assessed using the Mann–Whitney U-test and the Kruskal–Wallis test ~~for~~  
221 ~~differences between two or more independent groups, respectively~~. Spearman's

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222 correlation test was performed to determine a possible relationship between both  
223 trace elements. As usual, the level of statistical significance was set at *p-value* = 0.05.

224

### 225 3. RESULTS AND DISCUSSION

226

227 Individual and descriptive statistics of Hg and Se concentrations in the blubber and  
228 liver of bottlenose dolphins (mean ± SD, median and range) are shown in [Table 2](#). The  
229 data are expressed as  $\mu\text{g g}^{-1}$  (ppm) on a dry weight (dw) basis. However, to allow  
230 comparison with other reports, the wet weight (ww) results were also determined  
231 using conversion factors calculated for each sample based on their respective  
232 percentages of dry residue ([Table 2](#)). The mean correction factor for blubber and liver  
233 tissue (0.48 and 0.28 respectively) are comparable with values reported in the  
234 literature ([Becker et al., 1995; Mackey et al., 1995](#)). Mean and median Hg values of  
235 83.36 and 80.83  $\mu\text{g g}^{-1}$  dw were found in the blubber, which were lower than the mean  
236 and median Hg results in the liver (261.56 and 223.77  $\mu\text{g g}^{-1}$  dw, respectively). For Se,  
237 the mean and median levels of 8.96 and 7.29  $\mu\text{g g}^{-1}$  dw in the blubber were much  
238 lower than the Se concentration in the liver, in which the mean value of 211.20  $\mu\text{g g}^{-1}$   
239 dw was quite far from the median of 68.63  $\mu\text{g g}^{-1}$  dw because of the data dispersion.

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240

241 Bioaccumulation of contaminants in marine mammals has been reported to be highly  
242 dependent on both biotic and abiotic factors, such as sex, age, diet and pollution

243 gradients in the aquatic environment (Storelli et al., 2005). Thus, an analysis of Hg and  
244 Se concentrations against different variables (Table 1) is essential to fully understand  
245 the effects of these elements on the specimens studied.

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246  
247 *3.1. Influence of sex and age on mercury and selenium levels*

248  
249 Age is the most important biotic factor for Hg and Se accumulation. Albeit an  
250 increasing hepatic concentration of both trace elements was observed, the small and  
251 unequal sample sizes of age categories discouraged any statistical test assessment.

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252 This enables us to use the body length as a surrogate for age class. Testing despite a  
253 lack of statistical significance in their levels between consecutive age groups ( $p > 0.05$ ).

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254 Because of the small sample size of the age categories, it was interesting to analyze  
255 the correlations between of the pollutant levels with and the length of the animal.

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256 Thus, the hepatic Hg was found to be positively correlated against this variable with a  
257 Spearman coefficient ( $r_s$ ) of 0.769 (Fig. 1A). In accordance with previous authors  
258 (Bellante et al., 2012), an increasing trend throughout the life of cetaceans was

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259 observed (Fig. 1), probably due to bioaccumulation from the continuous uptake of Hg  
260 in the diet and the decreasing ability to excrete this metal and storage in stable forms  
261 such as HgSe (Aguilar et al., 1999; Mackey et al., 1995; Wagemann et al., 2000). Other

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262 authors also found an upward trend of hepatic Hg in sharks and rays with age  
263 (Gutierrez-Mejia et al., 2009; Storelli and Marcotrigiano, 2002), suggesting a higher

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264 rate of assimilation than excretion of Hg and a lower efficiency of detoxification.  
265 Moreover, the largest specimens may capture bigger prey, which are more likely to  
266 contain higher levels of Hg. Although essential trace elements are regulated via

267 homeostasis in marine mammals (Mendez-Fernandez et al., 2014), Se levels in liver  
268 samples were notably higher in adults than in the youngest individuals (Fig. 1) and  
269 were also correlated with animal length ( $r_s=0.764$ ) [in accordance with other studies](#)  
270 (Woshner et al., 2001). [The high level of Se may result from its accumulation with Hg](#)  
271 [during the detoxification process or from a highly concentrated diet because most](#)  
272 [ocean fish are Se-rich, as was reported in a study in the Faroe Islands \(Budtz-Jorgensen](#)  
273 [et al., 2007\).](#)

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274  
275 Fig. 1B also shows tissue distribution [of Hg and Se throughout the lifetime of the](#)  
276 [animals,](#) highlighting the concentration of both elements mostly in the liver, with a  
277 statistically significant difference between tissues ( $p=0.000$ ). Therefore, the liver  
278 appeared to be the preferential tissue, as indicated in previous studies (Bellante et al.,  
279 2012; Frodello et al., 2000). [-However, it is also interesting to stress that the newborn](#)  
280 [and calf specimens of this research accumulated greater levels of Hg in the blubber](#)  
281 [than the liver, especially for the calf. Concerning the Se, the newborn showed equal](#)  
282 [levels in both tissues and the calf individual doubled the concentration in the blubber.](#)  
283 [These results were in contrast to the following sampling ages where the liver showed](#)  
284 [higher levels of both trace elements with an evident increasing trend. One could](#)  
285 [hypothesize that this variation on the tissue distribution is due to the different](#)  
286 [pollutant sources. Thus, the Hg and Se were initially transferred through placental](#)  
287 [barrier entering the fetal circulation to be transported to the liver for metabolism, and](#)  
288 [then distributed to the blubber for accumulation. The calf showed the highest](#)  
289 [concentrations of both elements in the blubber among all the animals studied, likely as](#)  
290 [a result of exposure through lactation. On the contrary, the juvenile group had the](#)

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291 [lowest level of Hg and Se in the blubber compared to the rest of the age categories.](#)  
292 [This may be due to the release of Hg and Se from the blubber into the circulation at](#)  
293 [weaning stage, which could be considered a period of negative energy balance](#) (Louis  
294 et al., 2015). [However, little is known concerning the factors that affect mobilization of](#)  
295 [pollutants from adipose tissue](#) (Louis et al., 2014). [Therefore, this finding should be](#)  
296 [interpreted with caution also because only one newborn and one calf of bottlenose](#)  
297 [dolphins were available for this study.](#)

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298  
299 In this study no influence of sex on Hg or Se accumulation was observed, as has been  
300 found in other marine areas of the North Atlantic Ocean (Mendez-Fernandez et al.,  
301 2014).

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### 302 303 3.2. Influence of stranding location on mercury and selenium levels

304  
305 Clear differences for Hg and Se concentrations in the liver between the geographical  
306 areas, in which animals stranded, were found (with ~~p-value~~p-values of 0.060 and 0.046  
307 respectively). The results for the eastern Canary Islands of Lanzarote (LZ),  
308 Fuerteventura (FV) and Gran Canaria (GC) showed the highest Se and Hg levels as  
309 compared to the western islands of Tenerife (TF) and La Gomera (LG). Thus, median  
310 hepatic Hg molar concentration from animals stranded in the eastern islands (1824.21  
311 nmol g<sup>-1</sup> dw) was 4 times greater than that from specimens stranded in the western  
312 islands (427.97 nmol g<sup>-1</sup> dw). This difference between both Canary regions was even  
313 more prominent for Se, which reached a 22-fold hepatic molar concentration in the  
314 eastern region of the Archipelago. In fact, there is a decreasing trend from the nearest

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315 to the furthest island from the African continent (Supp. Fig. 1). This finding could be  
316 related to geographical differences of natural and/or anthropogenic sources, but is  
317 more likely affected by the age of the animals at the various locations. The youngest  
318 individuals were found stranded in the western Canary Islands, so these results should  
319 be cautiously considered.

320

### 321 3.3. Temporal trends of mercury and selenium concentrations

322

323 Figure 2 illustrates the total Hg and Se concentrations in the blubber of individuals  
324 grouped according to the year of stranding (between 1999 and 2013). For this context,

325 it is preferable to [analyze the blubber samples because no influence of age or length](#)

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326 [on Hg or Se levels in this tissue was obtained in the present study \(see Fig. 1B\)](#). Despite

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327 the low sample size of the groups (1999, n=2; 2000, n=1; 2002, n=1; 2003, n=2; 2004,  
328 n=1; 2005, n=5; 2007, n=1; 2008, n=4; 2009, n=2; 2010, n=2; 2011, n=5; 2012, n=2;

329 2013, n=1) [required careful interpretation](#), an increasing temporal trend of Hg in the

330 blubber can be seen throughout the study period. [In addition, the Mann-Whitney U-](#)

331 [test revealed a significant difference \( \$p=0.016\$ \) between 2005 and 2011, each with 5](#)

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332 [specimens available. Excluding the samples from years with only one value, the](#)

333 [Kruskal-Wallis test revealed a significant difference \( \$p=0.029\$ \)](#). From the individual

334 stranded in 1999 to the last one in 2013, the Hg level has tripled in the blubber,

335 consistent with a recently published report (Lamborg et al., 2014). Lamborg's group

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336 found that deep and intermediate North Atlantic waters are abnormally enriched in

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337 Hg, probably because of anthropogenic activities such as mining and fossil fuel

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338 combustion. Furthermore, no temporal differences were observed in the Se burden in  
339 blubber (Fig. 2) or the hepatic levels of either trace element (data not shown).

340

#### 341 3.4. Study of the relationship between mercury and selenium

342

343 As previously mentioned, a common approach to assess the risk of exposure to Hg is to  
344 determine the molar ratio of Hg and Se in the body (Se/Hg). A high positive correlation  
345 between Hg and Se with an equimolar ratio in the liver as well as the protective effect  
346 of Se against Hg toxicity is well documented (Cuvin-Aralar and Furness, 1991; Geraci,  
347 1989; Koeman et al., 1973; Yang et al., 2007).

348

349 In the present study, the results showed that increasing Hg levels were associated with  
350 increasing Se concentrations, as described for other dolphin populations (Palmisano et  
351 al., 1995). Spearman's correlation coefficient ( $r_s$ ), calculated between molar  
352 concentrations of hepatic Hg and Se, showed a strong positive relationship (Fig. 3).

353 Excluding the outlier data for Se (CET 407), the correlation slightly decreased from  
354 0.960 to 0.955, although the coefficient of determination ( $R^2$ ) for linear regression  
355 increased from 0.592 to 0.807. It is remarkable that the strongest linear association  
356 ( $R^2=0.973$ ) between these two elements occurred below  $1500 \text{ nmol g}^{-1}$  ( $300 \mu\text{g g}^{-1} \text{ dw}$ )  
357 of Hg in the liver, comparable to a total Hg threshold of  $100 \mu\text{g g}^{-1} \text{ ww}$ , as obtained by

358 other authors (Palmisano et al., 1995). Above this concentration, the level of hepatic  
359 Se significantly exceeded the Hg concentration, so the Se/Hg molar ratio was higher  
360 than 1 (Fig. 3B and 4). Se was in molar excess of Hg in 11 of 29 livers evaluated (37.9%).

361 Other publications reported a similar levels of Se compared to Hg in both pelagic fish

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362 (Kaneko and Ralston, 2007) and cetaceans (Mendez-Fernandez et al., 2014), and the  
363 authors stated that this excess reflects the good health status of individuals or a high  
364 proportion of young animals. In the present study, individuals with a Se/Hg molar ratio  
365 above 1 were all included in older categories (Fig. 4), contrary to such statement and  
366 other results obtained for several cetaceans species (Caceres-Saez et al., 2013;  
367 Palmisano et al., 1995; Yang et al., 2007). Regarding the place of stranding, LZ and FV  
368 showed a Se/Hg molar ratio over 1; by contrast, GC, TF and LG had a median ratio  
369 below 1 (Supp. Fig. 1, insert panel). However, the limited sample size per group  
370 undercut any conclusion.

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371  
~~372 Recently, different criteria have been developed to determine whether the  
373 consumption of certain foods by humans presents a health risk, so dietary  
374 recommendation can be issued. The positive values correspond to a Se/Hg molar ratio  
375 above 1, but are also associated with a hidden risk for Se deficiency and poisoning.  
376 Zhang and colleagues attempted to develop a new criterion for an exposure  
377 assessment to Se/Hg that they called the benefit risk value (BRV), taking into account  
378 the amount of Se required for normal biological functions and the threshold intake  
379 value for Se poisoning. Unfortunately, such a calculation could not be performed in  
380 this study, due to a lack of specific information for marine mammals essential for the  
381 development of the equation, which might not be applicable to an assessment of  
382 health risk in bottlenose dolphins.~~

383  
384 3.5. Assessment of the health risk of mercury and selenium

385

386 Wagemann and Muir (1984), established a threshold for hepatic damage in marine  
387 mammals in a range of 100-400  $\mu\text{g g}^{-1}$  ww Hg burden (Wagemann and Muir, 1984). In  
388 supplementary figure 2, the total hepatic Hg concentration is individually compared to  
389 this threshold. The results indicate that 10 of 29 livers of stranded individuals (34.5%)  
390 exceeded the minimum Hg tolerance level, and 4 had values just below 105  $\mu\text{g g}^{-1}$  ww.  
391 All these samples were from subadult and adult specimens, corresponding to 45.5% of  
392 the total of subadults and adults in this study. Other authors obtained comparable  
393 results for stranded bottlenose dolphins in Australian and Floridian waters (Lavery et  
394 al., 2008; Stavros et al., 2011). These results coincided with animals with a Se/Hg molar  
395 ratio greater than 1.

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396  
397 ~~Experimental studies suggest that Hg intoxication could cause weight loss, toxic~~  
398 ~~hepatitis, renal failure and death in marine mammals. Additionally, high Hg levels have~~  
399 ~~been associated with parasitic infection and pneumonia, resulting in a lower~~  
400 ~~resistance to infectious diseases. After an experimental intoxication of seals, the~~  
401 ~~hepatic Hg level reached more than 500  $\mu\text{g g}^{-1}$  dw after death. Other reports~~  
402 ~~associated chronic Hg accumulation with liver abnormalities observed in stranded~~  
403 ~~bottlenose dolphins from the Atlantic Ocean. Rawson's group found a correlation~~  
404 ~~between the lipofuscin pigment with the hepatic Hg concentration. Large deposits~~  
405 ~~were observed when the Hg level exceeds 60  $\mu\text{g g}^{-1}$  ww.~~

406  
407 ~~The organic form of Hg (MeHg) appears to be the form of Hg most toxic to animals.~~  
408 ~~However, after the results of *in vitro* studies carried out by Betti and Nigro, 1996, an~~  
409 ~~adaptation acquired by dolphins to counteract the toxic effects of MeHg was~~

410 ~~suggested . As mentioned earlier, the formation of insoluble tiemannite granules~~  
411 ~~provides the ability to endure high Hg exposures to odontocetes . Thus, Hg and Se~~  
412 ~~levels above 2000  $\mu\text{g g}^{-1}$  dw were reported in animals with no signs of poisoning~~  
413 ~~because of the protection provided by the combined presence of both trace elements;~~  
414 ~~however, the energy cost of the detoxification is difficult to assess.~~

415

416 The results discussed above indicate that the youngest and oldest bottlenose dolphins  
417 may be of greater toxicological concern (Fig. 4). Although the newborn and the calf  
418 among the animals studied had the lowest Hg content, they were deficient in Se which  
419 could lead to Hg toxicity and they also had a Se/Hg molar ratio less than 1, indicating a  
420 limited protection by Se. This result is consistent with human studies in which authors  
421 argue that prenatal and postnatal Hg exposure negatively affects central nervous  
422 system functions (Rasmussen et al., 2005). Additionally, a molar ratio of 1 or lower  
423 may indicate that all of the available Se is bound to Hg, conferring a possible oxidative  
424 stress risk (Caceres-Saez et al., 2013). However, these results must be carefully  
425 considered because there was only one specimen available from each, newborn and  
426 calf categories. By contrast, the older animals had the highest concentrations of both  
427 Hg and Se in the liver and Se/Hg ratios greater than 1 suggesting a Se molar excess  
428 which could become toxic at high levels (O'Hara et al., 2003). Nevertheless, the inter-  
429 relationships between the Hg and Se concentrations, age, nutritional status and  
430 disease are complex (Law et al., 2012) and the limits of deficiency, essentiality, and  
431 poisoning is quite difficult to assess and not well studied.

432

433 *3.6. Mercury and selenium in bottlenose dolphins from different marine areas*

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434

435 A comparison of Hg and Se levels in bottlenose dolphins from the Canary Islands (this  
436 study) and from different marine areas worldwide was made (Table 3). All these  
437 published results, compiled as ranges of Hg concentrations in the liver, were plotted  
438 on a map (Fig. 5), as well as others previously performed using Hg content in the hair  
439 of pinnipeds (McHuron et al., 2014) and organic pollutants in cetaceans (Aguilar et al.,  
440 2002). Bottlenose dolphins from the Mediterranean Sea had greater Hg concentrations  
441 than published values elsewhere, as was previously reported for striped dolphins  
442 (Andre et al., 1991). Even within the same marine area some differences in the Hg  
443 content were observed. Thus, the Ligurian and Tyrrhenian Sea, showed the maximum  
444 measured of hepatic Hg (13150  $\mu\text{g g}^{-1}$  dw) ever reported before, followed by the  
445 Adriatic Sea which appears to be significantly more polluted by Hg than the less-  
446 contaminated Eastern Mediterranean coast. The Hg levels from the North Sea and the  
447 Northeast Atlantic Ocean, including the Canary Islands (results from this study), were  
448 similar to mean concentrations in bottlenose dolphins from the Western Atlantic  
449 Ocean and from several locations in the Pacific Ocean (Hong Kong and east coast of  
450 Australia), but not to results from the south coast of Australia where a higher Hg  
451 contamination occurred that nearly matched the Adriatic Sea values (see Table 3 for  
452 references). Thus, this last sea displayed 6 times higher Hg burden compared with the  
453 results from bottlenose dolphins from the Canary Islands (present study).  
454 Furthermore, the Tyrrhenian Sea showed the highest Hg value obtained in the  
455 literature, more than 50 times greater than the values obtained in this study. On the  
456 other hand, *T. truncatus* and *T. aduncus* from South of Australia had 3 to 7 times  
457 higher Hg levels respectively, than the specimens from the Canary Archipelago.

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458

459 It has been reported that Mediterranean prey had higher Hg levels than Atlantic prey  
460 [\(Lahaye et al., 2006\)](#), which explains the Hg enrichment in the Mediterranean food  
461 webs, and also in the liver of bottlenose dolphins. The authors suggest that this might  
462 be due to natural Hg sources in the Mediterranean Sea [\(Andre et al., 1991\)](#) and high  
463 anthropogenic Hg emissions especially from France [\(Bellante et al., 2012\)](#).

464

465 There are not many studies on Se levels in the liver of bottlenose dolphins ([Table 3](#)).  
466 The mean hepatic concentration of Se was below 50  $\mu\text{g g}^{-1}$  ww in most marine areas  
467 worldwide, but two locations far exceeded this value, which also corresponded to  
468 places that had the highest Hg burdens, the Ligurian Sea [\(Capelli et al., 2008\)](#) and the  
469 liver of *Tursiops aduncus* in the south of Australia [\(Lavery et al., 2008\)](#). The bottlenose  
470 dolphins from both regions showed 8 and 3-fold greater Se levels, respectively, than  
471 the results obtained in this study. These geographical differences are difficult to  
472 explain because Se is an essential element and many factors, such as dietary intake or  
473 natural sources, but also differences in physiologic needs or the retention of Se for  
474 detoxification processes, might influence its concentrations, [\(McHuron et al., 2014\)](#).

475

#### 476 **4. Conclusions** \_\_\_\_\_

477

478 The present study contains the first reported evidence of Hg and Se concentrations in  
479 the blubber [of bottlenose dolphins stranded along the coasts of the Canary Islands](#) and  
480 [broadens the data previously available in liver tissue of bottlenose dolphins from the](#)

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481 [Canary Islands. In addition, it represents the first Se/Hg molar ratio assessment in](#)  
482 [cetaceans from this marine area.](#)

483

484 There is an increasing temporal trend of Hg concentration during the period of the  
485 study (1997-2013) and is consistent with recently published results for Hg in Atlantic  
486 waters (Lamborg et al., 2014).

487

488 Hg and Se accumulate in the liver of dolphins during their lifetime and are strongly  
489 positively correlated with each other. Hg increases with [age-body length](#) probably  
490 because of continual dietary uptake and Se due to detoxification processes or from  
491 eating Se-rich fish. Individuals with Se/Hg molar ratios over 1 are all subadults and  
492 adults. Conversely, young animals have lower Hg burdens and are also deficient in Se.

493 Thus, according to our results, the youngest and oldest animals seem to be of greater  
494 toxicological concern. [In addition, variation on these two elements in the blubber](#)  
495 [between the earliest stages of life \(newborn and calf\) and the following ages, likely](#)  
496 [indicates the influence of lactation and weaning on the lipophilic pollutant](#)  
497 [accumulation. Nevertheless, this finding must be carefully discussed considering the](#)  
498 [limited data available per age group.](#)

499

500 A comparison of the present study with literature values from other worldwide marine  
501 areas indicates that hepatic Hg results from this part of the Northeast Atlantic Ocean  
502 are similar to those obtained in bottlenose dolphins from the North Sea, the Western  
503 Atlantic Ocean and several locations in the Pacific Ocean. The Mediterranean Sea and  
504 the South of Australia are hot spot contaminated areas for both elements; by contrast,

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505 the median results of this study show that the bottlenose dolphin population from the  
506 Canary Islands is not especially threatened by Hg or Se. However, it must be  
507 emphasized that the concentrations of the elements were highly variable between  
508 specimens; some fall into the Hg threshold established for hepatic damage, and others  
509 are Se deficient. In light of these results, further work is required to assess the  
510 individual effects of high loads of Hg and either large amounts or a deficiency of Se. In  
511 addition, an evaluation of the possible toxic impact of chronic exposure is also  
512 necessary.

513

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519 Silvi for their technical assistance.

520

#### 521 **FIGURE CAPTIONS**

522

523 Figure 1. (A) Hepatic mercury (above) and selenium (below) concentrations in  
524 bottlenose dolphins correlated with the total length of the specimens with Spearman  
525 correlations ( $r_s$ ) of 0.769 and 0.764 respectively. (B) Mean concentrations ( $\mu\text{g g}^{-1}$  dw) of  
526 mercury (above) and selenium (below) in the blubber and liver of bottlenose dolphins  
527 comparing age groups. Sample size (n) is in brackets. Note that there is only one  
528 animal in the newborn, calf and old age categories.

529

530 Figure 2. Temporal distribution of total mercury and selenium concentrations ( $\mu\text{g g}^{-1}$   
531 dw) in blubber samples of bottlenose dolphins stranded from 1999 to 2013 [on the](#)  
532 [Canary Islands coasts](#). The plot represents the mean [level](#) with standard deviation (SD).  
533 [Sample size in each group: 1999 \(n=2\); 2000 \(n=1\); 2002 \(n=1\); 2003 \(n=2\); 2004 \(n=1\);](#)  
534 [2005 \(n=5\); 2007 \(n=1\); 2008 \(n=4\); 2009 \(n=2\); 2010 \(n=2\); 2011 \(n=5\); 2012 \(n=2\);](#)  
535 [2013 \(n=1\). Inset: temporal trends of both trace elements \( \$\mu\text{g g}^{-1}\$  dw, individual or](#)  
536 [mean values\) in the liver of the animals studied. Sample size in each group: 1997 \(n=1\);](#)  
537 [1999 \(n=2\); 2000 \(n=1\); 2002 \(n=1\); 2003 \(n=2\); 2005 \(n=5\); 2007 \(n=1\); 2008 \(n=4\);](#)  
538 [2009 \(n=2\); 2010 \(n=2\); 2011 \(n=5\); 2012 \(n=2\); 2013 \(n=1\).](#)

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539

540 Figure 3. (A) Correlation between mercury and selenium molar concentrations ( $\text{nmol g}^{-1}$   
541 dw) in the liver of bottlenose dolphins from the Canary Islands. Spearman correlation  
542 ( $r_s=0.955$ ) excluding the outlier data (CET 407) with a graphic representation of linear  
543 ( $R^2=0.807$ ) and potential cubic regression ( $R^2=0.880$ ). Linear regression of Hg molar  
544 concentration below 1500 ( $R^2=0.973$ ). (B) Dependence of the Se/Hg molar ratio on the  
545 total mercury ( $\mu\text{g g}^{-1}$  dw) in liver samples of bottlenose dolphins. Spearman correlation  
546 ( $r_s=0.943$ ) considering all samples ( $n=29$ ) and its potential cubic regression ( $R^2=0.937$ ).  
547 Spearman correlation excluding the outlier value ( $r_s=0.937$ ) and its linear regression  
548 ( $R^2=0.789$ ).

549

550 Figure 4. Trends for hepatic mercury and selenium levels with the age of bottlenose  
551 dolphins (median of molar concentrations,  $\text{nmol g}^{-1}$  dw); the selenium outlier data (CET  
552 407) is excluded. Sample size (n) is in brackets.

553

554 Figure 5. Relative mercury concentration in the liver of bottlenose dolphins from the  
555 Canary Islands (the present study) and other marine areas worldwide (see Table 3 for  
556 references). 1, < 50 ( $\mu\text{g g}^{-1}$  ww); 2, 50-100 ( $\mu\text{g g}^{-1}$  ww); 3, 100-500 ( $\mu\text{g g}^{-1}$  ww); 4, > 500  
557 ( $\mu\text{g g}^{-1}$  ww).

558

559 Supplementary Figure 1. Median molar concentrations ( $\text{nmol g}^{-1}$  dw) of mercury and  
560 selenium in the liver of bottlenose dolphins stranded on several islands of the Canary  
561 Archipelago. The graphic inserted at the top right represents plot boxes of hepatic  
562 Se/Hg molar ratios comparing location groups. LZ, Lanzarote (n=5); FV, Fuerteventura  
563 (n=2); GC, Gran Canaria (n=8); TF, Tenerife (n=13); LG, La Gomera (n=1).

564

565 Supplementary Figure 2. Total mercury and selenium concentrations ( $\mu\text{g g}^{-1}$  ww) in 29  
566 livers of stranded bottlenose dolphins compared to the 100-400  $\mu\text{g g}^{-1}$  ww range  
567 threshold proposed by Wagemann and Muir, 1984.

568

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